1 Characterization of uranium redox state in organic-rich Eocene sediments

- 2
- 3 Susan A Cumberland^{1,2,3}, Barbara Etschmann², Joël Brugger², Grant Douglas⁴, Katy Evans⁵,
- 4 Louise Fisher⁶, Peter Kappen³, John W. Moreau¹
- 5 1 School of Earth Sciences, University of Melbourne, Parkville, Victoria 3100, Australia
- 6 2 School of Earth, Atmosphere and Environment, Monash University, Clayton 3800, Victoria,7 Australia
- 8 3 ANSTO Australian Synchrotron, 800 Blackburn Road, Clayton 3168, Victoria, Australia
- 9 4 CSIRO Land and Water, Floreat, Western Australia, Australia
- 10 5 Western Australian School of Mines, Curtin University, Bentley, Western Australia,11 Australia
- 12 6 CSIRO Mineral Resources, Bentley, Western Australia, Australia
- 13

14 Abstract

15 The presence of organic matter (OM) has a profound impact on uranium (U) redox cycling, 16 either limiting or promoting the mobility of U via binding, reduction, or complexation. To understand the interactions between OM and U, we characterised U oxidation state and 17 speciation in nine OM-rich sediment cores (18 samples), plus a lignite sample from the 18 19 Mulga Rock polymetallic deposit in Western Australia. Uranium was unevenly dispersed 20 within the analysed samples with 84% of the total U occurring in samples containing 21 >21 wt. % OM. Analyses of U speciation, including x-ray absorption spectroscopy and 22 bicarbonate extractions, revealed that U existed predominately (~71%) as U(VI), despite the 23 low pH (4.5) and nominally reducing conditions within the sediments. Furthermore, low

extractability by water, but high extractability by a bi-carbonate solution, indicated a strong association of U with particulate OM. The unexpectedly high proportion of U(VI) relative to U(IV) within the OM-rich sediments implies that OM itself does not readily reduce U, and the reduction of U is not a requirement for immobilising uranium in OM-rich deposits. The fact that OM can play a significant role in limiting the mobility and reduction of U(VI) in sediments is important for both U-mining and remediation.

30 Keywords: Mulga Rock, uranium, mobility, organic matter, oxidation state, x-ray
 31 absorption spectroscopy.

32 Highlights

- 84% of U in Mulga Rock OM-bearing sediments occurred within the OM-richest
 samples (>21% total organic carbon; 9 of 18 samples).
- 71% of U was present in the U(VI) oxidation state.
- Higher proportions of U occur as U(VI) in mature OM-rich sediments than previously
 thought.
- OM strongly complexes U(VI), limiting its mobility and reactivity.
- U reduction processes are not ubiquitous within OM-rich sediments.

40 1 Introduction

The mobility of uranium (U) in natural environments is controlled by its oxidation state and solubility, with the uranyl ion $(U^{VI}O_2^{2^+})$ generally more soluble than U(IV). Ligands such as $CO_3^{2^-}$, OH⁻, PO₄³⁻ and SO₄²⁻ can either increase (via formation of stable aqueous complexes) or decrease (via formation of insoluble minerals) U mobility (Cumberland et al., 2016). Uranium mobility is also influenced by organic matter (OM) in both dissolved and colloidal/particulate forms (Wood, 1996; Cumberland et al., 2016). Depending on pH, 47 dissolved OM such as humic and fulvic acids can bind U, facilitate its transportation, and 48 prevent its sorption to mineral surfaces (Luo and Gu, 2008; Zhao et al., 2012; Tinnacher et 49 al., 2013). In contrast, particulate or solid phase OM can adsorb and accumulate U, resulting 50 in the formation of OM-rich sedimentary U deposits (Greenwood et al., 2013; Cumberland 51 et al., 2016). While considerable information exists on U mobility in inorganic systems 52 (Grenthe et al., 2004), datasets for organic-rich systems are comparatively deficient (Bargar 53 et al., 2013). The complex molecular-scale associations of U and OM, and the impact of OM 54 on U mobility/immobilisation, are still being unravelled (Bargar et al., 2013). These 55 knowledge gaps limit development of in situ mining technologies (Zammit et al., 2014), as 56 well as environmental management and remediation strategies for U mining sites.

57 Recent studies have been directed towards understanding U mobility in OM-rich modern 58 sediments, including the pathways through which U potentially accumulates (Tokunaga et al., 2005; Law et al., 2011). Peat, for example, can concentrate U by up to factor of 10,000 59 from groundwater containing ppb level U concentrations (e.g. (Idiz et al., 1986; Read et al., 60 61 1993; Owen and Otton, 1995; Lidman et al., 2012)). Bryan et al. (2012) and Warwick et al. 62 (2005) identified U sorption to carboxylic and phenolic functional groups on insoluble 63 humics and fulvics as the likely scavenging mechanism. Furthermore, U accumulation in peat and other OM-rich environments is commonly conceptualised as involving initial U(VI) 64 adsorption followed by reduction to U(IV), with the latter process resulting in long-term U 65 66 immobilisation (Spirakis, 1996). A suite of factors may facilitate U reduction in sediments, 67 including increased temperature (Nakashima 1992), electron donors such as H_2S or Fe^{2+} , and 68 direct enzymatic reduction (Newsome et al., 2014; Campbell et al., 2015). Mineral surfaces 69 (e.g., pyrite; Fe-Ti-oxides) and insoluble OM may also play roles in reduction of sorbed U(VI),

70 catalysing rate-limiting reaction steps by facilitating electron transfer from underlying 71 material or co-adsorbates (Renock et al., 2013; Latta et al., 2014). Most U in fossil economic 72 U deposits, including OM-rich types (Min et al., 2000; Deditius et al., 2008), has been 73 assumed to be associated with U(IV)-minerals, such as coffinite, or other non-crystalline 74 forms (Bhattacharyya et al., 2017). Interestingly, synchrotron X-ray absorption spectroscopy 75 (XAS) data have revealed the presence of both U(IV) and U(VI) in some OM-rich U deposits, with U(IV) constituting of 35-68% of total U (Mikutta et al., 2016). Regenspurg et al. (2010) 76 77 found a mixture of U(VI) and U(IV) in alpine soils from the Dischma Valley (Switzerland), 78 based on combined carbonate extractions and data from X-ray Absorption Near Edge 79 Structure (XANES) analyses. These results point to U(VI) being an important component of 80 U bound to OM surfaces. From these recent findings, a picture is emerging for U-OM association in modern sediments that emphasizes the complexity and variability of U redox 81 82 state and chemical speciation (Mikutta et al., 2016; Bone et al., 2017).

83 Synchrotron x-ray absorption spectroscopy (XAS) provides a suitable analytical technique for 84 obtaining U-oxidation state and speciation in OM-rich sediments, since it is sensitive to both 85 'invisible' (e.g., colloidal; adsorbed; associated with OM, typically sub > 10 nm) and mineral-86 bound U (Denecke et al., 1998a; Denecke et al., 1998b; Mikutta et al., 2016; Bhattacharyya 87 et al., 2017). XAS can be applied to samples with a complex composition and containing a 88 wide range of U concentrations (from a few ppm to wt.%). Here our objective was to 89 characterise the physical and chemical nature of U within an OM-rich Eocene sedimentary U 90 deposit, Mulga Rock (Western Australia), in order to determine the relative abundance and 91 origins of observed U(IV) or U(VI). Our results inform strategies for sustainable mining

92 (subsurface extraction) or environmental remediation of U in OM-rich sedimentary or93 aqueous environments.

94 2 Geological setting and core sampling

95 The Mulga Rock (MR) U deposit is located near Kalgoorlie (Western Australia; Figure 1), 96 hosted within a series of paleochannels and lacustrine beds formed during the Eocene (56-34 Ma), now buried to shallow depths (30 to 50 m). According to the petrographic analyses 97 98 of Jaraula et al. (2015), MR sediments contain a mixture of particulate and non-particulate 99 OM consisting of woody (lignite) material combined with aquatic algal and bacterial 100 biomass, consistent with accumulation from a combination of forested, wetland and 101 lacustrine environments. Over time, the OM layers and deeper sediments became anoxic, 102 with considerable evidence for microbial production of authigenic minerals, most 103 prominently Fe-sulphides. Uranium accumulation likely resulted from groundwater flow 104 through permeable sandstones within palaeochannels, and scavenging of the carried U (present day average 8 µg L⁻¹ U in local groundwater) by OM-rich layers (Douglas et al., 105 106 2011; Jaraula et al., 2015). U-Pb isotope systematics and U-Th disequilibria studies suggest 107 that U and the associated metals originated from local lamproitic or carbonatitic sources 108 (Douglas et al., 2011; Jaraula et al., 2015).



110

To date, four orebodies with potential U resources have been delineated at MR: Ambassador, Emperor, Princess and Shogun (Figure 1). Ambassador, discovered in 1979, has been the main subject of industry and academic investigations, with resources estimated at 13,000 tonnes U (Douglas et al., 1993; Douglas et al., 2011). The most recent orebody, Princess, was discovered in March 2012. The wider MR deposit is estimated to contain 66.5 MT of ore at 520 mg kg⁻¹ U₃O₈ (Vimy, 2016c). The geological setting and close association of U with OM-rich sediments at MR make it an excellent site for studying the

121 nature of U in ore-grade OM-rich environments.

<sup>Figure 1. Map of Mulga Rock, deposit sites, core locations and detail from core 5613
showing lithology, oxidation state and radioactivity (obtained from down-hole gamma log
counts (Vimy, 2014)). See also SI Table 1 for locations.</sup>

We analysed 18 samples from nine air- and diamond-drilled cores obtained from the March 122 123 2012 drilling project: Ambassador (4), Emperor (1), Princess (3) and Shogun (1), together 124 with a section of lignite (CD 1577) extracted during a previous exploration campaign from the Ambassador deposit (Vimy, 2014, 2015, 2016a, b). Lignite sample CD 1577 was 125 126 immediately preserved in epoxy resin to prevent oxidation. The March 2012 samples were 127 selected on-site as they were extracted (sample locations given in Error! Reference source 128 **not found.**) and immediately stored in airtight containers, taking care to minimize exposure 129 to air and to leave no air gap. The samples were selected to represent a range of redox 130 states, mineralisation types and grades, as estimated on-site by a handheld Niton XL3t X-ray 131 Fluorescence (XRF) spectrometer, and on the radiometrics determined using a down-hole gamma probe (AusLog, #T125, 33mm diameter) calibrated with a Cs¹³⁷ source (Vimy, 2014, 132 133 2016b). Laboratory based geochemical analysis was completed 6-12 months after sampling 134 and preservation; synchrotron X-ray analysis was performed within two years of sample 135 preservation.

136 3 Geochemical Analysis

Six analytical methods were used: bulk core sample analysis (ICP-MS); chemical extractions
(ICP-MS); electron microscopy; and synchrotron-based XAS and X-ray fluorescence
microscopy (XFM).

140 3.1 Sample preparation and measurements of loss on ignition (LOI) and pH

Samples were characterised using fresh, oven-dried, ashed or dried-homogenised material. To obtain dry masses, sub-samples and replicates were oven-dried at least overnight at 104 °C and weighed until constant mass. The resulting oven-dried samples either underwent LOI analysis, or were homogenised to a very fine powder in an agate-lined ball mill and dry145 stored in glass vials at ambient temperatures until further analysis. The LOI analysis was 146 undertaken on a subset of samples (replicates = 5, n = 12) at 550 °C for 4 hr, then 900 °C for 147 2 hr, using a Thermoline muffle furnace (Dean, 1974; Heiri et al., 2001). A reduced number 148 of samples were analysed for LOI (12 of 18) to conserve material. Powdered samples were 149 analysed for total nitrogen (TN), organic carbon (TOC) and sulphur (TS) by an Elementar III 150 Elemental Analyser (University of Kiel) without further pre-treatment. Sediment pH was 151 analysed within 3 months of sample recovery using the method of (Rowell, 1994). 152 Measurements were conducted on duplicate samples using a Thermo portable meter with 153 standard calomel electrode. Six grams of undried sample was mixed with 15 ml of water, the 154 slurry was shaken by hand for 15 minutes, and the pH was then measured immediately and 155 again after 10 minutes.

156 3.2 Total and extractable metals including U

157 Total metals, U and rare earth elements (REE) concentrations were obtained from acid-158 digested samples using a method adopted from Kamber et al. (2003) and Eggins et al. 159 (1997). Both dried and ashed samples were analysed for U (SI Table 5). 2 mL of triple-160 distilled HF and 1 mL triple-distilled HNO₃ were added to 100 mg of sample in capped Teflon 161 vessels and digested at 135 °C overnight on a hotplate. The dehydrated material was 162 refluxed twice with concentrated HNO₃, then dissolved overnight in 3N HNO₃. The solutions 163 were transferred to transparent polycarbonate tubes, and diluted with water. Solutions 164 were diluted further to a factor of 5000 with a solution of 1.6 % HNO₃, then analysed using 165 inductively coupled plasma-mass spectrometry (quadrupole ICP-MS, Agilent 7700x). Results of digested samples are reported as mg kg⁻¹ of dry mass. There was a significant correlation 166 167 $(R^2 = 0.999, p < 0.005)$ between the ashed (n=12) and unashed samples, indicating confidence in that the organic matrix did not interfere with the ICP-MS analysis. 168

169 Bi-carbonate (HCO₃⁻) (Zhou and Gu, 2005) and pure water extractions (ultra pure, 18.2 MΩ) 170 were undertaken to further constrain U speciation and oxidation state on 18 sub-samples. 171 Typically, bi-carbonate extraction assumes that in the absence of oxygen, U(VI) forms strong complexes with carbonate ions and hence becomes highly soluble, whereas U(IV), including 172 173 unoxidised biogenic U(IV), remains uncomplexed and insoluble (Guillaumont et al., 2003; 174 Zhou and Gu, 2005). In some circumstances, monomeric (i.e., sorbed) biogenic U(IV) can be 175 released in concentrated bicarbonate solutions, without affecting uraninite/coffinite 176 stability (Alessi et al., 2012). Thus in our work the extracted U(VI) may also contain a small 177 fraction of mobile, highly reactive biogenic U(IV). The dissolved U(VI) can be separated by 178 filtration (i.e. 0.2 µm) of the aqueous phase, and the U(IV) concentration can be estimated 179 by subtraction of the amount of soluble U analysed by ICP-MS from the total U value. This 180 provides a minimum U(VI) concentration, since some U(VI) minerals, in particular silicates, 181 phosphates and vanadates are sparingly soluble (Tokunaga et al., 2012; Kanematsu et al., 182 2014; Mehta et al., 2014). Bi-carbonate extractions were performed in duplicate under 183 anaerobic (N₂) conditions as follows: 20 ml of 1 M NaHCO₃ (de-gassed) was added to ~1.7 g 184 fresh hand-ground sediment; pH was recorded; and the sample was placed in 100 mL crimped and sealed serum bottles (acid washed, 10% HNO₃) with the headspace replaced 185 186 with N₂ gas to avoid oxidation. Blanks were prepared as above. The suspensions were rotated on an orbital shaker (150 rpm, 25 °C) continuously for one week, and then settled 187 for 10 minutes, after which they were vacuum-filtered through 0.2 µm cellulose nitrate 188 189 membranes (Whatman). Filtration was generally fast, with additional filters used where 190 necessary. Filtrate solutions were acidified (1% HNO₃) and then analysed by ICP-MS. The pH 191 of the bi-carbonate/sediment suspension was 8.5 ± 0.1 . Water extractions were performed

under the same conditions but filtered using a standard 1.2 μm GF/C Whatman glass filter
paper to include a potentially mobile U colloidal fraction.

194 3.3 Imaging U distribution and elemental association using XFM and SEM

195 To determine the microscale spatial distribution and elemental/mineralogical association of 196 U, subsamples of unground grains from core NNA 5613 (Ambassador ore body) were 197 selected based on total activity (measured by Geiger counter) and mounted onto 0.5 cm² 198 SiN windows using a 5% glucose solution. Due to the finely dispersed nature of the U in the 199 MR samples, TEM techniques proved unsuccessful in detecting U in the samples. One 200 sample, MR5613a, yielded better SEM results and gave higher radioactive counts per 201 second, likely due to higher U (5000 mg kg⁻¹) concentrations. This sample was therefore favoured for subsequent micro-analytical characterisation. 202

203 Subsample grains from MR5613a were analysed at the XFM Beamline (Australian 204 Synchrotron) at 18.5 keV incident photon energy and 2 μm beam size (Paterson et al., 205 2011). Fluorescence signals were collected using a 384-pixel Maia detector (Ryan et al., 206 2014). The fluorescence spectra collected were processed using the GeoPIXE software to 207 produce semi-quantitative elemental maps (Ryan and Jamieson, 1993; Ryan, 2000; 208 Etschmann et al., 2010; Li et al., 2016).

Backscatter electron images (sample MR5613a from core NNA 5613) were obtained on thin
sections using a Philips FEI XL30 environmental scanning electron microscope (ESEM)
equipped with an OXFORD INCA energy-dispersive X-ray spectrometer (EDS) (Earth Sciences,
University of Melbourne), and EDS maps were collected with a Field Emission Philips XL30
ESEM (BIO21, University of Melbourne).

214 The mineralogy of lignite sample CD 1577 was characterised using µ-XRD. Spectra were 215 collected using a Bruker general area detector diffraction system (GADDS) with Cu K_{α} 216 radiation and 200 or 300 µm collimators (CSIRO, Clayton, Melbourne; SI Figure 1).

217

3.4 Uranium speciation via XANES and EXAFS spectroscopy.

218 X-ray absorption near edge structure (XANES) and extended x-ray absorption fine structure 219 (EXAFS) spectra were collected at the U-L₃ absorption edge at the XAS Beamline (Australian 220 Synchrotron). Spectra were collected for eight powdered core samples (MR5076a, 221 MR5076b, MR5077, MR5613a, MR5613b, MR5766b, MR5766c, MR5766d,) selected because 222 they had sufficient U concentrations (from ICP-MS results) to yield good quality spectra, and 223 nine U-bearing minerals provided by the Melbourne Museum serving as standards. Data on 224 these standards are presented in the supplementary information (Error! Reference source 225 not found., SI Figure 2 and SI Figure 3). Homogenised powdered samples or mineral 226 standards diluted with boron nitride were packed into Al holders, and sealed with Kapton 227 tape. For analysis, samples were loaded into a He cryostat (pulse-tube Optistat, Oxford 228 Instruments; T~10K). In addition, spectra were taken from a fresh section of the core 229 subsample CD 1577 at six random points. The incident energy was controlled using a Si(111) 230 double crystal monochronomator and calibrated at the Zr-K absorption edge using a Zr 231 metal foil (first maximum of the first derivative at 17998 eV). Repeat measurements of this 232 Zr foil established that there was no measurable energy shift during the experiments. 233 Uranium was detected concurrently in fluorescence (100 element Canberra HP-Ge 234 fluorescence detector) and transmission modes using ion chambers (Oken, N₂ flow at ~0.3 235 L/min, U = 2.1 kV). Aluminium foil was used to reduce parasitic fluorescence count rate from 236 other fluorescence. Radiation induced changes to U valence or chemistry, assessed over repeat XANES runs, were negligible. The XANES scans were acquired as follows: baseline, 237

238 17010 to 17141 eV in 3 eV steps; over the edge-step, 17144 to 17204 eV in 0.3 eV steps; and 239 with constant steps in k-space ($\delta k = 0.035 \text{ Å}^{-1}$) above the edge. For EXAFS measurements the 240 energy range was extended to ~18140 eV (k~16 A⁻¹).

Spectra normalisation and linear combination fitting (LCF) were performed using the DEMETER software package (Ravel and Newville, 2005). Most spectra were truncated after 17969 eV (14 Å⁻¹) to avoid interference from the Zr K edge. The LCF was performed using two reference spectra, synthetic uraninite for U(IV) (Tsarev et al., 2016) and uranopilite for U(VI) (see also SI Figure 2). The uranium oxidation state was determined from the LCF data.

The EXAFS data were processed and refined using the HORAE package (Ravel and Newville, 2005) with the theoretical standards calculated by FEFF9 (Rehr et al., 2010). The k^{n} weighted data (n = 1, 2, 3) were refined in *R* space. The amplitude-reduction factor (S₀²) was estimated to be 0.85 by fitting UO₂(NO₃)₂•6H₂O (structure confirmed by powder XRD; SI Table 3 and **Error! Reference source not found.**); this value was used in the refinement of all the sample data.

252 4 Results

253 4.1 Core geochemistry

Core sample appearance varied from sandy and dry, to dark, peaty and moist, with a mean moisture content of ~12 wt. % (range 0.3 to 38 wt. %). Mean sample pH was 4.5 (range 3.5 to 7). The physicochemical data are presented in SI Table 4. The TOC concentrations of the samples averaged 22.5 wt. % (range 1 to 59 wt. %), reflecting substantial vertical *in situ* variation in the MR deposit. Samples LOI₅₅₀ ranged from 1 to 57 wt. %, and correlated with TOC (R^2 = 0.922, slope = 1.04, p < 0.005; n = 10). Further LOI performed at 900 °C (LOI₉₀₀)

- 260 ranged from 1 to 78.3 wt. % (average = 31 wt. %). This mass loss between 550 and 900 $^{\circ}$ C
- 261 can be attributed to decomposition of inorganic carbonates (Dean, 1974; Heiri et al., 2001).



264

265 Figure 2. Uranium distribution and correlations with organic matter and trace elements. (A) Bar chart showing uranium distribution in samples according to carbon content. The 266 267 data (n = 18) are split on the basis of their carbon contents (TOC) into two equal groups 268 (n = 9 for each group); one group has < 21 wt. % TOC, the other 22-60 wt. % TOC. Error bars 269 are standard deviations of uranium concentration: samples containing < 21% TOC contain 270 1600 mg U kg⁻¹ and those with 22-60 wt. % TOC 4500 mg U kg⁻¹. (**B**) Pearson correlation 271 coefficients for all core samples (n = 18); (B-1) between TOC and metals; and (B-2) U and 272 metals, where height of bar corresponds to R² values. Stars indicate statistical significance: 273 *** p < 0.005; ** p < 0.05; * p < 0.5, ns (no star) p > 0.51. Note that Fe was not analysed in 274 these samples. (C and D) Detail of sediment grains from the MR core NNA 5613: (C-275 1) Scanning Electron Microscope (SEM) image; (C-2) is enlargement of (C-1), showing U-rich 276 and Zn-rich particles of similar sizes. (D) Synchrotron x-ray fluorescence (XFM) image 277 showing U, Cu and Sr distribution in representative grain from the same sample as (C); U = 278 red, Cu = green and Sr = blue.

279 [full colour figure in print]

Uranium in the core samples ranged from 2 to 15,000 mg kg⁻¹ U (mean = 2800 mg kg⁻¹, n = 281 282 18), and correlated with chemical digests performed on ashed samples (900 °C, n = 12) when normalised to mg kg⁻¹ dry mass ($R^2 = 0.998$, slope = 1.00, p < 0.0005; Pearson) (see SI 283 284 Table 5). Core samples contained average element concentrations in the order 285 Ti>U>Zn>Ca>total REE>Cu>Ni>Pb>Cr>Co>Sc>V>As>Th (SI Table 6). Correlations between U and TOC, and U and other metals were weak (p < 0.5), with the strongest relationships 286 occurring between U and Pb, U and total REEs, Cd and As (p < 0.005, R² values shown in 287 288 Figure 2B and SI Table 7). Slight correlations were observed between TOC and Ca, Sr or V (p < 0.05), but most elements show no significant correlation with TOC. Detailed analyses on 289 290 the organic chemistry of samples from the Mulga Rock deposit are presented elsewhere 291 (Jaraula et al 2015).

292 The spatial distribution of U in the core samples was highly variable, with U present either in 293 diffuse form (no identifiable minerals > 10 nm) or as U-rich micro- to nano- particles (see 294 XFM and SEM images in Figure 2 C and D). Sample MR5613a (from core NNA 5613, Figure 2 D), with a bulk U concentration of 5,000 mg kg⁻¹, showed substantial variation in the form 295 296 and distribution of U between the four subsamples scanned. Two subsamples contained 297 disseminated U, while U was present as discrete µm-sized particles in the other two. The 298 lignite sample CD 1577 was structurally more coherent than the March 2012 cores. The 299 mineral matrix consisted mainly of quartz (see Error! Reference source not found. for µXRD 300 analysis). SEM elemental mapping and BSE imaging (Figures 3, 4 and SI Figure 4 for EDS 301 spectrum) show that U exists in particulate form (μ m-sized), but also forms large (up to 302 >1 mm) aggregates around Fe-sulfide minerals. The U-pyrite association is common at MR 303 (Douglas et al., (2011) as well as in many sandstone-hosted U deposits (e.g., Bonnetti et al.

304 (2015; 2017); Wülser et al., (2011)), and probably reflects an association between bacterial 305 sulphate reduction and U reduction (Bonnetti et al., 2017). Micro-XRD data confirm the 306 predominance of pyrite, and show that the main U-phase surrounding the pyrite 307 grains/aggregates is coffinite [USiO₄] (**Error! Reference source not found.**). The MR coffinite 308 has unit cell parameters (a = b = 6.943(2) Å; c = 6.262(2) Å; V = 301.9(3) Å³) that are close to 309 those of the Tertiary sandstone hosted Beverly deposit in South Australia (a = b = 6.971(2) Å; 310 c = 6.255(2) Å; V = 304.0(3) Å³; Wülser et al. (2011)).





Figure 3. Chemical maps (U, Al, C, O, Na, S and Si) based on SEM-Energy Dispersive Spectrometry (EDS) of lignite sample CD 1577. Elemental concentrations are represented by intensity, where bright colours show high concentrations and dark or black areas represent areas where elements are absent or at low concentrations. Sulfur corresponds to pyrite

- 316 distribution. Si is associated with U (coffinite according to XRD results), bright spots
- 317 correspond to quartz. Scale bar is $100 \ \mu m$.
- 318 [full colour figure in print]
- 319



320

Figure 4. Composite ESEM backscatter image, showing distribution of uranium in the lignite section from core CD 1577. Uranium is present throughout the whole sample with the brightest phase illustrating the highest U concentration. The EXAFS spectra were taken at random points which showed that U was present in both U(IV) and U(VI) oxidation states across the sample. Inserts (B) and (C) are enlargements of indicated areas in (A).

327 4.2 Leach tests and XANES linear combination fits

328 In all samples analysed, U was present in both U(IV) and U(VI) oxidation states, with usually 329 more U(VI) than U(IV). Bicarbonate extraction of U (U_{CO3}), a process conducted under 330 anaerobic (N₂) conditions which targets leachable U(VI) ions, indicated that an average of 331 69% (SD = \pm 18, n = 18) of total U is in the U(VI) form. The oxidation state of U (U(IV) or U(VI)) and total concentrations of U are highly correlated ($R^2 = 0.996$, p < 0.005) with the 332 333 percentage of U(VI) increasing with total U concentration, giving a slope of 0.79 for U(VI) versus total U (Figure 5A). Figure 5B shows that samples with low TOC (<8%) have a highly 334 335 variable oxidation state, whereas samples with high TOC (i.e. > 20%) tend to have a

relatively constant oxidation state, dominated by U(VI) (75 \pm 7% U(VI)). Interestingly, extractions using only deionised (ultra-pure, 18.2 M Ω) water under the same anaerobic conditions as bi-carbonate method, only removed 2.5 % of the total U, showing that the majority of U was not readily water-soluble.

The U oxidation state (U_{XAS}) was independently derived from the LCF results as described above (see also **Error! Reference source not found.** and Figure 5 C and D for the normalized XANES spectra). The LCF data (**Error! Reference source not found.**) show that the uranium is predominately present (i.e. 74 %) in the higher oxidation state, +6. These results are consistent with the concentration values from CO₃₋extraction. Comparison of XAS (74%) and bi-carbonate extraction datasets (69%) showed good agreement, giving an overall mean across the two methods of 71% U(VI).



348

Figure 5. Uranium speciation as derived from bicarbonate extraction (A and B); and 349 350 synchrotron-XANES analyses (C and D). (A) Bi-carbonate extracted U(VI) and U(IV) data in mg kg⁻¹ (y axis) plotted as a function of total U (mg kg⁻¹). Black squares are U(VI) ($R^2 = 0.996$. 351 y = 0.7896x - 49.5 for U(VI)) and red dots are U(IV) (R² = 0.21 x 49.5). (B) U(IV) and U(VI) 352 oxidation states as percent of total U (U(VI)/(U(VI)+U(IV)) x 100 plotted as a function of TOC 353 354 wt. %. Black squares =U(VI) ,and red dots =U(IV). (C) XAS-XANES spectra of MR bulk core samples and the lignite sample CD1577. The overlain dotted line is the fitted spectra from 355 356 LCF The arrow marks the position of the O=U=O U(VI) shoulder. (D) are the derivatives of 357 (C).

358

359 **4.3 EXAFS fits**

360 Two types of XANES spectra were obtained from six measurements performed on the lignite

361 sample CD 1577 (Figure 4). These spectra were grouped, merged and analysed separately,

362 with one type showing a uranyl-like ('oxidised-CD1577') spectrum and the other a reduced

363 ('reduced-CD1577) U spectrum (Figure 6). The EXAFS data also differ greatly among the two
364 types (Figure 6), and different fitting strategies were used for each type. EXAFS data from
365 the samples MR5766c and MR5613a are similar to the 'oxidised-CD1577' lignite spectra, and
366 were refined simultaneously.

367 4.3.1 Fitting of the EXAFS spectra with uranyl-like XANES

EXAFS spectra from MR5766c, MR5613a and 'oxidised-CD1577' were analysed with a 368 369 common E₀, using selected paths from the uranopilite structure (amcsd 0005730; Burns 370 (2001)), and one uranium-carbon single-scattering path from the structure of C₁₅H₁₀O₈U 371 from de Lill and Chan (2013). The advantage of this approach is that the 'oxidised-CD1577' 372 spectra could be fitted by constraining the distances and Debye-Waller factors to be the 373 same as those for MR5766c. To obtain the best fit for the 'oxidised-CD1577' spectra, the 374 number of oxygen atoms (O15 only) was refined, allowing additional oxygen to be fitted 375 that was not present in the MR5766c and MR5613a spectra. Note that the choice of C rather than another light element (e.g. O, N, P) cannot be justified by EXAFS alone. Carbon was 376 377 chosen because the refined bond-distance and coordination are consistent with previous 378 studies (see below); SEM did not reveal a correlation between U and P, or any of the poorly 379 soluble uranyl phosphates minerals; and the samples are rich in C but poor in N (C/N≥100; 380 Jaraula et al. 2015). The final results are shown in Table 1.

Two models were used for comparison: a simple model where the two equally short U-O distances (uranyl ion) were constrained to be the same (Table 1); and a second model where the two short U-O distances were refined independently (SI Table 9). While more complex multiple scattering (MS) paths (U-O_{ax2}-U-O_{ax2}, U-O_{ax2}-O_{ax1} and U-O_{ax2}-U-O_{ax1}) could be employed to fit the data out to a distance of 3.58 Å in the second model, the two short U-O

bonds were the same within error (1.75(3) and 1.77(3) Å). Furthermore, the goodness-of-fit parameters for both models were effectively identical, indicating that there was no significant advantage to the second model, so that the simpler model (Table 1) is preferred.

In all cases, the fit resulted in two carbon atoms at distances of 2.90(7) to 2.93(3) Å. This 389 390 configuration suggests a bidentate coordination to functional groups of the solid OM. All U(VI)-dominant spectra lacked a peak at distance of 3.8 Å, which would indicate the 391 392 presence of U-U bonds. This shows that most U is not present in the form of uranyl 393 minerals. Consequently, the most U(VI) in the studied samples is present predominantly as a 394 monomeric U(VI) complex bound to OM. This is further supported by the flattening of the oscillation between 6.8 and 7.9 Å⁻¹ in the k^2 -weighted EXAFS data (arrow, Figure 6A), a 395 396 feature which appears in uranyl-humic compounds (Denecke et al., 1998a; Denecke et al., 397 1998b; Mikutta et al., 2016).

somelo	licend	N	D (A)	oo (A2)	fraction	N _{tot} =	A.E.	red V2	r fastar	k-range		k waishting	60 2
sample	ligano	IN	к (А)	55 (A-)	Traction	Niraction	ΔE_0	red X-	r-lactor	(A -)	K-range (A)	k-weighting	30-
	Oax2	2	1.773(5)	0.003 (fix)			9.2(5)	113	0.023	2 - 12	1.3 - 4	1,2,3	0.85
MR5766c	015	2	2.22(1)	0.003 (fix)									
	014	2	2.36(1)	0.003 (fix)									
	OH20	2	2.42(2)	0.003 (fix)									
	С	2	2.93(3)	0.003 (fix)									
	U-Oax2-U-Oax2 (MS)		3.545(7)	0.006 (fix)									
	Oax2	2	1.77(2)	0.003 (fix)			11(3)	113		3 - 11.5	1.3 - 4	1,2,3	0.85
MR5613a	015	2	2.24(4)	0.003 (fix)									
	014	2	2.39(4)	0.003 (fix)									
	OH20	2	2.45(6)	0.003 (fix)									
	С	2	2.90(7)	0.003 (fix)									
	U-Oax2-U-Oax2 (MS)		3.55(3)	0.006 (fix)									
oxidised-CD1577*	Oax2	2.3(2)	1.773(5)	0.003 (fix)			9(1)	113		2.5 - 12	1 - 4	1,2,3	0.85
	015	2.8(4)	2.22(1)	0.003 (fix)									
	014	2	2.36(1)	0.003 (fix)									
	OH20	2	2.41(2)	0.003 (fix)									
	С	2	2.91(4)	0.003 (fix)									
	U-Oax2-U-Oax2 (MS)		3.545(7)	0.006 (fix)									
Reduced-CD1577	Oax2	2**	1.74(2)	0.003 (fix)	0.47(10)	0.94	8(3)	199		3 – 10.5	1 - 4	1,2,3	0.85
	015	2**	2.35(7)	0.003 (fix)	0.47(10)	0.94							
	014	2**	2.33 (fix)	0.003 (fix)	0.47(10)	0.94							
	OH20	2**	2.37 (fix)			0.94							
	O1 (coffinite)	2**	2.25(3)	0.003 (fix)	0.53(10)	1.06							
	O2 (coffinite)	4**	2.83(3)	0.003 (fix)	0.53(10)	2.12							
	U1 (coffinite)	4**	3.79(4)	0.005(4)	0.53(10)	2.12							

399 Table 1 Fitting parameters obtained from U L₃-Edge of EXAFS Spectra.

400 * These distances were constrained to be the same as those of MR5766c and were fitted together.

401 ** The number of ligands was fixed, the fraction of each mineral component was refined.



402 403

Figure 6 EXAFS data and fits, shown in (A) k-space; (B) R-space; (C) R- Imaginary. The uranyllike spectra are shown at the top of each graph for MR5766c, MR5613a and CD1577
oxidised spectra (ox-MR) with relevant paths shown below. The reduced spectra (red-MR)
from CD1577 is shown below in Red and relevant paths underneath.

408

409 4.3.2 Fitting of the reduced-CD1577 EXAFS spectra

410 The spectrum of the reduced component from the lignite (reduced-MR, CD 1577) was fitted 411 using a wide variety of models. We report here a simple model (SI Table 3) based on a 412 combination of paths corresponding to uranopilite $[(UO_2)_6(SO_4)O_2(OH)_6(H_2O)6\cdot 8H_2O]$, as a proxy for the uranyl component (amcsd 0005730; (Burns, 2001); and coffinite [USiO₄] (icsd 413 414 15484; (Fuchs and Gebert, 1958), which is the main U mineral according to micro-XRD (SI 415 Figure 1). The best fit was obtained using 47(10) % uranopilite and 53(9.8) % coffinite. The presence of atoms at ~3.8 Å distance indicates a second shell of U consistent with the crystal 416 structure of coffinite (EXAFS fit U-U distance of 3.79(4); crystallographic distance of 417 418 3.83(2) Å). Furthermore, Bhattacharyya et al. (2017) report a U-C distance of 2.90 Å for their 419 organic U(IV) structure, but this contribution is absent from the reduced-MR sample. Thus,

420 in contrast to U(VI), which is present in mostly monomeric form within the organic matrix,

421 U(IV) is present in mineral form (predominantly coffinite in CD 1577).

422 **5 Discussion**

423 5.1 U-OM association at MR

Accumulation of U by OM by 10's of 1000's of times relative to U-concentrations in 424 groundwater is a common and long established feature. Analyses of groundwater by 425 Douglas et al. (2011) show that modern waters at MR contain ppb concentrations of U 426 427 (8 μ g L⁻¹), hence implying concentrations of up to > 6 orders of magnitude in the OM layers 428 (mean = 2500 mg kg⁻¹ U, max = 15,000 mg kg⁻¹ U). The highest concentrations of U at MR 429 were associated with high concentrations of organic carbon, with 84% of the total U in half 430 the sample set (n=9) containing >22-60 wt. % TOC. According to Douglas et al. (2011), 431 uranium and the associated metals precipitated syngenetically with OM as it was deposited 432 during a humid phase in the Late Eocene. Subsequent small-scale mobilisation during 433 diagenesis and/or weathering concentrated the metals in the upper 2 m of the lignite, with 434 the latest episode in the last 300,000 years. Continuous upgrading via interaction with groundwater containing low concentrations of dissolved U could also have contributed to 435 436 the high grades observed today.

A characteristic of the OM-rich layers is that the proportion of U(VI) is relatively consistent
(e.g. 77.2% ± 18 from extractions), despite the fact that absolute U concentrations vary
greatly. Correlation coefficients between U and OM in soils and sediments in the literature
are highly variable (Meunier et al., 1989; Landais, 1996; Min et al., 2000; Regenspurg et al.,

2010; Cumberland et al., 2016; Och et al., 2016) and Mulga Rock is no exception. At MR the OM-rich samples contained most of the U identified across all the studied samples, yet no statistically significant correlation existed between grade and OM content. We infer that U introduced in oxidized form (uranyl complexes) by groundwater flowing through the OMrich sediments sorbed strongly and immediately to the available sites on the organic carbon.

446 We examined MR samples for U oxidation state, as this parameter controls the potential for 447 U mobility and biotoxicity, as well as providing necessary data to inform design of the ore-448 forming process. The results of both XAS and HCO₃-chemical extractions show that on 449 average 71% of the total U (mean of 69% HCO₃ extraction and 74% XAS LCF) was in the U(VI) 450 oxidation state in the MR samples. That most of the U was observed as U(VI) was surprising 451 given the long-held assumption that for long-term U immobilisation and accumulation to 452 occur within OM-rich sediments, reduction would be more important than adsorption (Spirakis, 1996). In the case of MR specifically, the presence of high proportions of reduced 453 454 U might also have been expected, given the low to neutral pH (3.5 to 7), reducing 455 conditions, and abundant coffinite/uraninite particles within the ores (Douglas et al., 2011). 456 However, while U phase stability modelling at low pH and Eh predicts that uraninite (UO₂) 457 and coffinite (under quartz-supersaturation conditions; Brugger et al., (2005)) are the dominant phases in the absence of OM (Langmuir, 1978), models that include OM as 458 459 humate suggest a U(VI)-humate phase dominating at ~pH 6 (Shanbhag and Choppin, 1981; 460 Lenhart et al., 1997; Cumberland et al., 2016).

Recent studies have shown that modern OM-rich soils contain a high proportion of U(VI) (32
-65 % U(VI) in peats, and 51 to 100 % U(VI) in organic-rich alpine soils; (Regenspurg et al.,
2010; Mikutta et al., 2016). Hence, the predominance of U(VI) at MR might be explained by

short U residence times and slow reduction kinetics. However, Jaraula et al. (2015) showed 464 465 that U has been present at MR long enough to cause radiolytic damage to organic 466 biomarkers. We therefore interpret that since the first sediments were deposited, U(VI) 467 reduction has not played a major role in immobilizing U, despite the fact that MR is a 468 shallow (<50 m) deposit undergoing deep weathering, and has remained in contact with groundwater containing µg L⁻¹ levels of U. Furthermore, the ores exist in an environment 469 where both inorganic and microbial redox processes are expected to be promoted - as 470 471 indeed demonstrated locally by the abundance of biogenic sulphide minerals.

472 **5.2** Uranium - organic matter complexes

The poor capability of pure water to extract U compared to the bicarbonate leach (2.5% versus 69% extraction) suggests poor U(VI) solubility and strong U(VI) association to particulate OM (i.e. > 1.2 μ m filter-size) or possibly finely dispersed silica. We therefore postulate that particulate OM has helped to retard U mobility at MR, especially since only very low concentrations of U persisted within neighbouring sandstone sediments that were lower in organic carbon.

Mikutta et al. (2016) reported that uranyl present in modern peats (e.g. <13,000 years) is 479 surrounded by 0.9 to 2.0 carbons at 2.89-2.93 Å. Therefore the speciation in the MR lignite, 480 481 where we identified a bidentate C coordination at similar distance (2.90(7) to 2.93(3) Å) 482 appears to be surprisingly similar to that in modern peat. This U-carbon model is also consistent with that of Denecke et al. (1998a; 1998b), representing poorly-ordered UO2²⁺ 483 484 bonding to COOH⁻ (carboxylic acid) groups present in organic macromolecules (Kaplan et al., 485 2016). Such bonding is likely at MR, considering that Douglas et al. (2011) found that up to 486 13.1% of the carbon is present in carboxyl groups within the MR OM-rich sediments.

Bhattacharyya et al. (2017) showed that uranium could remain attached to these carboxyl ligands even if U was reduced *in situ* to U(IV); however at MR, U reduction appears to be mainly associated with the formation of coffinite and/or uraninite particles.

490 6 Conclusion

491 Our findings show that uranium at Mulga Rock was mostly (85%) associated with the 492 organic-rich sediments, where most of it is present in the oxidised state of U(VI) (71% of 493 total U). It is commonly assumed that the flow of groundwater largely determines the 494 distribution of the U; however, this study found that U was associated with higher organic 495 matter concentrations, and therefore the mobility of U can be inhibited by OM. Results 496 from EXAFS analyses on the oxidised CD1577 spectra are consistent with the interpretation 497 that uranium predominantly forms a bidendate U(VI) complex with two organic (likely 498 carboxyl) groups.

499 Despite the dominant presence of OM, U(VI) was found to be reduced only incompletely, 500 even under low pH conditions. Hence, MR-OM itself did not readily reduce U, and reduction 501 of U was not required for immobilisation of uranium over geological time frames as 502 applicable to Mulga Rock. Tetravalent U, which displayed consistency with a coffinite-like 503 structure, was found to be widely present at Mulga Rock; these species most likely result 504 from *in situ* U(VI) reduction as a consequence of heterogeneous secondary processes, such 505 as biomediated sulfate reduction and pyrite mineralisation and/or microbial enzymatic 506 activity.

507 Our results extend a growing body of evidence that U speciation in OM-rich geological 508 materials involves a complex mixture of U(IV) and U(VI), representing potentially

509 overprinting generations of U delivery by groundwater and subsequent biologically-510 mediated U(VI) reduction. These results have direct implications for the development of U-511 OM chemical extraction methods and post-mining management practices: In the context of in-situ recovery, oxidation is not necessarily required for mobilising U, as U(VI) is present in 512 513 higher proportions than previously assumed. In the case of remediation, environmental 514 management strategies can shift from a focus on U reduction to using OM to capture U in its 515 uranyl form, as a potentially equally effective or preliminary approach to U remediation. 516 Future research into different forms of environmental U-OM interactions, in the presence of 517 other potentially reactive components such as mineral surfaces and microorganisms, is 518 necessary, to understand the conditions under which U can be released or immobilised in 519 OM-rich waters and sediments.

520 7 Acknowledgements

521 This work was funded through a grant to JWM from the CSIRO Organic Geochemistry of 522 Mineral Systems (OGMS) Research Cluster. Parts of this research were undertaken on the 523 XAS and XFM beamline at the Australian Synchrotron, part of ANSTO. We would like to 524 thank: Xavier Moreau (Vimy Resources) for MR samples; Dr Alan Grieg, Graham Hutchinson 525 and Dr Hong Vu for laboratory assistance; Prof. Lorenz Schwark (University of Kiel) for TOC, 526 N and S data analysis; Dr Martin De Jonge and Dr Daryl Howard for beamline assistance at 527 AS XFM, Museum Victoria for comparative minerals, and Steven Henderson for his editorial 528 contributions. SC would like to thank the Australian Synchrotron (AS) for beamtime and the 529 AS and Monash University for support in preparation of this manuscript through a joint 530 post-doctoral fellowship.

531

532 8 References

- Alessi, D.S., Uster, B., Veeramani, H., Suvorova, E.I., Lezama-Pacheco, J.S., Stubbs, J.E., Bargar, J.R.,
 Bernier-Latmani, R., 2012. Quantitative Separation of Monomeric U(IV) from UO2 in Products of
 U(VI) Reduction. Environmental Science & Technology 46, 6150-6157, doi:10.1021/es204123z.
- Bargar, J.R., Williams, K.H., Campbell, K.M., Long, P.E., Stubbs, J.E., Suvorova, E.I., Lezama-Pacheco,
 J.S., Alessi, D.S., Stylo, M., Webb, S.M., Davis, J.A., Giammar, D.E., Blue, L.Y., Bernier-Latmani, R.,
 2013. Uranium redox transition pathways in acetate-amended sediments. Proceedings of the
- 540
 National Academy of Sciences 110, 4506-4511, doi:10.1073/pnas.1219198110.
- Bhattacharyya, A., Campbell, K.M., Kelly, S.D., Roebbert, Y., Weyer, S., Bernier-Latmani, R., Borch, T.,
 2017. Biogenic non-crystalline U(IV) revealed as major component in uranium ore deposits. Nature
 Communications 8, 15538, doi:10.1038/ncomms15538.
- 544 Bone, S.E., Dynes, J.J., Cliff, J., Bargar, J.R., 2017. Uranium(IV) adsorption by natural organic matter in 545 anoxic sediments. Proceedings of the National Academy of Sciences 114, 711–716, 546 doi:10.1073/pnas.1611918114.
- Bonnetti, C., Cuney, M., Michels, R., Truche, L., Malartre, F., Liu, X., Yang, J., 2015. The Multiple Roles
 of Sulfate-Reducing Bacteria and Fe-Ti Oxides in the Genesis of the Bayinwula Roll Front-Type
 Uranium Deposit, Erlian Basin, NE China. Economic Geology 110, 1059-1081.
- Bonnetti, C., Liu, X., Zhaobin, Y., Cuney, M., Michels, R., Malartre, F., Mercadier, J., Cai, J., 2017.
 Coupled uranium mineralisation and bacterial sulphate reduction for the genesis of the Baxingtu
 sandstone-hosted U deposit, SW Songliao Basin, NE China. Ore Geology Reviews 82, 108-129,
 doi:10.1016/j.oregeorev.2016.11.013.
- 554 Brugger, J., Long, N., McPhail, D.C., Plimer, I., 2005. An active amagmatic hydrothermal system: The 555 Paralana hot springs, Northern Flinders Ranges, South Australia. Chemical Geology 222, 35-64, 556 doi:10.1016/j.chemgeo.2005.06.007.
- 557 Bryan, N.D., Abrahamsen, L., Evans, N., Warwick, P., Buckau, G., Weng, L., Van Riemsdijk, W.H., 558 2012. The effects of humic substances on the transport of radionuclides: Recent improvements in 559 the prediction of behaviour and the understanding of mechanisms. Applied Geochemistry 27, 378-560 389, doi:10.1016/j.apgeochem.2011.09.008.
- 561 Burns, P.C., 2001. A new uranyl sulfate chain in the structure of uranopilite. The Canadian 562 Mineralogist 39, 1139-1146.
- 563 Campbell, K.M., Gallegos, T.J., Landa, E.R., 2015. Biogeochemical aspects of uranium mineralization,
 564 mining, milling, and remediation. Applied Geochemistry 57, 206-235,
 565 doi:10.1016/j.apgeochem.2014.07.022.
- 566 Cumberland, S.A., Douglas, G., Grice, K., Moreau, J.W., 2016. Uranium mobility in organic matter-rich
- sediments: A review of geological and geochemical processes. Earth-Science Reviews 159, 160-185,
 doi:10.1016/j.earscirev.2016.05.010.
- de Lill, D.T., Chan, B.C., 2013. Structure and luminescence of a one-dimensional uranium
 coordination polymer assembled through benzophenone-4,4'-dicarboxylate. Inorganica Chimica Acta
 404, 215-218, doi:https://doi.org/10.1016/j.ica.2013.03.006.
- 572 Dean, W.E., 1974. Determination of carbonate and organic matter in calcareous sedimentry rocks by
- 573 loss of ignition:Comparison with other methods. Journal of Sedimentary Petrology 44, 242-248,
- 574 doi:doi:10.1306/74D729D2-2B21-11D7-8648000102C1865D
- 575 Deditius, A.P., Utsunomiya, S., Ewing, R.C., 2008. The chemical stability of coffinite, USiO4.*n*H(2)O;
- 576 0<n<2, associated with organic matter: A case study from Grants uranium region, New Mexico, USA.
- 577 Chemical Geology 251, 33-49, doi:10.1016/j.chemgeo.2008.02.009.
- 578 Denecke, M.A., Reich, T., Bubner, M., Pompe, S., Heise, K.H., Nitsche, H., Allen, P.G., Bucher, J.J.,
- 579 Edelstein, N.M., Shuh, D.K., 1998a. Determination of structural parameters of uranyl ions complexed

- 580 with organic acids using EXAFS. Journal of Alloys and Compounds 271, 123-127, doi:10.1016/s0925-581 8388(98)00038-3.
- 582 Denecke, M.A., Reich, T., Pompe, S., Bubner, M., Heise, K.H., Nitsche, H., Allen, P.G., Bucher, J.J.,
- 583 Edelstein, N.M., Shuh, D.K., Czerwinski, K.R., 1998b. EXAFS investigations of the interaction of humic
- 584 acids and model compounds with uranyl cations in solid complexes. Radiochimica Acta 82, 103-108.
- 585 Douglas, G.B., Butt, C.R.M., Gray, D.J., 2011. Geology, geochemistry and mineralogy of the lignite-586 hosted Ambassador palaeochannel uranium and multi-element deposit, Gunbarrel Basin, Western 587 Australia. Mineralium Deposita 46, 761-787, doi:10.1007/s00126-011-0349-4.
- 588 Douglas, G.B., Gray, D.J., Butt, C.M., 1993. Geochemistry, minerology and hydrogeochemistry of the 589 Ambassador multi-element lignite deposit, Western Australia: With additional investigations on the
- 590 characterization of organic matter. CSIRO, Australia.
- 591 Eggins, S.M., Woodhead, J.D., Kinsley, L.P.J., Mortimer, G.E., Sylvester, P., McCulloch, M.T., Hergt,
- 592 J.M., Handler, M.R., 1997. A simple method for the precise determination of \geq 40 trace elements in 593 geological samples by ICPMS using enriched isotope internal standardisation. Chemical Geology 134, 594 311-326, doi:https://doi.org/10.1016/S0009-2541(96)00100-3.
- 595
- Etschmann, B.E., Ryan, C.G., Brugger, J., Kirkham, R., Hough, R.M., Moorhead, G., Siddons, D.P., De
- 596 Geronimo, G., Kuczewski, A., Dunn, P., Paterson, D., de Jonge, M.D., Howard, D.L., Davey, P., Jensen,
- M., 2010. Reduced As components in highly oxidized environments: Evidence from full spectral 597 598 XANES imaging using the Maia massively parallel detector. American Mineralogist 95, 884-887, 599 doi:10.2138/am.2010.3469.
- 600 Fuchs, L.H., Gebert, E., 1958. X-Ray Studies of Synthetic Coffinite, Thorite and Uranothorites. 601 American Mineralogist 43, 243-248.
- 602 Greenwood, P., Brocks, J., Grice, K., Schwark, L., Jaraula, C., Dick, J., Evans, K., 2013. Organic 603 Geochemistry and Mineralogy: I. Characterisation of Organic Matter Associated with Metal Deposits. 604 Ore Geology Reviews, 1-27, doi:10.1016/j.oregeorev.2012.10.004.
- 605 Grenthe, I., Fuger, J., Konings, R.J.M., Lemire, R.J., Muller, A.B., Nguyen-Trung, C., Wanner, H., 2004. 606 Chemical Thermodynamics of Uranium. OECD- NEA, Paris, France.
- 607 Guillaumont, R., Fanghanel, T., Neck, V., Fuger, J., Palmer, D., Grenthe, I., Rand, M.H., 2003. Update 608 on the chemical thermodynamics of uranium, neptunium, plutonium, americium, and technetium. 609 OECD, Nuclear Energy Agency, Paris, France, p. 959.
- 610 Heiri, O., Lotter, A., Lemcke, G., 2001. Loss on ignition as a method for estimating organic and 611 carbonate content in sediments: reproducibility and comparability of results. Journal of 612 Paleolimnology 25, 101-110, doi:10.1023/a:1008119611481.
- 613 Idiz, E.F., Carlisle, D., Kaplan, I.R., 1986. Interaction between organic matter and trace metals in a 614 uranium rich bog, Kern County, California, U.S.A. Applied Geochemistry 1, 573-590, 615 doi:10.1016/0883-2927(86)90065-X.
- 616 Jaraula, C.M.B., Schwark, L., Moreau, X., Pickel, W., Bagas, L., Grice, K., 2015. Radiolytic alteration of
- 617 biopolymers in the Mulga Rock uranium deposit. Applied Geochemistry 52, 97-108, 618 doi:10.1016/j.apgeochem.2014.11.012.
- 619 Kamber, B.S., Greig, A., Schoenberg, R., Collerson, K.D., 2003. A refined solution to Earth's hidden 620 niobium: implications for evolution of continental crust and mode of core formation. Precambrian
- 621 Research 126, 289-308, doi:https://doi.org/10.1016/S0301-9268(03)00100-1.
- 622 Kanematsu, M., Perdrial, N., Um, W., Chorover, J., O'Day, P.A., 2014. Influence of Phosphate and 623 Silica on U(VI) Precipitation from Acidic and Neutralized Wastewaters. Environmental Science & 624 Technology 48, 6097-6106, doi:10.1021/es4056559.
- 625 Kaplan, D.I., Xu, C., Huang, S., Lin, Y., Tolić, N., Roscioli-Johnson, K.M., Santschi, P.H., Jaffé, P.R., 2016.
- 626 Unique Organic Matter and Microbial Properties in the Rhizosphere of a Wetland Soil.
- 627 Environmental Science & Technology 50, 4169-4177, doi:10.1021/acs.est.5b05165.
- 628 Landais, P., 1996. Organic geochemistry of sedimentary uranium ore deposits. Ore Geology Reviews
- 629 11, 33-51, doi:10.1016/0169-1368(95)00014-3.

- Langmuir, D., 1978. Uranium solution-mineral equilibria at low temperatures with applications to
 sedimentary ore deposits. Geochimica et Cosmochimica Acta 42, 547-569, doi:10.1016/00167037(78)90001-7.
- Latta, D.E., Mishra, B., Cook, R.E., Kemner, K.M., Boyanov, M.I., 2014. Stable U(IV) Complexes Form
- at High-Affinity Mineral Surface Sites. Environmental Science & Technology 48, 1683-1691,doi:10.1021/es4047389.
- Law, G.T.W., Geissler, A., Burke, I.T., Livens, F.R., Lloyd, J.R., McBeth, J.M., Morris, K., 2011. Uranium
- 637 Redox Cycling in Sediment and Biomineral Systems. Geomicrobiology Journal 28, 497-506, 638 doi:10.1080/01490451.2010.512033.
- Lenhart, J.J., Figueroa, L.A., Honeyman, B.D., Kaneko, D., 1997. Modeling the adsorption of U(VI)
 onto animal chitin using coupled mass transfer and surface complexation. Colloids and Surfaces A:
 Physicochemical and Engineering Aspects 120, 243-254, doi:10.1016/S0927-7757(96)03865-4.
- Li, K., Etschmann, B., Rae, N., Reith, F., Ryan, C.G., Kirkham, R., Howard, D., Rosa, D.R.N., Zammit, C.,
- Pring, A., Ngothai, Y., Hooker, A., Brugger, J., 2016. Ore Petrography Using Megapixel X-Ray Imaging:
 Rapid Insights into Element Distribution and Mobilization in Complex Pt and U-Ge-Cu Ores. Economic
 Goology 111, 487, 501, doi:10.2112/ocongoo.111.2.487
- 645 Geology 111, 487-501, doi:10.2113/econgeo.111.2.487.
- Lidman, F., Mörth, C.M., Laudon, H., 2012. Landscape control of uranium and thorium in boreal
 streams spatiotemporal variability and the role of wetlands. Biogeosciences 9, 4773-4785,
 doi:10.5194/bg-9-4773-2012.
- Luo, W., Gu, B., 2008. Dissolution and Mobilization of Uranium in a Reduced Sediment by Natural
 Humic Substances under Anaerobic Conditions. Environmental Science & Technology 43, 152-156,
 doi:10.1021/oc8013070
- 651 doi:10.1021/es8013979.
- Mehta, V.S., Maillot, F., Wang, Z., Catalano, J.G., Giammar, D.E., 2014. Effect of co-solutes on the products and solubility of uranium(VI) precipitated with phosphate. Chemical Geology 364, 66-75, doi:10.1016/j.chemgeo.2013.12.002.
- 655 Meunier, J.D., Trouiller, A., Brulhert, J., Pagel, M., 1989. Uranium and organic matter in a 656 paleodeltaic environment; the Coutras Deposit (Gironde, France). Economic Geology 84, 1541-1556.
- Mikutta, C., Langner, P., Bargar, J.R., Kretzschmar, R., 2016. Tetra- and Hexavalent Uranium Forms
 Bidentate-Mononuclear Complexes with Particulate Organic Matter in a Naturally Uranium-Enriched
- Peatland. Environmental Science & Technology 50, 10465–10475, doi:10.1021/acs.est.6b03688.
- Min, M.Z., Meng, Z.W., Sheng, G.Y., Min, Y.S., Liu, X., 2000. Organic geochemistry of paleokarsthosted uranium deposits, South China. Journal of Geochemical Exploration 68, 211-229,
 doi:10.1016/s0375-6742(99)00085-0.
- 663 Newsome, L., Morris, K., Lloyd, J.R., 2014. The biogeochemistry and bioremediation of uranium and 664 other priority radionuclides. Chemical Geology 363, 164-184, doi:10.1016/j.chemgeo.2013.10.034.
- 665 Och, L.M., Müller, B., März, C., Wichser, A., Vologina, E.G., Sturm, M., 2016. Elevated uranium 666 concentrations in Lake Baikal sediments: Burial and early diagenesis. Chemical Geology 441, 92-105,
- 667 doi:10.1016/j.chemgeo.2016.08.001.
- 668 Owen, D.E., Otton, J.K., 1995. Mountain wetlands: Efficient uranium filters potential impacts.
 669 Ecological Engineering 5, 77-93, doi:10.1016/0925-8574(95)00013-9.
- 670 Paterson, D., de Jonge, M.D., Howard, D.L., Lewis, W., McKinlay, J., Starritt, A., Kusel, M., Ryan, C.G.,
- 671 Kirkham, R., Moorhead, G., Siddons, D.P., 2011. The X-ray Fluorescence Microscopy Beamline at the
- Australian Synchrotron. In: McNulty, I., Eyberger, C., Lai, B. (Eds.). 10th International Conference on
 X-Ray Microscopy, pp. 219-222.
- 674 Ravel, B., Newville, M., 2005. ATHENA, ARTEMIS, HEPHAESTUS: data analysis for X-ray absorption
- 675 spectroscopy using IFEFFIT. Journal of Synchrotron Radiation 12, 537-541, 676 doi:10.1107/s0909049505012719.
- 677 Read, D., Bennett, D.G., Hooker, P.J., Ivanovich, M., Longworth, G., Milodowski, A.E., Noy, D.J., 1993.
- 678 The migration of uranium into peat-rich soils at Broubster, Caithness, Scotland, U.K. Journal Of
- 679 Contaminant Hydrology 13, 291-308, doi:10.1016/0169-7722(93)90067-3.

- Regenspurg, S., Margot-Roquier, C., Harfouche, M., Froidevaux, P., Steinmann, P., Junier, P., BernierLatmani, R., 2010. Speciation of naturally-accumulated uranium in an organic-rich soil of an alpine
 region (Switzerland). Geochimica et Cosmochimica Acta 74, 2082-2098,
 doi:10.1016/j.gca.2010.01.007.
- 684 Rehr, J.J., Kas, J.J., Vila, F.D., Prange, M.P., Jorissen, K., 2010. Parameter-free calculations of X-ray 685 spectra with FEFF9. Physical Chemistry Chemical Physics 12, 5503-5513, doi:10.1039/b926434e.
- 686 Renock, D., Mueller, M., Yuan, K., Ewing, R.C., Becker, U., 2013. The energetics and kinetics of uranyl
- reduction on pyrite, hematite, and magnetite surfaces: A powder microelectrode study. Geochimica
 et Cosmochimica Acta 118, 56-71, doi:10.1016/j.gca.2013.04.019.
- 689 Rowell, D.L., 1994. Soil Science Methods and Applications. Longman UK, Harlow, England.
- Ryan, C., 2000. Quantitative trace element imaging using PIXE and the nuclear microprobe.International Journal of Imaging Systems and Technology 11, 219-230.
- Ryan, C., Jamieson, D., 1993. Dynamic analysis: on-line quantitative PIXE microanalysis and its use in
 overlap-resolved elemental mapping. Nuclear Instruments and Methods in Physics Research Section
 B: Beam Interactions with Materials and Atoms 77, 203-214.
- Ryan, C.G., Siddons, D.P., Kirkham, R., Li, Z.Y., Jonge, M.D.d., Paterson, D.J., Kuczewski, A., Howard,
- 696 D.L., Dunn, P.A., Falkenberg, G., Boesenberg, U., Geronimo, G.D., Fisher, L.A., Halfpenny, A., Lintern,
- M.J., Lombi, E., Dyl, K.A., Jensen, M., Moorhead, G.F., Cleverley, J.S., Hough, R.M., Godel, B., Barnes,
- S.J., James, S.A., Spiers, K.M., Alfeld, M., Wellenreuther, G., Vukmanovic, Z., Borg, S., 2014. Maia X ray fluorescence imaging: Capturing detail in complex natural samples. Journal of Physics:
- 700 Conference Series 499, 012002.
- Shanbhag, P.M., Choppin, G.R., 1981. Binding of uranyl by humic acid. Journal of Inorganic and
 Nuclear Chemistry 43, 3369-3372, doi:10.1016/0022-1902(81)80117-0.
- Spirakis, C.S., 1996. The roles of organic matter in the formation of uranium deposits in sedimentary
 rocks. Ore Geology Reviews 11, 53-69, doi:10.1016/0169-1368(95)00015-1.
- Tinnacher, R.M., Nico, P.S., Davis, J.A., Honeyman, B.D., 2013. Effects of Fulvic Acid on Uranium(VI)
 Sorption Kinetics. Environmental Science & Technology 47, 6214-6222, doi:10.1021/es304677c.
- 707 Tokunaga, T.K., Kim, Y., Wan, J., Yang, L., 2012. Aqueous Uranium(VI) Concentrations Controlled by
- 708 Calcium Uranyl Vanadate Precipitates. Environmental Science & Technology 46, 7471-7477, 709 doi:10.1021/es300925u.
- 710 Tokunaga, T.K., Wan, J.M., Pena, J., Brodie, E.L., Firestone, M.K., Hazen, T.C., Sutton, S.R., Lanzirotti,
- A., Newville, M., 2005. Uranium reduction in sediments under diffusion-limited transport of organic
 carbon. Environmental Science & Technology 39, 7077-7083, doi:10.1021/es050221a.
- 713 Tsarev, S., Collins, R.N., Fahy, A., Waite, T.D., 2016. Reduced Uranium Phases Produced from
- Anaerobic Reaction with Nanoscale Zerovalent Iron. Environmental Science & Technology 50, 2595-
- 715 2601, doi:10.1021/acs.est.5b06160.
- 716 Vimy, 2014. Mulga Rock Uranium Project Resource Upgrade. Perth.
- 717 Vimy, 2015. Pre-Feasibility Study Reaffirms Mulga Rock Project as one of Australia's Leading
- 718 Undeveloped Uranium Projects
- 719 Vimy, 2016a. Resource Update for Mulga Rock Project.
- 720 Vimy, 2016b. Significant Resource Upgrade for Mulga Rock Project.
- 721 Vimy, 2016c. Vimy Resources Limited; Annual Report 2016. Perth, p. 98.
- 722 Warwick, P., Evans, N., Hall, A., Walker, G., Steigleder, E., 2005. Stability constants of U(VI) and U(IV)-
- humic acid complexes. Journal of Radioanalytical and Nuclear Chemistry 266, 179-190,
 doi:10.1007/s10967-005-0890-7.
- Wood, S.A., 1996. The role of humic substances in the transport and fixation of metals of economic
- 726 interest (Au, Pt, Pd, U, V). Ore Geology Reviews 11, 1-31, doi:10.1016/0169-1368(95)00013-5.
- 727 Wülser, P.-A., Brugger, J., Foden, J., Pfeifer, H.-R., 2011. The Sandstone-Hosted Beverley Uranium
- 728 Deposit, Lake Frome Basin, South Australia: Mineralogy, Geochemistry, and a Time-Constrained
- 729 Model for Its Genesis. Economic Geology 106, 835-867, doi:10.2113/econgeo.106.5.835.

Zammit, C.M., Brugger, J., Southam, G., Reith, F., 2014. In situ recovery of uranium — the microbial
 influence. Hydrometallurgy 150, 236-244, doi:10.1016/j.hydromet.2014.06.003.

Zhao, D., Wang, X., Yang, S., Guo, Z., Sheng, G., 2012. Impact of water quality parameters on the
sorption of U(VI) onto hematite. Journal of Environmental Radioactivity 103, 20-29,
doi:10.1016/j.jenvrad.2011.08.010.

- 735 Zhou, P., Gu, B.H., 2005. Extraction of oxidized and reduced forms of uranium from contaminated
- 736 soils: Effects of carbonate concentration and pH. Environmental Science & Technology 39, 4435-
- 737 4440, doi:10.1021/es0483443.
- 738
- 739

1 Characterization of uranium redox state in organic-rich

2 Eocene sediments

- 3 Susan A Cumberland^{1,2,3}, Barbara Etschmann², Joël Brugger², Grant Douglas⁴, Katy Evans⁵,
- 4 Louise Fisher⁶, Peter Kappen³, John W. Moreau¹
- 5 1 School of Earth Sciences, University of Melbourne, Parkville, Victoria 3100, Australia
- 6 2 School of Earth, Atmosphere and Environment, Monash University, Clayton 3800, Victoria,7 Australia
- 8 3 ANSTO Australian Synchrotron, 800 Blackburn Road, Clayton 3168, Victoria, Australia
- 9 4 CSIRO Land and Water, Floreat, Western Australia, Australia
- 10 5 Western Australian School of Mines, Curtin University, Bentley, Western Australia, Australia
- 11 6 CSIRO Mineral Resources, Bentley, Western Australia, Australia

12

SUPPLEMENTARY INFORMATION 14

15 SI Table 1 Location and drill depth of cores analysed

						Core Depth	
Hole Id	Deposit	Samples from core	Northing	Easting	RL	(m)	Туре
						()	
NNA 5612	Princess	MR5612a,b	6684044.69	578834.44	342.8	54	AC
NNA 5613	Princess	MR5613a,b	6684001.28	578902.04	341.48	54	AC
NNA 5636	Princess	MR5636	6684253.49	579486.8	345.52	72	AC
NNA 5706	Emperor	MR5706a,b	6690894.00	552782.00	334.5	62	AC
NNA 5728	Shogun	MR5728a,b	6686912.00	563587.00	318.4	51.0	AC
NNA 5766	Ambassador	MR5766a,b,c,d	6682660.00	576969.00	342.1	64	AC
NND 5076	Emperor	MR5076a,b	6682834.90	579131.20	339.5	48.5	DDH
NND 5077	Ambassador	MR5077	6682189.32	576147.80	333	54.3	DDH
NND 5078	Ambassador	MR5078	6682605.99	577075.96	342.4	57.0	DDH
CD1577	Ambassador		6682708.94	579942.81	331.4	43.5	DDH

16 AC = air core; DDH = diamond drill hole, RL = Reduced Level – elevation above sea level in metres of

17 the drill hole collar (Vimy, 2014, 2016b)

Supplementary Information

20 SI Table 2 Uranium mineral standards and location where known

Number	Name	Formula	Location
M 33701	andersonite	Na ₂ CaUO ₂ (CO ₃) ₃ · 6H ₂ O	
M 21676	autunite	Ca(UO ₂) ₂ (PO ₄) ₂	
M 30851	boltwoodite	(Na, K)(UO ₂) ₂ (HSiO ₄) · H ₂ O	
M 33175	brannerite	U TiO ₄	
M 32495	coffinite	USiO ₄	
M 28752	coffinite	USiO ₄	USA, New Mexico Gransted, McKinley Country, Paddy Murphy Mines (35 8 50N, 107 51 3 W)
M 33214	curienite	Pb ₂ (UO ₂) ₂ (V ₂ O ₈) ₂ · 5H ₂ O	
M 22869	davidite	(La, Ce)(Y, U, Fe)(Ti,Fe3+) ₂₀ (O,HO) ₃₈	
M 45514	saleeite	$Mg(UO_2)_2(PO_4)_2 \cdot 10H_2O$	
M 44367	schoepite	(UO ₂)O ₂ (OH) ₁₂ · 12H ₂ O	Democratic Republic of Congo (Zaire), Shaba (Katanga) Province, Kolwezi Munsonoi Mine (10 45 S 25 25 E)
M 17631	torbernite	Cu(UO ₂) ₂ (PO ₄) ₂ · 8 - 12 H ₂ O	
M 27299	soddyite	(UO ₂)SiO ₄ (H ₂ O) ₂	Democratic Republic of Congo (Zaire), Shaba (Katanga) Province, Swamba
M 26415	uraninite	UO ₂	
M 19266	uraninite	UO ₂	Australia, Northern Territory, South Alligator River
M 34470	uranophane	$Ca(UO_2)_2(SiO_3)(OH)_2 \cdot 5H_2O$	
M 40774	uranopilite	(UO ₂) ₆ (SO ₄)O ₂ (OH) ₆ (H ₂ O) ₆	
M 34904	uranopilite	(UO ₂) ₆ (SO ₄)O ₂ (OH) ₆ (H ₂ O) ₆	El Sherana mine, South Alligator River, NT, AU (13 31 S 132 31 E)
M 25802	zippeite	Mg $(UO_2)_6 (SO_4)_3 (OH)_{10} (H_2 O)_n$	

21 Source: Museum Victoria, formulas are generic

22

Supplementary Information

24 SI Table 3 EXAFS fitting and paths for the uranyl nitrate standard

Ligand	N	R (Å)	ss (Ų)	ΔE ₀	Red X ²	k-range	R-range	k-weighting	SO ²
01	2 (fix)	1.787(7)	0.002(1)	10(1)	978	1-4.5 Å ⁻¹	2-12.5 Å	1,2,3	0.8 5
02	2 (fix)	2.34(3)	0.008(7)						
N1	2 (fix)	2.47(4)	0.003 (fix)						
03	2 (fix)	2.53(3)	0.003 (fix)						
O4_2(path 37)	4 (fix)	4.30(3)	0.003 (fix)						
MS paths	21: 01-01	3.57(1)	0.002(1)						
	22: 01-01	3.57(1)	0.002(1)						
	24: N1-O1	3.65(4)	0.0025						
	27: 03-01	3.71(3)	0.0025						
error (quadrature	e) for MS paths								
0.009899495									
0.040607881									
0.030805844									

27 SI Table 4 Summary of physicochemical data in MR cores n = 18.

Sample ID	Core	Depth (m)	Deposit	U (mg kg ⁻ 1)	рН	тос %	LOI ₅₅₀ %	LOI ₉₀₀ %	Carbonate % (**)	Total N %
CD1577	1577	43.5	Amb	nd	nd	nd	nd	nd	nd	nd
MR5076a	5076	44.3-44.4	Amb	1877	5.24	39	NA	NA	NA	0.46
MR5076b	5076	44.7-44.8	Amb	14930	3.09	30	NA	NA	NA	0.42
MR5077	5077	42.2-42.5	Amb	5760	4.12	44	NA	NA	NA	0.37
MR5078	5078	51.5-51.5	Amb	3720	3.02	34	NA	NA	NA	0.27
MR5612a	5612	41-42	Pri	2	4.7	60	57.7	78.3	17.3	0.35
MR5612b	5612	43-44	Pri	3.2	4.34	3.8	NA	NA	NA	0.04
MR5612c	5612	45-46	Pri	65	3.22	3.3	NA	NA	NA	0.05
MR5613a*	5613	40-41	Pri	5816	5.57	33.5	43.4	55.9	17.1	0.23
MR5613b*	5613	41-43	Pri	4588	5.61	39.5	NA	NA	NA	0.23
MR5636	5636	55-56	Pri	278	4.96	3.2	NA	NA	NA	0.05
MR5706a	5706	42-43	Emp	115	3.15	22	19.1	30.6	15.7	0.33
MR5706b	5706	55-56	Emp	40	7.06	0.72	1	1	0.1	0.02
MR5728a	5728	28.5-29	Sho	5590	3.33	26	37.9	41.9	5.1	0.29
MR5728b	5728	34-34.5	Sho	113	3.64	3.6	3.3	3.5	0.2	0.03
MR5766a	5766	53-53.5	Amb	5012	4.35	15.1	19.4	28.9	13	0.18
MR5766b*	5766	55.5-56	Amb	623	5.43	16	21.5	24.6	4.2	0.1
MR5766c*	5766	56.5-57	Amb	1206	4.60	21	22.5	33.2	14.6	0.14
MR5766d	5766	58-58.5	Amb	621	4.72	13	10.5	12.6	2.9	0.09
Mean				2798	4.5	22.5	23.6	31.1	9	0.2
Min				1.9	3	0.7	1	1	0.1	0
Max				14932	7.1	58.8	57.7	78.3	17.3	0.5
SD				3805	1.1	16.8	17.9	23.6	7.1	0.1
Median				914.2	4.5	21.2	20.4	29.8	9.1	0.2

28 *EXAFS Data available

^{**}NB estimated CO₃²⁻was calculated by LOI₉₀₀ – LOI₅₅₀ x MW C/ 100 (Dean, 1974; Heiri et al., 2001). Sample ID is the lab number

30 given by Melbourne University. NA = not analysed

32 SI Table 5 Elemental data for cores (1/2)

Sample number	Core	Depth	Deposit	U (XRF)	U ashed	U	As	Ве	Ca	Со	Cr	Cu
		m		mg kg⁻¹	mg kg⁻¹	mg kg-1	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg-1	mg kg⁻¹	mg kg⁻¹
MR5076a	5076	44.3-44.4	Amb	NA	ND	1880	17	16	12900	282	261	624
MR5076b	5076	44.7-44.8	Amb	NA	ND	14900	296	128	5290	4300	163	7290
MR5077	5077	42.2-42.5	Amb	NA	ND	5760	49	3	2950	142	880	458
MR5078	5078	51.5-51.5	Amb	NA	ND	3720	157	19	34	541	337	1500
MR5766a	5766	53-53.5	Amb	5280	5198.0	5010	50	20	2450	2960	535	696
MR5766b	5766	55.5-56	Amb	642	586.4	623	4	3	912	70	886	<lod< td=""></lod<>
MR5766c	5766	56.5-57	Amb	1160	1223.0	1210	4	6	1240	359	1080	<lod< td=""></lod<>
MR5766d	5766	58-58.5	Amb	686	641.2	621	2	2	266	640	1200	3390
MR5706a	5706	42-43	Emp	132	129.2	115	6	2	884	149	191	53
MR5706b	5706	55-56	Emp	28.5	21.7	40	0	0	50	7	520	8
MR5612a	5612	41-42	Pri	<lod< td=""><td>5.8</td><td>2</td><td>6</td><td>3</td><td>4250</td><td>81</td><td>21</td><td><lod< td=""></lod<></td></lod<>	5.8	2	6	3	4250	81	21	<lod< td=""></lod<>
MR5612b	5612	43-44	Pri	<lod< td=""><td>ND</td><td>3</td><td>3</td><td>1</td><td>744</td><td>223</td><td>19</td><td>1</td></lod<>	ND	3	3	1	744	223	19	1
MR5612c	5612	45-46	Pri	80.5	<lod< td=""><td>65</td><td>6</td><td>1</td><td>557</td><td>653</td><td>22</td><td>2</td></lod<>	65	6	1	557	653	22	2
MR5613a	5613	40-41	Pri	5385.0	5614.0	5820	65	7	7000	83	2550	5120
MR5613b	5613	41-43	Pri	6296.0	<lod< td=""><td>4590</td><td>128</td><td>4</td><td>4650</td><td>289</td><td>1300</td><td>5780</td></lod<>	4590	128	4	4650	289	1300	5780
MR5636	5636	55-56	Pri	<lod< td=""><td>ND</td><td>278</td><td>13</td><td>3</td><td>603</td><td>291</td><td>67</td><td>300</td></lod<>	ND	278	13	3	603	291	67	300
MR5728a	5728	28.5-29	Sho	903.0	5560	5590	6	1	974	0	1040	115
MR5728b	5728	34-34.5	Sho	<lod< td=""><td>125.1</td><td>113</td><td>3</td><td>0</td><td>166</td><td>1</td><td>99</td><td>109</td></lod<>	125.1	113	3	0	166	1	99	109

Supplementary Information

35 SI Table 5 Elemental data for cores (2/2)

36

sample number	core	depth	deposit	Ga	Li	Ni	Pb	Rb	Sc	Ti	Th	V	W	Zn	total REE+Y
		m		mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹	mg kg⁻¹
MR5076	5076	44.3-44.4	Amb	3	81	1250	158	1.9	192	5270	44	145	1.1	1750	6270
MR5076	5076	44.7-44.8	Amb	<lod< td=""><td>13</td><td>8310</td><td>5290</td><td>2.4</td><td>1820</td><td>1850</td><td>10</td><td>312</td><td>0.6</td><td>3850</td><td>15400</td></lod<>	13	8310	5290	2.4	1820	1850	10	312	0.6	3850	15400
MR5077	5077	42.2-42.5	Amb	5	1	172	3650	0.8	427	10300	12	144	2.2	14	642
MR5078	5078	51.5-51.5	Amb	24	17	2050	2750	8.5	459	45500	49	493	9.2	770	4910
MR5766a	5766	53-53.5	Amb	26	28	10800	379	2.4	915	24900	33	409	6.9	12800	5530
MR5766b	5766	55.5-56	Amb	4	4	241	236	0.9	46	7460	13	48	3.1	1810	525
MR5766c	5766	56.5-57	Amb	4	3	940	600	0.8	88	6080	12	66	3.1	2560	698
MR5766d	5766	58-58.5	Amb	3	6	1250	458	1.6	51	5010	20	30	2.8	2280	234
MR5706a	5706	42-43	Emp	19	22	85	63	11.8	24	4970	24	98	2.6	10	291
MR5706b	5706	55-56	Emp	1	4	24	8	0.4	8	474	1	4	2.2	29	22
MR5612a	5612	41-42	Pri	2	9	232	7	1.2	29	6620	5	195	1.6	3810	601
MR5612b	5612	43-44	Pri	2	4	365	4	0.3	15	6940	10	39	4.0	2050	191
MR5612c	5612	45-46	Pri	18	16	1370	11	4.5	12	3100	7	35	5.0	7840	147
MR5613a	5613	40-41	Pri	5	3	362	2290	1.3	825	11100	201	148	3.9	1620	1710
MR5613b	5613	41-43	Pri	4	5	614	2090	1.1	330	7820	33	143	2.0	<lod< td=""><td>948</td></lod<>	948
MR5636	5636	55-56	Pri	4	15	605	135	3.7	34	3550	11	30	1.5	1590	400
MR5728a	5728	28.5-29	Sho	49	6	4	465	4.0	377	8800	244	153	4.0	13	514
MR5728b	5728	34-34.5	Sho	1	2	1	6	0.4	7	1190	39	12	2.4	<lod< td=""><td>15</td></lod<>	15

38 SI Table 6 Summary of metals within deposits shown as average data in parts per million (mg kg⁻¹)

- ·	Mean	Min	Max	Ambassador	Emperor	Princess	Shogun
Element				n=8	n=2	n=6	n=2
As	45.2	0.3	295.7	72.3	3	36.7	4.6
Ве	12.2	0.1	128.3	24.6	0.8	3.3	0.3
Са	2553	33.6	12950	3260	466.9	2966	569.9
Со	614.8	0.5	4300	1161	78.2	270.1	0.6
Cr	621	19.4	2552	668.8	355.7	662.7	570.1
Cu	1696	0.9	7286	2325	30.1	2241	111.8
Ga	9.6	0	48.8	8.3	10.1	6	24.8
Li	13.4	0.9	80.5	19.2	13.1	8.8	4.1
Ni	1594	0.7	10800	3128	54.5	591.1	2.5
Pb	1033	3.7	5292	1690	35.5	754.7	235.9
Rb	2.7	0.3	11.8	2.4	6.1	2	2.2
Sc	314.4	6.8	1822	500	15.7	207.3	191.9
Th	42.6	1.1	244.1	24	12.4	44.4	141.6
Ті	8934	474.3	45450	13290	2720	6516	4997
U	2798	1.9	14930	4219	77.9	1792	2853
V	139	4.1	492.7	205.7	51.1	98.1	82.7
W	3.2	0.6	9.2	3.6	2.4	3	3.2
Zn	2673	10.1	12770	3225	19.6	3382	12.8
REE+Y	2168	14.6	15390	4274	156.8	665.3	264.5

40

Supplementary Information

41 SI Table 7 Correlations between U and metals; TOC and metals

Correlation Test	R	R ²	p-value	significance
U vs TOC	0.377	0.142	0.1225	*
U vs As	0.861	0.741	0.0000	***
U vs REE+Y	0.805	0.649	0.0001	***
U vs Ba	0.14	0.02	0.5786	ns
U vs Ca	0.345	0.119	0.1606	*
U vs Cd	0.816	0.666	0.0000	***
U vs Mo	0.195	0.038	0.4377	*
U vs Ni	0.602	0.363	0.0082	**
U vs Pb	0.875	0.765	0.0000	***
U vs Sn	0.234	0.055	0.3491	*
U vs Sr	0.278	0.077	0.2641	*
U vs Th	0.282	0.08	0.2567	*
U vs V	0.577	0.333	0.0121	**
U vs Zr	0.113	0.013	0.6548	ns
TOC vs As	0.344	0.118	0.1626	*
TOC vs REE+Y	0.258	0.067	0.3006	*
TOC vs Ba	0.302	0.091	0.2233	*
TOC vs Ca	0.591	0.35	0.0097	**
TOC vs Cd	0.145	0.021	0.5663	ns
TOC vs Mo	-0.069	0.005	0.7841	ns
TOC vs Ni	-0.01	0	0.9695	ns
TOC vs Pb	0.457	0.208	0.0568	*
TOC vs Sn	0.349	0.122	0.1552	*
TOC vs Sr	0.54	0.291	0.0208	**
TOC vs Th	0.184	0.034	0.4643	*
TOC vs V	0.481	0.231	0.0433	**
TOC vs Zr	0.021	0	0.9349	ns

42 *** p<0.005, ** p<0.05, * p <0.5, ns p >0.51

45	SI Table 8 Results from the linear combination fitting.	Two end-members were represented by synthetic uraninite (U(IV)) and mineral uranopilite (U(VI))
----	---	---

							uraninite		Uranopilite	
Data	Core	Rfactor	Chinu	chisqr	nvarys	scaleby	weight	error	weight	error
MR5766c	NNA 5766	0.0004	9.89E-05	0.017	1	1	0.14	0.0043	0.86	0.0043
MR5766d	NNA 5766	0.011	0.0027	0.452	1	1	0.35	0.0226	0.65	0.0226
MR5713b	NNA 5613	0.008	0.0019	0.316	1	1	0.36	0.0188	0.64	0.0188
MR5713a	NNA 5613	0.0005	0.00013	0.022	1	1	0.12	0.0049	0.88	0.0049
MR5076a	NND 5076	0.006	0.0012	0.204	1	1	0.33	0.0151	0.67	0.0151
MR5077	NND 5077	0.007	0.0016	0.266	1	1	0.24	0.0173	0.76	0.0173
MR5076b	NND 5076	0.002	0.0004	0.072	1	1	0.19	0.0090	0.81	0.0090
MR5766b	NNA 5766	0.003	0.0008	0.129	1	1	0.11	0.0120	0.89	0.0120
reduced_CD1577	CD 1577	0.015	0.0025	0.419	1	1	0.65	0.0217	0.35	0.0217
oxidised_CD1577	CD 1577	0.003	0.0005	0.091	1	1	0.16	0.0101	0.84	0.0101
U(IV) Synthetic uraninite	(Tsarev et al., 2016)						1			
U(VI) Natural uranopilite	M 40774								1	

Supplementary Information

48 SI Table 9 Alternative EXAFS model using more MS paths

								k-range			
Sample	Ligand	N	R (Å)	ss (Ų)	ΔE ₀	Red X ²	r-factor	(Å)	R-range (Å)	k-weighting	S0 ²
MR5766c	Oax2	1	1.75(4)	0.003 (fix)	7(1)	111	0.025	2 - 12	1.3 - 4	1,2,3	0.85
	Oax1	1	1.77(3)	0.003 (fix)							
	015	2	2.20(1)	0.003 (fix)							
	014	2	2.34(2)	0.003 (fix)							
	OH20	2	2.39(2)	0.003 (fix)							
	С	2	2.92(3)	0.003 (fix)							
	U-Oax2-U-Oax2 (MS)		3.51(6)	0.006 (fix)							
	U-Oax2-Oax1 (MS)		3.58(5)	0.006 (fix)							
	U-Oax2-U-Oax1 (MS)		3.58(5)	0.006 (fix)							
MR5713a	Oax2	1	1.77(10)	0.003 (fix)	9(3)	111	0.025	3 - 11.5	1.3 - 4	1,2,3	0.85
	Oax1	1	1.77(8)	0.003 (fix)							
	015	2	2.23(5)	0.003 (fix)							
	014	2	2.38(4)	0.003 (fix)							
	OH20	2	2.43(6)	0.003 (fix)							
	С	2	2.89(7)	0.003 (fix)							
	U-Oax2-U-Oax2 (MS)		3.54(11)	0.006 (fix)							
	U-Oax2-Oax1 (MS)		3.58(13)	0.006 (fix)							
	U-Oax2-U-Oax1 (MS)		3.58(13)	0.006 (fix)							
oxidised_CD1577*	Oax2	1	1.75(4)	0.003 (fix)	7(1)	111	0.025	2.5 - 12	1 - 4	1,2,3	0.85
	Oax1	1	1.77(3)	0.003 (fix)							
	015	2.8(4)	2.20(1)	0.003 (fix)							
	014	2	2.34(2)	0.003 (fix)							
	OH20	2	2.39(2)	0.003 (fix)							
	с	2	2.88(4)	0.003 (fix)							
	U-Oax2-U-Oax2 (MS)		3.50(6)	0.006 (fix)							
	U-Oax2-Oax1 (MS)		3.58(5)	0.006 (fix)							
	U-Oax2-U-Oax1 (MS)		3.58(5)	0.006 (fix)							

49

50 *these distances were constrained to be the same as those of MR5766c these 3 data sets were fitted together.

51 SI Figure 1 XRD spectra

52	A)	Uranyl nitrate Patterns were collected with a Bruker D8 Advance Eco diffractometer equipped with a
53		LYNXEYE XE linear position sensitive detector and using a Cu X-ray tube operated at 40 kV and 25 mA.
54		Data were collected from 3–80° 20 with a step size of 0.02°20 and a counting rate of 1 s/step. Minerals
55		were identified with reference to standard patterns from the ICDD Powder Diffraction File 2 (PDF-2)
56		database using the DIFFRAC.EVA v.4 software package (available from Bruker AXS).





- -

B) Background of CD1577. Data were measured using a Bruker GADDS microdiffractometer, using Cu Kα
 radiation from an X-ray tube operating at 40 mA, 40 kV. The incident beam was passed through crossed
 reflecting (Gobel) mirrors, resulting in a monochromatic, parallel, high-brilliance source which was then
 collimated through a 200 µm pinhole system. The spectrum was consistent with quartz alpha mixed with
 an unidentified minor phase.







Supplementary Information

Coffinite- Micro-XRD pattern of mineralised area of uranium, spot analysis was taken at the edge of a pyrite grain. Red lines show markers corresponding to U(IV) coffinite pattern (USiO₄) (RRUFF). Beam = 300 μm Cu Kα collimator.



SI Figure 2 U L3 edge XAS x-ray absorption of the near edge structure for spectra of bulk cores and of some
 mineral standards



SI Figure 3 U L3 edge XAS - Extended x-ray absorbance of the fine structure shown in K² space. Top four
spectra are fitted EXAFS of the bulk cores where solid line is the sample. The U(IV) and U(VI) lignite samples
were from a thin section of MR lignite (CD 1577). The bottom seven spectra are mineral standards donated by
Museum Victoria.



92 SI Figure 4 EDS-SEM map and spectra from CD1577

