

NOTICE: this is the author's version of a work that was accepted for publication in Organic Geochemistry. Changes resulting from the publishing process, such as peer review, editing, corrections, structural formatting, and other quality control mechanisms may not be reflected in this document. Changes may have been made to this work since it was submitted for publication. A definitive version was subsequently published in Organic Geochemistry, Vol. 75 (2014).
DOI: 10.1016/j.orggeochem.2014.04.013

1 **Salinity variations in the Northern Coorong Lagoon, South Australia:**
2 **Significant changes in the Ecosystem following human alteration to the**
3 **natural water regime**

4 *Svenja Tulipani^{a*}, Kliti Grice^{a*}, Evelyn Krull^b, Paul Greenwood^{a,c}, Andrew. T. Revill^d*

5 *^a WA Organic and Isotope Geochemistry Centre, Department of Chemistry, Curtin University, GPO Box U1987,*
6 *Perth, WA 6845, Australia*

7 *^b CSIRO Land and Water, Glen Osmond, SA 5064, Australia*

8 *^c Centre for Exploration Targeting and West Australian Biogeochemistry Centre, The University of Western*
9 *Australia 35 Stirling Highway, Crawley WA 6009, Australia*

10 *^d CSIRO Marine and Atmospheric Research, GPO Box 1538, Hobart 7001, Tasmania*

11 **Corresponding authors: s.tulipani@curtin.edu.au, +61 (0)8 9266 7628; k.grice@curtin.edu.au, +61 (0)8 9266*
12 *2474*

13 **Keywords:** Coorong Lagoon, δD , salinity, organic matter, perylene, 1-chloro-*n*-alkanes

14 **Abstract**

15 European settlement and drought have significantly impacted the hydrology of the Coorong a shallow
16 coastal lagoon complex in South Australia, which is part of a terminal wetland at the Mouth of the River
17 Murray. An increased salinity associated with lower water levels and progressive isolation from ocean flushes
18 contributed to a severe decline in ecological diversity over the past decades. Here we have conducted a
19 molecular and stable isotopic study of a sedimentary core from the northern Coorong Lagoon spanning more
20 than 5000 years to investigate the recent palaeoenvironmental history of the ecosystem. Major alterations were
21 evident in many biogeochemical parameters in sediments deposited after the 1950s coinciding with the
22 beginning of intensified water regulations. The most prominent shift occurred in $\delta^{13}C$ profiles of C_{21} - C_{33} *n*-
23 alkanes from average values of -23.5 ‰ to an average of -28.2 ‰. Further changes included decreases in carbon
24 preference index (CPI) and average chain length (ACL) of the *n*-alkane series as well as significant increases in
25 algal (e.g. C_{20} HBI, long-chain alkenes and C_{29} -alkadiene) and bacterial (e.g. ^{13}C depleted short-chain *n*-alkanes
26 and hopanoids, $\delta^{13}C$: -35.9 to -30.1 ‰) derived hydrocarbons. Long-chain *n*-alkanes with a strong odd-over-
27 even predominance as observed here are typically attributed to terrestrial plants. In the Coorong however,
28 terrestrial input to sedimentary OM is only minor. Therefore changes in the before mentioned parameters were

29 attributed to a source transition from a major contribution of macrophytes towards predominantly microalgae
30 and bacteria.

31 δD values of C_{21} - C_{33} *n*-alkanes showed a general trend towards more enriched values in younger sediments,
32 indicating an overall rising salinity. However, the most pronounced positive shift in these profiles again
33 occurred after the 1950s. Altogether this study demonstrates that the recent human induced changes of the
34 Coorong hydrology, compounded by a severe drought led to an increase in salinity and alterations of primary
35 production which have been much more significant than natural variations occurring throughout the Holocene
36 over several thousands of years.

37 **1. Introduction**

38 The Coorong is part of a terminal lake system in South Australia at the mouth of the River Murray. It
39 consists of two narrow coastal lagoons (North and South Lagoon) which are connected through a small channel
40 and together extend for more than 130 km in length (Fig. 1). The wetland is of great ecological significance, in
41 particular for water birds (migratory and native; Paton et al., 2009), and is therefore listed under the Ramsar
42 Convention on Wetlands and is protected as a national park. However, substantial water regulations imposed on
43 the River Murray and the lake system in the second half of the 20th century, compounded by a severe drought
44 from 2000-2010, have significantly altered the complex and dynamic hydrology of the ecosystem leading to a
45 marked increase of salinity in the Coorong and a severe decline of ecological diversity over the past several
46 decades (e.g. DEH, 2000; Shuttleworth et al., 2005; Paton et al., 2009; Dick et al., 2011). Although recent floods
47 in 2010 temporarily improved the ecological state of the Lower Lakes and led to a short term recovery of the
48 freshwater levels in the Coorong, the long term issue of providing continuous riverine input (especially under
49 future drought conditions) has not been resolved and an extensive water management plan is required to further
50 restore and maintain the diverse ecosystem of the Coorong (Kingsford et al., 2011).

51 Estuarine systems like the Coorong have multiple organic matter (OM) sources including *in situ* primary
52 production by organisms such as phytoplankton, microbial mats, macroalgae or aquatic and benthic plants as
53 well as allochthonous input from terrestrial plant debris and soil OM which enter the system mainly through
54 riverine inflow (e.g. Shuttleworth et al., 2005; Volkman et al., 2008). Changes in the type of OM – for example
55 caused by different flow regimes – can have a profound impact on the ecosystem since the base of the food-web
56 is affected (Krull et al., 2009). Efficient remediation measures for the Coorong require a profound understanding
57 of the effects of alterations in parameters such as water-level, salinity or quantity of riverine input (all of which
58 have occurred as an immediate result of human water management) on the ecosystem. This also involves

59 gaining detailed knowledge about ecology and palaeoenvironmental conditions prior to human influence, and
60 tracking changes in OM sources over time to establish their relationship with the key environmental parameters
61 previously mentioned.

62 The analysis of sedimentary biomarkers in combination with compound specific stable isotope analysis
63 (CSIA) is a powerful analytical tool for this purpose and has been frequently used to elucidate OM sources
64 and/or environmental conditions at the time of deposition (e.g. Freeman et al., 1994; Grice et al. 1996, 2001,
65 2005; Volkman et al., 2008; McKirdy et al., 2010). Compound specific δD analysis has a great potential for
66 reconstructions of climate and hydrological conditions in palaeoenvironments since D/H-ratios of biolipids in
67 autotrophic organisms are strongly influenced by the composition of the utilized source water (Sessions et al.,
68 1999; Andersen et al., 2001; Sauer et al., 2001; Dawson et al., 2004; Nabbefeld et al., 2010). δD signatures of
69 biomarkers derived from aquatic or benthic organisms thus reflect D/H ratios in the lake or seawater (Sachse et
70 al., 2006; Mügler et al., 2010) and therefore often strongly correlate with palaeosalinities (Andersen et al., 2001;
71 McKirdy et al., 2010; Mügler et al., 2010). Salinity in estuarine systems such as the Coorong co-varies with δD
72 signatures of lake water since it is largely determined by the balance of evaporation and seawater inflow (both
73 leading to a D-enrichment in the reservoir) and the input of D-depleted meteoric waters from precipitation or
74 riverine influx (Gat, 1996; Ingram et al., 1996).

75 A detailed investigation of OM sources in the recent Coorong was performed by Reville et al. (2009) who
76 analysed abundances and stable isotopic compositions of pigment and lipid biomarkers in the water-column and
77 surface sediments throughout both lagoons. McKirdy et al. (2010) carried out a comprehensive study which also
78 included bulk and compound specific isotope analysis to reconstruct variations in environmental conditions and
79 OM-sources in the Coorong from the mid to late Holocene. However, their main focus was a core from the
80 South Lagoon and only some elemental and bulk isotope data from the North Lagoon was included.
81 Nevertheless their data revealed significant differences between the ecosystems of the North and South Lagoon.
82 Krull et al. (2009) combined elemental, bulk $\delta^{13}C$ and ^{13}C -NMR analyses with radionuclide dating in several
83 cores from both lagoons spanning the past 100 years and found evidence for significant changes in sedimentary
84 OM, which coincided with the beginning of substantial human impact on the water regime.

85 However, so far there have been no biomarker studies in sediments from the North Lagoon spanning the
86 time periods before and after European settlement. Here we analysed bulk $\delta^{13}C_{OM}$ as well as abundances and
87 stable isotopic compositions ($\delta^{13}C$ and δD) of lipid biomarkers in a core from the North Lagoon dating back
88 more than 5000 years. The aim was to gain insight into the ecosystem of the North Lagoon in pre-European

89 times, reconstruct salinity variations over an extended timeframe and more robustly scrutinise the recent
90 changes in OM sources observed by Krull et al. (2009). Pollen analysis and different dating techniques
91 performed previously by Krull et al. (2009) and Fluin et al. (2007) in a core nearby allowed for a correlation of
92 our data with recent events in the Coorong.

93 **2. Experimental**

94 **2.1. Sampling location and core description**

95 *2.1.1. Environmental settings*

96 Together with Lakes Alexandrina and Albert the Coorong forms a terminal wetland at the mouth of the
97 River Murray (Fig. 1). The region experiences a Mediterranean climate, with comparatively cool and moist
98 winters and long dry summers. The elongated Coorong covers an area between 150-240 km² depending on
99 seasonal changes and features an average depth around 1 m with maximum depths of 4 m (Boon, 2000). In its
100 natural state the North Lagoon used to be connected to Lake Alexandrina which receives freshwater input from
101 the River Murray. The hydrology of the Coorong was determined by seawater exchange through the Murray
102 Mouth and freshwater input from Lake Alexandrina as well as groundwater and surface runoff *via* Salt Creek at
103 the southern end of the lagoons (Boon, 2000; Webster, 2005). Seasonal and tidal changes created various
104 dynamic habitats in and around the lagoons, supporting a diverse range of bird, fish and plant species (DEH,
105 2000; Boon, 2000).

106 Human water management with an impact on the Coorong commenced approximately 100 years ago in
107 the form of regulations and diversions from the River Murray for agricultural use and ship navigation
108 (Maheshwari et al., 1995). In 1940 barrages were installed between the Coorong and Lake Alexandrina which
109 largely prevented water exchange and reduced riverine input. However, there was still regular freshwater flow
110 into the North Lagoon until the 1950s as the barrages were regularly opened to prevent flooding of the
111 surrounding areas (Krull et al., 2009). After a major flood event in 1956 the inflow from the River Murray was
112 reduced significantly due to increased abstraction for agricultural and domestic use. A severe drought between
113 2000 and 2010 has led to further decline in freshwater input (Webster, 2005, 2010; Kingsford et al., 2011). The
114 reduced inflows contributed to siltation of the Murray Mouth limiting the seawater exchange with a temporary
115 closure occurring in 1981. During the recent drought in the beginning of this century an open connection has
116 only been maintained by regular dredging (Webster, 2005; Kingsford et al., 2011).

117 Depending on the amount of freshwater inflow through the barrages the salinities in the northern part of
118 the lagoon close to the Murray Mouth can vary from values approaching zero to seawater levels. A salinity

119 gradient shows an increase with distance from the Murray Mouth and a constant hypersalinity prevails in the
120 South Lagoon during periods of low freshwater inflow such as the drought from 2000 to 2010 (e.g. Webster,
121 2005, 2010). The salinity increase is usually considered the main cause for the extinction of key species such as
122 the macrophytes *Ruppia megacarpa* and other significant alterations to the ecosystem (Nicol, 2005; Dick et al.,
123 2011). Increased freshwater inflows after 2010 have caused a slight and temporary improvement of the
124 ecological state in the Coorong (Kingsford et al., 2011) but this is not addressed here since the presently
125 analysed samples were collected in 2005.

126 **2.1.2. Sample collection**

127 The core analysed in this study was part of a suite of samples collected throughout the Coorong in 2005
128 by University of Adelaide researchers for a project commissioned by the Department of Water, Land and
129 Biodiversity Conservation (DWL&BC) for the Upper South East (USE) Program. The sampling site was located
130 in the middle of the North Lagoon slightly south of the Murray Mouth (Fig. 1) and is equivalent to core “C4” in
131 Krull et al. (2009). Sample collection and preparation of the core are described elsewhere (e.g. Krull et al.,
132 2009). For this study the core was sampled between the depths of 0 and 115 cm. 1 and 5 cm intervals were
133 homogenized for bulk analysis ($\delta^{13}\text{C}$ and TOC) and biomarker analysis, respectively.

134 **2.1.3. Chronology**

135 In the present study we used chronological data obtained by Krull et al. (2009) and Fluin et al. (2007)
136 from a core referred to as “C3” (located in close vicinity to C4) to correlate observed alterations in biomarker
137 parameters and their stable isotopic compositions with significant events in the Coorong. Krull et al. (2009) used
138 a combination of *Pinus* pollen analysis (a species introduced by Europeans) and radionuclide-dating (^{137}Cs and
139 Pu isotopes) in various cores collected throughout the Coorong to establish chronologies over the past 50-60
140 years. Sediment ages in more basal parts of some of these cores, obtained from ^{14}C -accelerator mass
141 spectrometry (AMS), have been published by Fluin et al. (2007).

142 The first appearance of *Pinus* pollen in sediments from southeastern Australia is a common indicator for the
143 evolution of European settlement and is often used in combination with other dating techniques (Tibby, 2003).
144 In the Coorong its first occurrence roughly correlates with the 1950s (Krull et al., 2009). In core C3 and core C4
145 the first *Pinus* pollen were detected at depths of 23 and 27 cm, respectively (Krull et al., 2009). A detailed
146 description of the principles of ^{137}Cs and Pu radionuclide dating can be found in Krull et al. (2009).

147 **2.2. Analytical methods**

148 **2.2.1. Bulk stable isotope and TOC analysis**

149 Sediment samples were dried at 55 °C overnight, ground, homogenised and weighed into tin cups for
150 analysis. For carbonate removal sulfuric acid was added directly into the tin cups to prevent loss of acid-
151 soluble organic carbon (Verardo et al., 1990). After re-drying the samples were analysed for C content and $\delta^{13}\text{C}$
152 using a Carlo Erba NA1500 CNS analyser interfaced via a ConFlo II to a Finnigan Mat Delta S isotope ratio
153 mass spectrometer operating in the continuous flow mode. Combustion and oxidation were achieved at 1090 °C
154 and reduction at 650 °C. To ensure reproducibility, samples were analysed at least in duplicate. Results are
155 presented in the standard delta (δ) notation relative to Vienna Pee Dee Belemnite (VPDB).
156 TOC (%) was calculated from mass spectrometer peak areas using standards with a known C content. The
157 reproducibility of stable isotope measurements was ± 0.2 %.

158 **2.2.2. Sample preparation and extraction for biomarker analysis**

159 Samples were dried on a heated sandbath (55 °C) or in a cool oven (45 °C), powdered and extracted with
160 a mixture of 9:1 dichloromethane (DCM): methanol by accelerated solvent extraction (ASE) using a Dionex
161 ASE 200 (Dionex, Sunnyvale, CA, USA). The extracts were desulfurized overnight following addition of
162 activated copper powder. The filtered extracts were successively fractionated by silica gel chromatography
163 (column size: 20 cm x 0.9 cm I.D.) using approximately two column volumes of increasingly polar solvents.
164 Aliphatic and aromatic fractions relevant for this study were eluted with *n*-hexane and 30% dichloromethane in
165 *n*-hexane, respectively. An internal perdeuterated standard (*p*-terphenyl d^{14}) was added to the aromatic fraction
166 to assist semi quantitative analysis. Abundances of hopanoids, sterenes and steranes were calculated from peak
167 areas in m/z 191, m/z 215 and m/z 217 extracted ion chromatograms, respectively and reported relative to peak
168 areas of C_{21} - C_{31} *n*-alkanes from total ion chromatograms (TIC) without the consideration of response factors.

169 **2.2.3. Gas chromatography mass-spectrometry (GC-MS)**

170 GC-MS analyses were performed on an Agilent 5973 Mass-Selective Detector (MSD) interfaced to an
171 Agilent 6890 gas chromatograph (GC). For the separation of the analytes a capillary column (60 m x 0.25 mm)
172 coated with a 0.25 μm 5% phenyl 95% methyl polysiloxane stationary phase (DB-5MS, J & W scientific) was
173 used. The GC-oven was temperature programmed from 50 °C to 310 °C at a rate of 3 °C/min with initial and
174 final hold times of 1 and 20 minutes, respectively. Samples were injected (split/splitless injector) with a HP
175 6890 auto-sampler operated in a pulsed-splitless mode at 280 °C. Helium was used as the carrier gas at a

176 constant flow rate of 1.1 mL/min. Full scan (50 - 550 Daltons) 70 eV mass spectra were acquired typically with
177 an electron multiplier voltage of 1800 V and a source temperature of 230 °C.

178 **2.2.4. Gas chromatography-isotope ratio mass-spectrometry (GC-irMS)**

179 GC-irMS analyses were performed on a Micromass IsoPrime mass-spectrometer interfaced to either i) an
180 Agilent 6890 GC equipped with a 6890 autosampler for carbon isotope analysis; or ii) an Agilent 6890N GC
181 with 7683 autosampler for hydrogen isotope analysis. The GC- column, carrier gas, injector conditions and oven
182 temperature program were identical to the settings described for GC-MS. For $\delta^{13}\text{C}$ analyses an interface
183 consisting of a quartz tube packed with CuO-pellets (4 mm x 0.5 mm, isotope grade, Elemental Microanalysis
184 LTD.) maintained at 850 °C was used to oxidize the organic analytes to CO_2 and H_2O . The latter was
185 subsequently removed by a liquid nitrogen trap at -100 °C. Isotopic compositions were determined by
186 integration of the m/z 44, 45 and 46 ion currents of CO_2 peaks from each analyte and reported relative to CO_2
187 reference gas pulses of known ^{13}C -content. Isotopic values are given in the delta (δ) notation relative to the
188 international standard VPDB.

189 For hydrogen isotope (δD) analysis the gas chromatographically separated analytes were converted into
190 hydrogen gas in a pyrolysis furnace packed with chromium catalyst at 1050 °C. δD values were determined by
191 integration of the m/z 2 and 3 ion currents of H_2 peaks generated by the chromatographically separated analytes
192 and reported relative to H_2 reference gas pulses of known D/H content relative to Vienna Standard Mean Ocean
193 Water (VSMOW). A correction factor accounting for contributions from H_3^+ produced in the ion source was
194 determined by m/z 3 analyses at two different H_2 gas pressures.

195 Each sample was analysed in duplicate and average values and standard deviations were reported.
196 Standard deviations for C-CSIA in all reported results were < 0.5 ‰ and for H-CSIA always < 7 ‰ and in most
197 cases < 5 ‰. In house standard solutions containing a mixture of *n*-alkanes with a known isotopic composition
198 were regularly (after ~8 analyses) analysed to confirm accuracy of measured isotopic ratios.

199 Peaks of corresponding monoenes (if present) were included for the calculation of $\delta^{13}\text{C}$ and δD of *n*-alkanes
200 since co-elution prevented baseline separations. A further separation of the aliphatic fraction by procedures such
201 as 5A-molecular sieving (Grice et al., 2008a) or argentation-thin layer chromatography (TLC) was not possible
202 due to the limited amounts of sample material.

203 3. Results

204 Fig. 2 displays depth profiles of selected biomarker and stable isotope parameters indicative of changing
205 OM-sources in the North Lagoon in a chronological context. A pronounced excursion in the TOC profile shortly
206 below a depth of 20 cm marks the increased input of terrestrial OM being flushed into the lagoon during a large
207 flood event in 1956. A second excursion at ~ 50 cm might indicate an earlier and non-reported flood.

208 The first occurrence of *Pinus* pollen at a depth of 27 cm indicates the beginning of European influence in
209 the Coorong region and roughly coincides with the beginning of a restricted freshwater inflow into the North
210 Lagoon through the barrages (see section 2.1.1). Sediment ages also included in Fig. 2a were published by Krull
211 et al. (2009) and Fluin et al. (2007). Sediments < 20 cm, deposited after drastic alterations of the Coorong water
212 regime, are hereafter referred to as “recent sediments” and the remaining sediments deposited before a
213 significant European influence as “older sediments”.

214 3.1. Hydrocarbon abundances

215 3.1.1. *n*-Alkanes, *n*-alkenes and acyclic isoprenoids

216 The aliphatic fractions were dominated by a suite of mid- to long-chain *n*-alkanes (C₂₁-C₃₃) with a strong odd
217 over even predominance, typically maximising at C₂₇ (e.g. Fig. 3). In recent sediments (< 20 cm) their
218 distribution profiles showed significant alterations evident in decreasing carbon preference indexes (CPI, Eq. 1)
219 from an average of 6.13 (± 1.07 standard deviation) to 4.25 (± 0.43) and average chain length (ACL; Eq. 2) from
220 an average of 26.12 (± 0.29) to 25.89 (± 0.52; Fig. 2a). P_{aq} values (Eq. 3), determined according to a proxy to
221 distinguish input from terrestrial vs. aquatic plants (Ficken et al., 2000), varied between 0.37 and 0.61.

$$222 \quad \text{CPI} = \frac{\sum C_{\text{odd}}}{\sum C_{\text{even}}} \quad \text{Eq. 1}$$

223 where “C_{odd}” and “C_{even}” are peak areas from TIC chromatograms of *n*-alkanes with odd and even numbered
224 chain lengths, respectively over the range from C₂₁-C₃₀

$$225 \quad \text{ACL} = \frac{\sum (i \times C_i)}{\sum C_i} \quad \text{Eq. 2}$$

226 where “C_i” is the peak area of the *n*-alkane with carbon number “i” over the range from C₂₁-C₃₁

$$227 \quad \text{P}_{\text{aq}} = (C_{23} + C_{25}) / (C_{23} + C_{25} + C_{29} + C_{31}) \quad \text{Eq. 3}$$

228 where “C” is the peak area of *n*-alkanes

229 Another notable difference in recent sediments of the North Lagoon deposited after the ~1950s was a
230 significant increase in relative abundances of microalgal- and bacterial-derived hydrocarbons including the C₂₀
231 highly branched isoprenoid (HBI; **I**, see Appendix for structures), and short-chain *n*-alkanes (C₁₄-C₂₀, Figs 2a
232 and 3). Furthermore, the uppermost sediments (< 20 cm) contained high concentrations of unsaturated
233 aliphatics, with C₂₃-C₂₉ *n*-alkenes and a C₂₉ *n*-alkadiene of notably high abundance (Fig. 3).

234 **3.1.2. Hopanoids and steroids**

235 Bacterial-derived hopanoids (C₂₇-C₃₂) were abundant in all samples (Fig. 4a) and showed distributions
236 typical of immature sediments with a complex suite of hopenes (hop-17(21)-enes, *neohop*-13(18)-enes and hop-
237 22(29)-ene (diploptene)) and high abundances of 17β,21β-hopanes (biological configuration) relative to 17β,21α
238 and 17α,21β isomers. Most hopanoids exhibited highest abundances (*cf.* *n*-alkanes) in the uppermost 20 cm
239 indicating an increased bacterial input (e.g. Fig. 4a). The molecular distribution of hopanoids showed slight
240 variations throughout the core, however they were most pronounced in recent sediments.

241 The most abundant steroids in the aliphatic fractions were C₂₇-, C₂₈- and C₂₉-5α,14α,17α-20R-ster-2-
242 enes, which are early diagenetic products of biological steroids (e.g. Mackenzie et al., 1982). We also detected
243 low abundances of 5α,14α,17α-20R-cholestane, a product formed during slightly later stages of diagenesis
244 (Mackenzie et al., 1982), which showed increasing abundances with sediment age typical of a diagenetic
245 product (Fig. 4b). The ternary diagram in Fig. 5 showed distinct changes in ster-2-ene distributions in recent
246 sediments, with a significantly increased abundance of cholest-2-ene and other C₂₇-steroids.

247 **3.1.3. Other biomarkers**

248 The aliphatic fractions also contained a series of C₁₀-C₂₅ 1-chloro-*n*-alkanes which can be identified in
249 the *m/z* 91 mass chromatograms (Grossi and Raphael, 2003) illustrated in Fig. 6. This product series may extend
250 to higher MW homologues, but co-elutions with other analytes (i.e., alkenes, hopanoids and steroids) would
251 have compromised their detection. Nevertheless, C_{30:1} and C_{32:1} chloro-*n*-alkenes have been tentatively identified
252 by relative retention times which were consistent with previous reports (Zhang et al., 2011) and by mass spectral
253 features. The abundance profiles of the C₁₆-1-chloro-*n*-alkane (representative of other homologues) and the
254 putative C_{32:1} chloro-*n*-alkene throughout the core are displayed in Fig. 4c.

255 The PAH perylene (**II**) was only present in trace amounts in sediments of the uppermost 30 cm but
256 increased significantly in concentration at greater depths (Fig. 4d).

257 3.2. Stable isotope analysis

258 3.2.1. $\delta^{13}\text{C}$ analysis

259 $\delta^{13}\text{C}$ profiles of bulk organic carbon and selected biomarkers throughout the analysed core are displayed
260 in Fig. 2b. The profiles of mid to long-chain *n*-alkanes (and corresponding monoenes if present, see section
261 2.2.4) showed comparatively minor variations in older sediments representing several thousand years (25-115
262 cm; average $\delta^{13}\text{C}$: $-23.5 \pm 2.5\text{‰}$), but a marked negative shift of $\sim 5\text{‰}$ in recent sediments (< 20 cm; average
263 $\delta^{13}\text{C}$: $-28.2 \pm 2.0\text{‰}$), indicating significant changes in the ecosystem of the North Lagoon since the ~ 1950 s. The
264 $\delta^{13}\text{C}$ depth profiles for C_{17} - C_{20} *n*-alkanes and hopanoids largely resembled the longer chain *n*-alkanes, although
265 the isotopic shift in the most recent sediments is somewhat less pronounced, particularly compared to the odd-
266 chain *n*-alkanes (Fig. 2b). The bulk $\delta^{13}\text{C}_{\text{OM}}$ profile also exhibited a -5‰ shift coincident with this point but
267 showed an additional sharp negative spike shortly below 20 cm reflecting depleted terrestrial OM being flushed
268 into the system during the large flood event in 1956, consistent with the TOC-profile, however the second
269 potential flood event indicated by the TOC-profile was not evident in $\delta^{13}\text{C}_{\text{OM}}$ values. The direct effects of the
270 flood in 1956 were also not visible in the other depth profiles due to their lower resolution. The bulk $\delta^{13}\text{C}_{\text{OM}}$
271 signatures were generally slightly heavier than corresponding values obtained for *n*-alkanes, consistent with
272 lipids in autotrophs being generally more depleted compared to total biomass (Monson and Hayes, 1982). A
273 further difference in stable isotope ratios of *n*-alkanes between recent and older sediments were distinct
274 sawtooth patterns with ^{13}C -depleted even-carbon-numbered *n*-alkanes in older sediments (>25 cm), whereas
275 these features were absent in more recent samples (e.g. Fig. 3).

276 The $\delta^{13}\text{C}$ values of C_{17} - C_{20} *n*-alkanes (average $-29.1 \pm 2.5\text{‰}$), hop-17(21)-ene (average $-29.7 \pm 3.5\text{‰}$) and 17β ,
277 21β - 22R -homohopane (average $-31.8 \pm 2.2\text{‰}$) were generally lower than the values of $>\text{C}_{21}$ *n*-alkanes (average
278 $-24.8 \pm 3.1\text{‰}$). On the contrary, $\delta^{13}\text{C}$ signatures of the C_{29} *n*-alkadiene (average $-23.3 \pm 1.1\text{‰}$), which was
279 abundant only in recent sediments, were slightly more ^{13}C -enriched. The C_{20}HBI was significantly more ^{13}C -
280 enriched compared to the other hydrocarbons with an average $\delta^{13}\text{C}$ value of $-15.2 \pm 1.5\text{‰}$. Its sedimentary $\delta^{13}\text{C}$
281 depth profile also showed strong differences (Fig. 2b).

282 3.2.2. δD analysis

283 δD depth profiles of representative *n*-alkanes are displayed alongside other potential molecular salinity
284 indicators throughout the analysed core in Fig. 7. The measured δD values showed some variations but a general
285 trend towards a D-enrichment in younger sediments. The most pronounced positive shift was evident in recent

286 sediments (< 20 cm). Similar to $\delta^{13}\text{C}$ values, only the δD values of *n*-alkanes in older sediments (> 25cm)
287 exhibited a sawtooth pattern with D-enriched even-carbon-numbered *n*-alkanes (Fig. 3).

288 **4. Discussion**

289 **4.1. Salinity increase in the North Lagoon due to recent human impact**

290 The palaeosalinity markers displayed in Fig. 7 consistently indicate a recent increase in salinity following
291 the beginning of extensive human water-management in the 1950s, which was significantly more pronounced
292 than salinity variations over the previous thousands of years represented by the deeper sediments. In an
293 environmental setting like the Coorong lagoons, δD signatures of *in situ* produced biolipids generally reflect the
294 D/H abundance in the lake water (Sachse et al., 2006; Mügler et al., 2010), which is largely determined by the
295 extent of evaporation and freshwater input. Therefore more depleted δD values of these biolipids indicate lower
296 salinities. Since the *n*-alkanes detected here are largely produced by allochthonous sources (i.e. aquatic
297 macrophytes, algae and bacteria; see following sections), the general trend towards enriched δD values in
298 younger sediments, evident in all their profiles (e.g. Fig. 7), indicates an overall rising salinity, despite some
299 variations. Nevertheless, the most pronounced D-enrichment occurred in recent sediments (< 20 cm). The *n*-
300 alkanes deposited shortly after the flood event in 1956 showed a slight depletion in δD values, reflecting a
301 temporary freshening of the system. Similar variations in salinity, including the pronounced increase over the
302 past decades, have also been observed in studies of diatom assemblages in the North Lagoon (McKirdy et al.,
303 2010). Relatively large variations of δD values in older sediments also seem consistent with variations in aridity
304 in the Coorong region throughout the Late Holocene e.g. (Ahmad, 1996; Mee et al., 2007). Possibly these
305 changes would have been more pronounced in δD profiles at a higher time resolution. Furthermore, the shift in
306 δD values of *n*-alkanes in the recent sediments might also be enhanced by a change in source organisms, which
307 will be discussed in the following sections and is also indicated by the loss of the sawtooth pattern in δD
308 signatures of *n*-alkanes in recent sediments (Fig. 3).

309 Another potential parameter reflecting the rising salinity in recent sediments is the increasing abundance
310 of the hop-(17)21-ene relative to other hopanes (Fig. 7). Hop-(17)21-enes seem to be formed during early
311 diagenesis in hypersaline environments in preference to 17 α ,21 β -hopanes (ten Haven et al., 1985, 1988).
312 However, longer chain homologues (C₃₁-C₃₅), which are also indicative of high salinities (e.g. ten Haven et al.,
313 1985), were not abundant in the North Lagoon. The R₂₂-index, displaying the abundance of the C₂₂ *n*-alkane
314 relative to the C₂₁ and C₂₃ *n*-alkanes, can also be a marker for hypersaline depositional environments (ten Haven
315 et al., 1985, 1988). In the recent sediments from the North Lagoon it increases significantly, as it is expected for

316 rising salinities. However, the average recent sediment values of 0.32 are still much lower than those reported
317 by ten Haven et al. (1985) in hypersaline environments, which is to be expected since the increasing salinity in
318 the North Lagoon had not yet progressed to prevalent hypersalinity.

319 Unfortunately the parameters measured here do not reflect ecological changes of the most recent flood
320 event in 2010, since the analysed core was collected in 2005. However, the determined salinity indicators reflect
321 the high impact of the drastic reduction of freshwater inflow through the barrages, on the ecosystem of the North
322 Lagoon. The increase of salinity is generally considered the driving force for the observed ecological decline in
323 the Coorong.

324

325 **4.2. Changes of OM sources in the North Lagoon following human impact**

326 The marked shifts in the sedimentary depth profiles of $\delta^{13}\text{C}_{\text{OM}}$ and $\delta^{13}\text{C}$ of hydrocarbons (Fig. 2b),
327 which coincided with the intensification of human water management in the 1950s, reflect a major change in the
328 type of source organisms caused by the increasing salinity as well as the progressive isolation from riverine
329 inflow and seawater exchange. Although environmental factors can also influence $\delta^{13}\text{C}$ values in biomarkers
330 and sediments, this was most likely not the major cause for the isotopic shifts observed here. An elevated
331 salinity as reported in the Coorong over the past decades may influence $\delta^{13}\text{C}$ values, however it is usually
332 associated with heavier isotopic signatures in biolipids (Schidlowski et al., 1984; Grice et al., 1998; Andersen et
333 al., 2001) and the opposite (lighter isotopic signatures) is observed in the North Lagoon. Furthermore, an
334 increased abundance of dissolved CO_2 caused for example by enhanced heterotrophic activity after algal blooms
335 or by a slight decrease of the pH would lead to a ^{13}C -depletion in lipids of primary producers. With limited
336 supply of $[\text{CO}_2]_{\text{aq}}$ phytoplanktonic organisms show less discrimination against ^{13}C resulting in significantly
337 enriched biolipids whereas a high abundance of this carbon source leads to an enhanced discrimination and
338 more depleted biolipids (Takahashi et al., 1990; Freeman and Hayes, 1992). This effect may have also
339 intensified the observed $\delta^{13}\text{C}$ shifts in the North Lagoon. Although the Coorong remains an alkaline system
340 (Revill et al., 2009) temporary stratification with more acidic bottom waters (pH: 6.8) have been observed in the
341 past decades (Geddes, 2003).

342 Most shifts in the stable isotope profiles in the recent sediments described here were distinctly greater
343 in magnitude than natural variations over thousands of years, thus indicating unambiguous ecological changes in
344 recent sediments (younger than ~1950). However, it should be noted that sediments at greater depths of this core
345 had a very different time resolution (*cf.* radionuclide dates with radiocarbon age) and averaging effects were

346 much more significant for these compared to the recent sediments. Furthermore, bioturbation may have also
347 contributed to averaging of the signals.

348 Further indications for a change in source organisms was the disappearance of the sawtooth patterns in
349 the $\delta^{13}\text{C}$ and δD values of $>\text{C}_{21}$ *n*-alkanes (Fig. 3) in the recent sediments (< 20 cm). Similar sawtooth patterns
350 have been reported in various higher plant waxes due to different metabolic pathways used for the synthesis of
351 odd and even numbered homologues (Grice et al., 2008b; Zhou et al., 2010) as well as in sediments, where this
352 pattern may also be enhanced by or be the result of input from different source organisms such as a microalgal
353 bias to odd numbered *n*-alkanes (Logan et al., 1999; Grice et al., 2001; Dawson et al., 2004).

354 The isotopic differences evident throughout the studied core were also visible in a crossplot of $\delta^{13}\text{C}$ vs.
355 δD values of $>\text{C}_{21}$ *n*-alkanes in representative samples from the North Lagoon well before (105-110 cm) and
356 after (0-15 cm) human interferences had impacted the water regime (Fig. 8). Three groups (*n*-alkanes in recent
357 sediments and odd-chain and even-chain *n*-alkanes in older sediments) were clearly separated. The distinction
358 was however somewhat more obvious in $\delta^{13}\text{C}$ than in δD signatures since isotopic fractionations in different
359 metabolic pathways are usually reflected more strongly by $\delta^{13}\text{C}$ values (e.g. Chikaraishi and Naraoka, 2003;
360 Polissar and Freeman, 2010). The D-enrichment in the most recent samples is also a result of the rising salinity
361 and decreased freshwater input (see previous section).

362 **4.2.1. Significant input of aquatic macrophytes in older sediments**

363 The higher $\delta^{13}\text{C}$ values of $>\text{C}_{21}$ *n*-alkanes in older sediments are consistent with a predominant origin
364 from aquatic C3 plants such as seagrasses, which produce relatively ^{13}C -enriched hydrocarbons compared to
365 their terrestrial counterparts. Chikaraishi and Naraoka (2003) reported an average $\delta^{13}\text{C}$ value of *n*-alkanes
366 extracted from three marine seagrass species of -22.8 ± 1.0 ‰ and slightly more depleted values for aquatic
367 freshwater plants of -25.3 ± 1.9 ‰. There may also be a minor contribution from phytoplankton and a small
368 allochthonous input from terrestrial plants, potentially including species from halophytes (see section 4.4).

369 Aquatic plants mainly synthesise *n*-alkanes of mid-chain lengths typically maximising at C_{21} , C_{23} or C_{25}
370 (Botello and Mandelli, 1978; Cranwell, 1984; Ficken et al., 2000; Jaffé et al., 2001). Therefore the determined
371 ACLs of *n*-alkanes, which were lower than the 28 to 33 typical of terrestrial plants (Chikaraishi and Naraoka,
372 2003) are consistent with a major contribution from aquatic macrophytes. It is furthermore supported by the
373 range of P_{aq} values reported in these sediments, which are consistent with a major contribution from non-
374 emergent aquatic plants to the *n*-alkanes found in the North Lagoon. This is in accordance with the formerly
375 high abundance of aquatic macrophytes, in particular *Ruppia megacarpa*, which used to be a significant source

376 of OM before its recent extinction due to the rising salinity (Krull et al., 2009). However, particularly with a
377 microalgal contribution to sedimentary *n*-alkanes, which is likely the case here (see following section), P_{aq}
378 values have to be interpreted with care as the produced odd-chain *n*-alkanes might be in the same range to those
379 of aquatic plants.

380 A high contribution of aquatic macrophytes to the OM in older sediments in this and other cores from
381 the North Lagoon has also been suggested by Krull et al. (2008), to explain the lignin-rich but ^{13}C -enriched OM
382 in these samples. In recent sediments lignin was only present in trace amounts. However, relatively low
383 corresponding C/N values (8.4-10.1) reported in Krull et al. (2008), still indicate significant algal or
384 cyanobacterial input, although these values may be biased by a high input from inorganic nitrogen in the North
385 Lagoon, as noted by McKirdy et al. (2010).

386 Another indication for a significant contribution of macrophytes in the older sediments might be the
387 increasing concentration of perylene (Fig. 4d), which correlates with the relative lignin abundance reported by
388 Krull et al. (2008). However, the increase in concentrations with depth also reflects its diagenetic origin. Unlike
389 most other unsubstituted PAHs, which are mainly of a pyrogenic or thermogenic origin, perylene has been
390 associated with quinone pigments, particularly in wood/lignin degrading fungi but also in some other organisms
391 (Jiang et al., 2000; Grice et al., 2009; Suzuki et al., 2010), and typically shows different sedimentary abundances
392 compared to these combustion markers (Jiang et al., 2000; Atahan et al., 2007; Grice et al., 2009; Suzuki et al.,
393 2010). A link to fungal wood/lignin degradation is also supported by the absence of perylene in sediments and
394 crude oils predating the evolution of vascular plants and in marine sourced oils (Grice et al. 2009). The co-
395 occurrence of perylene and lignin in older sediments from the North Lagoon and the absence of both in recent
396 sediments do concur with the source correlation of perylene to fungal lignin degradation.

397 **4.2.2. Increased microalgal and bacterial input in recent sediments**

398 Although the ^{13}C depleted long-chain *n*-alkanes in the recent sediments of the North Lagoon are in the
399 range of terrestrial C3 plants waxes (e.g. Chikaraishi and Naraoka, 2003), a predominant origin from these
400 organisms seems implausible since a high allochthonous contribution to sedimentary OM in the North Lagoon,
401 especially in recent years, is unlikely. Several studies (using various analytical approaches) confirmed only a
402 minor terrestrial input to the sediments of the North Lagoon. These included biomarker and pigment analysis in
403 surface sediments and the water-column (Revill et al., 2009) as well as elemental (low C/N values; Krull et al.,
404 2009; McKirdy et al., 2010) and ^{13}C -NMR analysis (Krull et al., 2009) in sediment cores. According to Revill et
405 al. (2009) the primary source of OM in surface sediments of the North Lagoon were benthic diatoms and

406 cyanobacteria whereas OM in the water-column was mainly derived from green algae. Nevertheless, the
407 presence of pollen (e.g. *Pinus*, Krull et al., 2009) indicates a minor terrestrial input.

408 McKirdy et al. (2010) observed a similar negative shift in $\delta^{13}\text{C}$ values of *n*-alkanes in recent sediments
409 of the South Lagoon and attributed it to a primary source change from aquatic macrophytes to bacteria. Such a
410 source transition also seems plausible in the sediments analysed here. A ^{13}C -depletion would be expected from a
411 higher bacterial contribution to the long-chain *n*-alkanes, since also other bacterial-derived hydrocarbons such as
412 short-chain *n*-alkanes or hopanoids had isotopically lighter $\delta^{13}\text{C}$ signatures in these samples (Fig. 2b).

413 Long-chain *n*-alkanes with high CPIs are typically ascribed to vascular plants waxes (Eglinton and
414 Hamilton, 1967), but they do have other sources such as microalgae or bacteria, including sulfate reducers and
415 heterotrophs. Although most microalgae and bacteria predominantly produce *n*-alkanes of shorter chain lengths
416 (C_{14} - C_{20}), longer-chain *n*-alkanes or long-chain *n*-alkyl precursors such as fatty acids, *n*-alcohols or *n*-alkenes
417 have also been attributed to some species (e.g. Cranwell, 1982, 1984; Collister et al., 1994a; Lichtfouse et al.,
418 1994; Volkman et al., 1998; Logan et al., 1999; McKirdy et al., 2010). Whereas these bacteria (Oró et al., 1967;
419 Davis, 1968; Han and Calvin, 1969; Jones and Young, 1970) and some diatoms (Volkman et al., 1980 and
420 references therein; Nichols et al., 1988) generate long-chain *n*-alkane distributions without a pronounced odd-
421 over-even predominance, green algae often synthesise long-chain *n*-alkenes with odd carbon numbers, leading
422 to *n*-alkane distributions with high CPIs (Gelpi et al., 1970; Gelin et al., 1997; Allard and Templier, 2000). The
423 decreasing CPIs observed in the most recent sediments of the North Lagoon (Fig. 2a) could therefore indicate an
424 increasing abundance of bacteria, most likely sulphate reducers or heterotrophs and possibly diatoms. Low CPI
425 distributions are sometimes also associated with contamination from petroleum products (Peters et al., 2005).
426 However such a contamination in the Coorong seems unlikely since there has been no other indication for this in
427 the present project or in previous studies.

428 An increased phytoplanktonic contribution to the recent sedimentary OM is also consistent with the
429 significant increase of bacterial and microalgal biomarkers such as short chain *n*-alkanes (C_{14} - C_{20}), C_{27} - C_{29} -
430 alkadienes, the C_{20} HBI and some hopanes in the uppermost core-section (Figs 2a and 4a). C_{14} - C_{20} *n*-alkanes
431 often maximising at C_{17} are usually attributed to diatoms, green algae or bacteria, particularly cyanobacteria
432 (Han and Calvin, 1969; Gelpi et al., 1970; Cranwell, 1982). The relatively low $\delta^{13}\text{C}$ values of these *n*-alkanes in
433 the presently analysed sediments are in the same range as the $\delta^{13}\text{C}$ values of bacterial derived hopanoids (Fig.
434 2b), indicating a common source.

435 Long-chain *n*-alkenes with a strong odd-over-even predominance, which were abundant in sediments <
436 20 cm, are produced by some green microalgae and possibly cyanobacteria (Gelpi et al., 1970; Gelin et al.,
437 1997; Allard and Templier, 2000). The C₂₇-C₂₉ *n*-alkadienes were most likely derived from the A race of the
438 green algae *Botryococcus braunii* which is known to produce odd-carbon-numbered C₂₇-C₃₁ *n*-alkadienes and
439 minor amounts of the C₂₉ *n*-alkatriene (Metzger et al., 1986; Metzger and Largeau, 2005). *n*-Alkadienes of
440 similar chain lengths have also been isolated from some chlorococcales algae (Allard and Templier, 2000).
441 Although *Botryococcus braunii* is a freshwater species, there is reported evidence of its presence in hypersaline
442 environments, mainly due to salinity stratification in the water-column (e.g. Grice et al., 1998) which also
443 periodically occurs in the Coorong (Webster, 2005). Blooms of this species have previously been observed in
444 the Coorong region (Cane, 1976) and δ¹³C signatures of the C₂₉ *n*-alkadiene (-24.6 to -22.5 ‰) would also be
445 consistent with an origin from these algae (Grice et al., 1998). The high relative abundance of the unsaturated
446 aliphatics in recent sediments is in accordance with the high abundance of green algae in the water-column of
447 the present day Coorong (Revill et al., 2009). However, the difference to older sediments, in which these
448 compounds showed very low abundances or were absent, can also be partly attributed to their relatively unstable
449 structures resulting in an early transformation into more stable *n*-alkanes during diagenesis. A high abundance
450 of algal-derived OM in the most recent samples < 20 cm (possibly from the A race of *Botryococcus braunii*) is
451 also indicated by the relative ¹³C -enrichment of *n*-alkanes/alkenes > C₂₄ compared to their shorter chain
452 homologues (average difference: 4 ‰). However, the values were still lower than the δ¹³C signatures of the
453 algal-derived C₂₉ *n*-alkadiene, which indicates a mixed source of these *n*-alkanes in recent sediments with
454 contributions from bacteria as well as green algae and potentially also a minor input from terrestrial higher
455 plants. Revill et al. (2009) reported depleted values in green algal derived phytol in the water-column sometimes
456 approaching -30 ‰ which they attributed to a slow growth rate.

457 The enhanced abundance of the C₂₀HBI (Fig. 2a) presumably indicates an increased input from diatoms
458 or significant alterations in their population. The C₂₀HBI has been frequently assigned to a diatomaceous origin
459 based on stable isotopic compositions and abundance profiles as well as structural similarities to the C₂₅ and C₃₀
460 HBIs (III, IV), which are established biomarkers for these microalgae (Volkman et al., 1998; Atahan et al.,
461 2007; McKirdy et al., 2010). However, the C₂₀HBI or potential precursors have, unlike the larger analogues,
462 never been isolated from cultured organisms. The enriched δ¹³C values of the C₂₀HBI in sediments from the
463 North Lagoon (-18.8 to -13.9 ‰) are typical of lipids produced by diatoms since many species are capable of
464 assimilating enriched bicarbonate as opposed to more depleted CO₂ (Freeman and Hayes, 1992; Bieger et al.,

465 1997). Also the different sedimentary $\delta^{13}\text{C}$ depth profile of the C_{20}HBI compared to the other hydrocarbons,
466 reflect this different carbon source.

467 **4.2.3. Alterations in bacterial and microalgal populations in recent sediments**

468 The recent environmental changes in the North Lagoon, such as the increased salinity as well as the
469 enhanced turbidity caused by the extinction of macrophytes, also led to significant alterations in bacterial and
470 algal populations, which were evident in relative abundances and stable isotopic compositions of molecular
471 indicators for these organisms. The increased abundance of the C_{20}HBI (Fig. 2a) likely reflects significant
472 changes in diatom communities, which have also been observed in palynological analyses (Gell et al., 2007;
473 Haynes et al., 2007). In particular, the C_{20}HBI was the dominant aliphatic product in most sections of a core
474 from the South Lagoon, where conditions such as high salinities and a higher degree of isolation have prevailed
475 throughout the Holocene before being significantly enhanced by the recently employed interferences with the
476 water regime (McKirdy et al., 2010). The elevated abundance of the C_{20}HBI in recent sediments of the North
477 Lagoon indicates important changes in the environmental conditions and water regime, since a diatom
478 population similar to that previously present in the South Lagoon became supported.

479 Furthermore, changes in algal populations were reflected by variations in the steroid distributions.
480 Although C_{29} -desmethylsteroids are often assigned to a terrestrial plant origin (e.g. Huang and Meinschein,
481 1979) they have also been detected in significant amounts in microalgae, seagrasses and sediments without
482 higher plant input (Attaway et al., 1971; Volkman, 1986, 1998; Grosjean et al., 2009). Revill et al. (2009)
483 attributed all sterols they detected in water-column and surface sediments throughout the Coorong to aquatic
484 organisms, namely different types of green microalgae, diatoms (C_{29} - and C_{28} -sterols) and zooplankton (C_{27} -
485 sterols). The high relative abundance of cholest-2-ene in recent sediments of the North Lagoon may be an
486 indication for an increased population of zooplankton feeding on benthic microbial mats (McKirdy et al., 2010).
487 The primary source organisms of the C_{29} -ster-2-ene in the North Lagoon sediments are most likely microalgae
488 and aquatic plants (the latter particularly in older sediments) possibly with a minor contribution from terrestrial
489 plants. The relatively lower abundance of C_{29} -ster-2-enes in recent sediments might reflect the recent extinction
490 of aquatic macrophytes due to the rising salinities (Nicol, 2005; Krull et al., 2009; Dick et al., 2011). A slight
491 trend towards higher relative abundances of the C_{27} -ster-2-ene was already evident before the 1950s, as
492 indicated by the three black diamonds in the middle of the diagram, which correspond to samples between 25
493 and 40 cm. This likely indicates that already the smaller changes to the water-regime prior to the 1950s (see
494 section 2.1.1.) have had a noticeable impact on the algal or zooplankton communities and possibly macrophytes.

495 The clear shift in steroid distributions in the most recent samples also took place slightly earlier than the stable
496 isotopic shifts.

497 An indication for a change in bacterial populations are the negative $\delta^{13}\text{C}$ -shifts in the profiles of the C_{17}
498 *n*-alkane and the $17\beta,21\beta$ -22R-homohopane at 25 cm, although they were not as pronounced as in (odd-carbon-
499 numbered) long-chain *n*-alkanes (Fig. 2b). A possible explanation would be an increase of bacterial growth in
500 benthic microbial mats, often favoured in highly saline environments, which typically produce more ^{13}C -
501 depleted lipids compared to planktonic organisms, (Collister et al., 1994b; Freeman et al., 1994; Bieger et al.,
502 1997; Logan et al., 1999). Differences in bacterial population were also reflected by alteration in the hopanoid
503 distributions, which were most pronounced in the upper core-section.

504 **4.3. Significance of 1-chloro-*n*-alkanes**

505 Although a variety of organohalogenes are produced biosynthetically, in particular by marine organisms
506 but also by some terrestrial life forms including plants and higher animals (e.g. Gschwend et al., 1985; Gribble,
507 1996, 2003; Winterton, 2000), reports of mid- to long-chain 1-chloro-*n*-alkanes in sediments are sparse. Zhang
508 et al. (2011) detected $\text{C}_{30:1}$ and $\text{C}_{32:1}$ 1-chloro-*n*-alkenes, which we also tentatively identified in this study, in
509 sediments from a freshwater crater lake in the Galápagos Islands. A series of long-chain 1-chloro *n*-alkanes
510 (C_{19} - C_{29}) has been isolated from 3 genera (*Suaeda*, *Sarcocornia* and *Halimione*) of halophytic Chenopodiaceae
511 (Grossi and Raphel, 2003). Some of these species (including *Sarcocornia* and *Suaeda*) are also found in the
512 Coorong region (Boon, 2000 and references therein) and represent a potential source for the chlorinated
513 paraffins in the North Lagoon sediments. However, the series detected in this study did not show the odd-over-
514 even carbon number predominance that Grossi and Raphel (2003) had previously observed in halophytes and it
515 also comprised shorter-chain homologues (C_{12} - C_{18}), which have not been detected in the Chenopodiaceae.

516 Other sources of these compounds in the North Lagoon may include algae, seagrasses or cyanobacteria.
517 Although no long-chain-chloro-*n*-alkanes/ *n*-alkenes have been isolated from these organisms so far, they are
518 known to produce a variety of other chloro-organic compounds including volatile 1-chloro-*n*-alkanes in the
519 range from C_1 to C_5 (Mynderse and Moore, 1978; Gribble, 1996, 2003). The increased relative abundances of 1-
520 chloro-*n*-alkanes and 1-chloro *n*-alkenes in the most recent sediments of the North Lagoon (Fig.4c) point
521 towards an algal or cyanobacterial source since these organisms also were significantly more abundant in that
522 part of the core. Nevertheless, Chenopodiaceae have a high tolerance to salinity and may have become more
523 abundant in the lagoon catchment with the rising salinities in the Coorong.

524 Mid and long-chain chloro-paraffins (CPs; C₁₀ - C₃₀) are also known anthropogenic contaminants due to
525 many industrial applications (Tomy et al., 1998; Štejnarová et al., 2005). However, anthropogenically-sourced
526 CPs commonly exhibit complex distributions with various stereoisomers and different degrees of chlorination as
527 a result of the synthesis process (Tomy et al., 1998). Therefore the relative specificity of the distinct series of 1-
528 chloro-*n*-alkanes detected here precludes such an origin.

529 $\delta^{13}\text{C}$ values of the mono-chlorinated paraffins in the North Lagoon ranged from -31.8 to -24.5 ‰ (Table
530 1) and are consistent with both, a C3 plant source (such as the chloro-paraffin containing Chenopodiaceae
531 analysed by Grossi and Raphel (2003) and most Chenopodiaceae in the Coorong region) as well as with a
532 bacterial or microalgal origin (*cf.* $\delta^{13}\text{C}$ values of algal- and bacterial derived hydrocarbons).

533 **5. Conclusions**

534 The molecular and isotopic sedimentary record included evidence that human interference with the water
535 regime of the Coorong, namely a drastic reduction of the freshwater inflow due to installation of barrages and
536 extensive water abstractions from the River Murray, was immediately responsible for major changes in the types
537 of primary production, sedimentary OM and salinity in the North Lagoon over the past ~50 years. The
538 magnitude of these parameter changes has been significantly more pronounced than natural variations over
539 thousands of years. Aliphatic and aromatic biomarker analyses (including CSIA, C and H) in a sediment core
540 from the North Lagoon spanning more than 5000 years revealed changes in the populations of primary
541 producers contributing to sedimentary OM from predominantly macrophytes in sediments deposited prior to the
542 1950s towards bacteria and microalgae in more recent sediments. Furthermore, H-CSIA enabled the
543 reconstruction of salinity variation in the North Lagoon, showing dynamic changes with an overall rise of
544 salinity throughout the Holocene. However, a sharp increase took place shortly after the 1950s presumably due
545 to the restriction of freshwater inflow through the barrages. We also detected an interesting series of mid- to
546 long-chain 1-chloro-*n*-alkanes in these sediments. Potential sources of these compounds could be halophytic
547 Chenopodiaceae, cyanobacteria or microalgae.

548 **Acknowledgements**

549 Geoff Chidlow and Sue Wang are thanked for technical support. Curtin University is acknowledged for
550 providing a Curtin Strategic International Research Scholarship (CSIRS) to ST and University of Duisburg-
551 Essen for providing financial support. KG acknowledges the ARC for a QEII fellowship and infrastructure
552 grants. All authors acknowledge John de Laeter Centre for infrastructure support and The Institute for

553 Geoscience Researcher (TiGER) for a top-up scholarship. Two unknown reviewers are thanked for providing
554 helpful comments.

555 **References**

- 556 Ahmad, R., 1996. Late Holocene major Australian arid period revealed by direct sedimentological evidence
557 from lakes in the Coorong region of South Australia. *Geology* 24, 619-622.
- 558 Allard, B., Templier, J., 2000. Comparison of neutral lipid profile of various trilaminar outer cell wall (TLS)-
559 containing microalgae with emphasis on algaenan occurrence. *Phytochemistry* 54, 369-380.
- 560 Andersen, N., Paul, H.A., Bernasconi, S.M., McKenzie, J.A., Behrens, A., Schaeffer, P., Albrecht, P., 2001.
561 Large and rapid climate variability during the Messinian salinity crisis: Evidence from deuterium
562 concentrations of individual biomarkers. *Geology* 29, 799-802.
- 563 Atahan, P., Grice, K., Dodson, J., 2007. Agriculture and environmental change at Qingpu, Yangtze delta region,
564 China: a biomarker, stable isotope and palynological approach. *Holocene* 17, 507-515.
- 565 Attaway, D., Haug, P., Parker, P., 1971. Sterols in five coastal spermatophytes. *Lipids* 6, 687-691.
- 566 Bieger, T., Abrajano, T.A., Hellou, J., 1997. Generation of biogenic hydrocarbons during a spring bloom in
567 Newfoundland coastal (NW Atlantic) waters. *Organic Geochemistry* 26, 207-218.
- 568 Boon, P.I., 2000. Biological impacts of changes to water level and salinity in the Coorong. Melbourne, School
569 of Life Sciences and Technology, Victoria University.
- 570 Botello, A.V., Mandelli, E.F., 1978. Distribution of *n*-paraffins in sea-grasses, benthic algae, oysters and recent
571 sediments from terminos lagoon, Campeche, Mexico. *Bulletin of Environmental Contamination and*
572 *Toxicology* 19, 162-170.
- 573 Cane, R.F., 1976. The origin and formation of oil shale. In: eds Yen TF and Chilingarian GV (ed)
574 *Developments in Petroleum Science*, Elsevier, Amsterdam, pp 27-60
- 575 Chikaraishi, Y., Naraoka, H., 2003. Compound-specific δD - $\delta^{13}C$ analyses of *n*-alkanes extracted from terrestrial
576 and aquatic plants. *Phytochemistry* 63, 361-371.
- 577 Collister, J.W., Lichtfouse, E., Hieshima, G., Hayes, J.M., 1994a. Partial resolution of sources of *n*-alkanes in
578 the saline portion of the Parachute Creek Member, Green River Formation (Piceance Creek Basin,
579 Colorado). *Organic Geochemistry* 21, 645-659.
- 580 Collister, J.W., Rieley, G., Stern, B., Eglinton, G., Fry, B., 1994b. Compound-specific $\delta^{13}C$ analyses of leaf
581 lipids from plants with differing carbon dioxide metabolisms. *Organic Geochemistry* 21, 619-627.

582 Cranwell, P.A., 1982. Lipids of aquatic sediments and sedimenting particulates. *Progress in Lipid Research* 21,
583 271-308.

584 Cranwell, P.A., 1984. Lipid geochemistry of sediments from Upton Broad, a small productive lake. *Organic*
585 *Geochemistry* 7, 25-37.

586 Davis, J.B., 1968. Paraffinic hydrocarbons in the sulfate-reducing bacterium *Desulfovibrio desulfuricans*.
587 *Chemical Geology* 3, 155-160.

588 Dawson, D., Grice, K., Wang, S.X., Alexander, R., Radke, J., 2004. Stable hydrogen isotopic composition of
589 hydrocarbons in torbanites (Late Carboniferous to Late Permian) deposited under various climatic
590 conditions. *Organic Geochemistry* 35, 189-197.

591 Department of Environment and Heritage (DEH), 2000. Coorong and Lakes Alexandrina and Albert Ramsar
592 management plan. Government of South Australia, p. 63

593 Dick, J., Haynes, D., Tibby, J., Garcia, A., Gell, P., 2011. A history of aquatic plants in the Coorong, a Ramsar-
594 listed coastal wetland, South Australia. *Journal of Paleolimnology* 46, 623-635.

595 Eglinton, G., Hamilton, R.J., 1967. Leaf epicuticular waxes. *Science* 156, 1322-1335.

596 Ficken, K.J., Li, B., Swain, D.L., Eglinton, G., 2000. An *n*-alkane proxy for the sedimentary input of
597 submerged/floating freshwater aquatic macrophytes. *Organic Geochemistry* 31, 745-749.

598 Fluin, J., Gell, P., Haynes, D., Tibby, J., Hancock, G., 2007. Palaeolimnological evidence for the independent
599 evolution of neighbouring terminal lakes, the Murray Darling Basin, Australia. *Hydrobiologia* 591, 117-
600 134.

601 Freeman, K.H., Hayes, J.M., 1992. Fractionation of carbon isotopes by phytoplankton and estimates of ancient
602 CO₂ levels. *Global Biogeochemical Cycles* 6, 185-198.

603 Freeman, K.H., Wakeham, S.G., Hayes, J.M., 1994. Predictive isotopic biogeochemistry: hydrocarbons from
604 anoxic marine basins. *Organic Geochemistry* 21, 629-644.

605 Gat, J.R., 1996. Oxygen and hydrogen isotopes in the hydrologic cycle. *Annual Review of Earth and Planetary*
606 *Sciences* 24, 225-262.

607 Geddes, M.C., 2003. Survey to investigate the ecological health of the North and South Lagoons of the
608 Coorong, June/July 2003. Report Prepared for the Department of Environment and Heritage and the
609 Department of Water, Land and Biodiversity Conservation. South Australian Research and Development
610 Institute (Aquatic Sciences), Adelaide

611 Gelin, F., Boogers, I., Noordeloos, A.A.M., Sinnighe-Damsté, J.S., Riegman, R., De Leeuw, J.W., 1997.
612 Resistant biomacromolecules in marine microalgae of the classes Eustigmatophyceae and Chlorophyceae:
613 Geochemical implications. *Organic Geochemistry* 26, 659-675.

614 Gell, P., Haynes, D., Tibby, J., Dick, J., Hancock, G., 2007. A palaeoecological review of the ecological
615 character assessment for Ramsar listing: the case of the Coorong, South Australia. *Quat Int (Suppl)*. XVII
616 INQUA Congress 167–168:136

617 Gelpi, E., Schneider, H., Mann, J., Oró, J., 1970. Hydrocarbons of geochemical significance in microscopic
618 algae. *Phytochemistry* 9, 603-612.

619 Gribble, G.W., 1996. The diversity of natural organochlorines in living organisms. *Pure and Applied Chemistry*
620 68, 1699-1712.

621 Gribble, G.W., 2003. The diversity of naturally produced organohalogens. *Chemosphere* 52, 289-297.

622 Grice, K., Schaeffer, P., Schwark, L., Maxwell, J.R., 1996. Molecular indicators of palaeoenvironmental
623 conditions in an immature Permian shale (Kupferschiefer, Lower Rhine Basin, north-west Germany) from
624 free and S-bound lipids. *Organic Geochemistry* 25, 131-147.

625 Grice, K., Schouten, S., Nissenbaum, A., Charrach, J., Sinnighe-Damsté, J.S., 1998. A remarkable paradox:
626 Sulfurised freshwater algal (*Botryococcus braunii*) lipids in an ancient hypersaline euxinic ecosystem.
627 *Organic Geochemistry* 28, 195-216.

628 Grice, K., Audino, M., Boreham, C.J., Alexander, R., Kagi, R.I., 2001. Distributions and stable carbon isotopic
629 compositions of biomarkers in torbanites from different palaeogeographical locations. *Organic*
630 *Geochemistry* 32, 1195-1210.

631 Grice, K., Cao, C., Love, G.D., Böttcher, M.E., Twitchett, R.J., Grosjean, E., Summons, R.E., Turgeon, S.C.,
632 Dunning, W., Jin, Y., 2005. Photic zone euxinia during the Permian-Triassic superanoxic event. *Science*
633 307, 706-709.

634 Grice, K., Mesmay, R.d., Glucina, A., Wang, S., 2008a. An improved and rapid 5A molecular sieve method for
635 gas chromatography isotope ratio mass spectrometry of *n*-alkanes (C₈-C₃₀₊). *Organic Geochemistry* 39,
636 284-288.

637 Grice, K., Lu, H., Zhou, Y., Stuart-Williams, H., Farquhar, G.D., 2008b. Biosynthetic and environmental effects
638 on the stable carbon isotopic compositions of *anteiso*- (3-methyl) and *iso*- (2-methyl) alkanes in tobacco
639 leaves. *Phytochemistry* 69, 2807-2814.

640 Grice, K., Lu, H., Atahan, P., Asif, M., Hallmann, C., Greenwood, P., Maslen, E., Tulipani, S., Williford, K.,
641 Dodson, J., 2009. New insights into the origin of perylene in geological samples. *Geochimica et*
642 *Cosmochimica Acta* 73, 6531-6543.

643 Grosjean, E., Love, G.D., Stalvies, C., Fike, D.A., Summons, R.E., 2009. Origin of petroleum in the
644 Neoproterozoic–Cambrian South Oman Salt Basin. *Organic Geochemistry* 40, 87-110.

645 Grossi, V., Raphel, D., 2003. Long-chain (C₁₉-C₂₉) 1-chloro-*n*-alkanes in leaf waxes of halophytes of the
646 Chenopodiaceae. *Phytochemistry* 63, 693-698.

647 Gschwend, P.M., Macfarlane, J.K., Newman, K.A., 1985. Volatile halogenated organic compounds released to
648 seawater from temperate marine macroalgae. *Science* 227, 1033-1035.

649 Han, J., Calvin, M., 1969. Hydrocarbon distribution of algae and bacteria and microbiological activity in
650 sediments. *Proceedings of the National Academy of Sciences* 64, 436-443.

651 Haynes, D., Gell, P., Hancock, G., 2007. Diatom inferred changes to water quality and source, the Coorong,
652 South Australia. *Quat Int (Suppl)*. XVII INQUA Congress 167–168:158

653 Huang, W.Y., Meinschein, W.G., 1979. Sterols as ecological indicators. *Geochimica et Cosmochimica Acta* 43,
654 739-745.

655 Ingram, B.L., Conrad, M.E., Ingle, J.C., 1996. Stable isotope and salinity systematics in estuarine waters and
656 carbonates: San Francisco Bay. *Geochimica et Cosmochimica Acta* 60, 455-467.

657 Jaffé, R., Mead, R., Hernandez, M.E., Peralba, M.C., DiGuida, O.A., 2001. Origin and transport of sedimentary
658 organic matter in two subtropical estuaries: a comparative, biomarker-based study. *Organic Geochemistry*
659 32, 507-526.

660 Jiang, C., Alexander, R., Kagi, R.I., Murray, A.P., 2000. Origin of perylene in ancient sediments and its
661 geological significance. *Organic Geochemistry* 31, 1545-1559.

662 Jones, J.G., Young, B.V., 1970. Major paraffin constituents of microbial cells with particular references to
663 *Chromatium* sp. *Archives of Microbiology* 70,82-88.

664 Kingsford, R.T., Walker, K.F., Lester, R.E., Young, W.J., Fairweather, P.G., Sammut, J., Geddes, M.C., 2011.
665 A Ramsar wetland in crisis - the Coorong, Lower Lakes and Murray Mouth, Australia. *Marine and*
666 *Freshwater Research* 62, 255-265.

667 Krull, E., Haynes, D., Lamontagne, S., Gell, P., McKirdy, D., Hancock, G., McGowan, J., Smernik, R., 2009.
668 Changes in the chemistry of sedimentary organic matter within the Coorong over space and time.
669 *Biogeochemistry* 92, 9-25.

670 Lichtfouse, É., Derenne, S., Mariotti, A., Largeau, C., 1994. Possible algal origin of long chain odd *n*-alkanes in
671 immature sediments as revealed by distributions and carbon isotope ratios. *Organic Geochemistry* 22,
672 1023-1027.

673 Logan, G.A., Calver, C.R., Gorjan, P., Summons, R.E., Hayes, J.M., Walter, M.R., 1999. Terminal Proterozoic
674 mid-shelf benthic microbial mats in the Centralian Superbasin and their environmental significance.
675 *Geochimica et Cosmochimica Acta* 63, 1345-1358.

676 Mackenzie, A.S., Brassell, S.C., Eglinton, G., Maxwell, J.R., 1982. Chemical fossils: the geological fate of
677 steroids. *Science* 217, 491-504.

678 Maheshwari, B.L., Walker, K.F., McMahon, T.A., 1995. Effects of regulation on the flow regime of the river
679 Murray, Australia. *Regulated Rivers: Research & Management* 10:15-38.

680 McKirdy, D.M., Thorpe, C.S., Haynes, D.E., Grice, K., Krull, E.S., Halverson, G.P., Webster, L.J., 2010. The
681 biogeochemical evolution of the Coorong during the mid- to late Holocene: An elemental, isotopic and
682 biomarker perspective. *Organic Geochemistry* 41, 96-110.

683 Mee, A.C., McKirdy, D.M., Williams, M.A.J., Krull, E.S., 2007. New radiocarbon dates from sapropels in three
684 Holocene lakes of the Coorong coastal plain, southeastern Australia. *Australian Journal of Earth Sciences*
685 54, 825-835.

686 Metzger, P., Largeau, C., 2005. *Botryococcus braunii*: a rich source for hydrocarbons and related ether lipids.
687 *Applied Microbiology and Biotechnology*, 66, 486-496.

688 Metzger P., Templier J., Largeau C., Casadevall E. (1986) An *n*-alkatriene and some *n*-alkadienes from the A
689 race of the green alga *Botryococcus braunii*. *Phytochemistry* 25, 1869-1872.

690 Monson, K.D., Hayes, J.M., 1982. Carbon isotopic fractionation in the biosynthesis of bacterial fatty acids.
691 Ozonolysis of unsaturated fatty acids as a means of determining the intramolecular distribution of carbon
692 isotopes. *Geochimica et Cosmochimica Acta* 46, 139-149.

693 Mügler, I., Gleixner, G., Günther, F., Mäusbacher, R., Daut, G., Schütt, B., Berking, J., Schwalb, A., Schwark,
694 L., Xu, B., Yao, T., Zhu, L., Yi, C., 2010. A multi-proxy approach to reconstruct hydrological changes and
695 Holocene climate development of Nam Co, Central Tibet. *Journal of Paleolimnology* 43, 625-648.

696 Mynderse, J.S., Moore, R.E., 1978. The isolation of (-)-E-1-chlorotridec-1-ene-6,8-diol from a marine
697 cyanophyte. *Phytochemistry* 17, 1325-1326.

698 Nabbefeld, B., Grice, K., Twitchett, R.J., Summons, R.E., Hays, L., Böttcher, M.E., Asif, M., 2010. An
699 integrated biomarker, isotopic and palaeoenvironmental study through the Late Permian event at
700 Lusitaniadalen, Spitsbergen. *Earth and Planetary Science Letters* 291, 84-96.

701 Nichols, P.D., Palmisano, A.C., Volkman, J.K., Smith, G.A., White, D.C., 1988. Occurrence of an isoprenoid
702 C₂₅ diunsaturated alkene and high neutral lipid content in Antarctic sea-ice diatom communities. *Journal of*
703 *Phycology* 24, 90-96.

704 Nicol, J., 2005. The ecology of *Ruppia* spp. in South Australia, with reference to the Coorong. A literature
705 review. SARDI Aquatic Sciences Publication No. RD04/0247-2. SARDI Research Report Series No. 88.

706 Oró, J., Tornabene, T.G., Nooner, D.W., Gelpi, E., 1967. Aliphatic hydrocarbons and fatty acids of some marine
707 and freshwater microorganisms. *Journal of Bacteriology* 93, 1811-1818

708 Paton, D.C., Rogers, D.J., Hill, B.M., Bailey, C.P., Ziemicki, M., 2009. Temporal changes to spatially
709 stratified waterbird communities of the Coorong, South Australia: implications for the management of
710 heterogeneous wetlands. *Animal Conservation* 12, 408-417.

711 Peters, K.E., Walters, C.C., Moldowan, J.M., 2005. The biomarker guide: Interpreting molecular fossils in
712 petroleum and ancient sediments. Prentice-Hall, New Jersey

713 Polissar, P.J., Freeman, K.H., 2010. Effects of aridity and vegetation on plant-wax δD in modern lake sediments.
714 *Geochimica et Cosmochimica Acta* 74, 5785-5797.

715 Revill, A.T., Leeming, R., Volkman, J.K., Clementson, L., 2009. Sources of Organic Matter In the Coorong.
716 CSIRO: Water for a Healthy Country National Research Flagship

717 Sachse, D., Radke, J., Gleixner, G., 2006. δD values of individual *n*-alkanes from terrestrial plants along a
718 climatic gradient – Implications for the sedimentary biomarker record. *Organic Geochemistry* 37, 469-483.

719 Sauer, P.E., Eglinton, T.I., Hayes, J.M., Schimmelmann, A., Sessions, A.L., 2001. Compound-specific D/H
720 ratios of lipid biomarkers from sediments as a proxy for environmental and climatic conditions.
721 *Geochimica et Cosmochimica Acta* 65, 213-222.

722 Schidlowski, M., Matzigkeit, U., Krumbein, W.E., 1984. Superheavy organic carbon from hypersaline microbial
723 mats. *Naturwissenschaften* 71, 303-308.

724 Sessions, A.L., Burgoyne, T.W., Schimmelmann, A., Hayes, J.M., 1999. Fractionation of hydrogen isotopes in
725 lipid biosynthesis. *Organic Geochemistry* 30, 1193-1200.

726 Shuttleworth, B., Woidt, A., Paparella, T., Herbig, S., Walker, D., 2005. The dynamic behaviour of a river-
727 dominated tidal inlet, River Murray, Australia. *Estuarine Coastal and Shelf Science* 64, 645-657.

728 Štejnarová, P., Coelhan, M., Kostrhounová, R., Parlar, H., Holoubek, I., 2005. Analysis of short chain
729 chlorinated paraffins in sediment samples from the Czech Republic by short-column GC/ECNI-MS.
730 Chemosphere 58, 253-262.

731 Suzuki, N., Yessalina, S., Kikuchi, T., 2010. Probable fungal origin of perylene in Late Cretaceous to Paleogene
732 terrestrial sedimentary rocks of northeastern Japan as indicated from stable carbon isotopes. Organic
733 Geochemistry 41, 234-241.

734 Takahashi, K., Yoshioka, T., Wada, E., Sakamoto, M., 1990. Temporal variations in carbon isotope ratio of
735 phytoplankton in a eutrophic lake. Journal of Plankton Research 12, 799-808.

736 ten Haven, H.L., De Leeuw, J.W., Schenck, P.A., 1985. Organic geochemical studies of a Messinian evaporitic
737 basin, northern Apennines (Italy) I: Hydrocarbon biological markers for a hypersaline environment.
738 Geochimica et Cosmochimica Acta 49, 2181-2191.

739 ten Haven, H.L., de Leeuw, J.W., Sinninghe-Damsté, J.S., Schenck, P.A., Palmer, S.E., Zumberge, J.E., 1988.
740 Application of biological markers in the recognition of palaeohypersaline environments. Geological
741 Society, London, Special Publications 40:123-130.

742 Tibby, J., 2003. Explaining lake and catchment change using sediment derived and written histories: an
743 Australian perspective. Science of The Total Environment 310, 61-71.

744 Tomy, G.T., Fisk, A.T., Westmore, J.B., Muir, D.C.G., 1998. Environmental chemistry and toxicology of
745 polychlorinated *n*-alkanes. Reviews of Environmental Contamination and Toxicology, pp. 53-128.

746 Verardo, D.J., Froelich, P.N., McIntyre, A., 1990. Determination of organic carbon and nitrogen in marine
747 sediments using the Carlo Erba NA-1500 analyzer. Deep-Sea Research 37,157-165.

748 Volkman, J.K., Johns, R.B., Gillan, F.T., Perry, G.J., Bavor Jr., H.J., 1980. Microbial lipids of an intertidal
749 sediment - I. Fatty acids and hydrocarbons. Geochimica et Cosmochimica Acta 44, 1133-1143.

750 Volkman, J.K., 1986. A review of sterol markers for marine and terrigenous organic matter. Organic
751 Geochemistry 9, 83-99.

752 Volkman, J.K., Barrett, S.M., Blackburn, S.I., Mansour, M.P., Sikes, E.L., Gelin, F., 1998. Microalgal
753 biomarkers: a review of recent research developments. Organic Geochemistry 29, 1163-1179.

754 Volkman, J.K., Revill, A.T., Holdsworth, D.G., Fredericks, D., 2008. Organic matter sources in an enclosed
755 coastal inlet assessed using lipid biomarkers and stable isotopes. Organic Geochemistry 39, 689-710.

756 Webster, I.T., 2005. An Overview of the Hydrodynamics of the Coorong and Murray Mouth. Technical Report
757 No. #/2005. CSIRO: Water for a Healthy Country National Research Flagship.

758 Webster, I.T., 2010. The hydrodynamics and salinity regime of a coastal lagoon – The Coorong, Australia –
759 Seasonal to multi-decadal timescales. *Estuarine Coastal and Shelf Science* 90, 264-274.
760 Winterton, N., 2000. Chlorine: the only green element - towards a wider acceptance of its role in natural cycles.
761 *Green Chemistry* 2, 173-225.
762 Zhang, Z., Metzger, P., Sachs, J.P., 2011. Co-occurrence of long chain diols, keto-ols, hydroxy acids and keto
763 acids in recent sediments of Lake El Junco, Galápagos Islands. *Organic Geochemistry* 42, 823-837.
764 Zhou, Y., Grice, K., Stuart-Williams, H., Farquhar, G.D., Hocart, C.H., Lu, H., Liu, W., 2010. Biosynthetic
765 origin of the saw-toothed profile in $\delta^{13}\text{C}$ and $\delta^2\text{H}$ of *n*-alkanes and systematic isotopic differences between
766 *n*-, *iso*- and *anteiso*-alkanes in leaf waxes of land plants. *Phytochemistry* 71, 388-403.

767

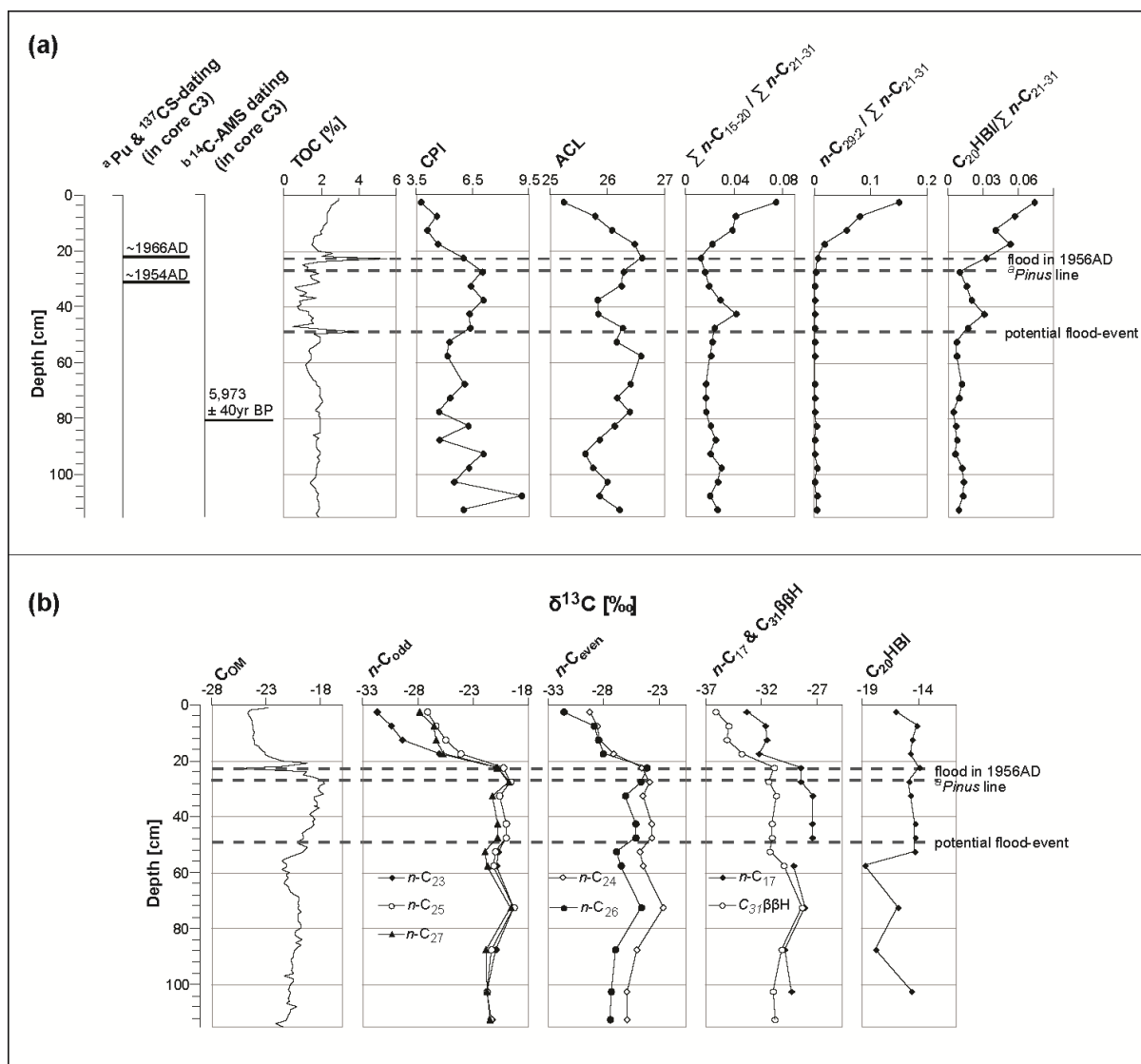
768 **Figure captions**



769

770 **Fig. 1** Location of the Coorong including the sampling site

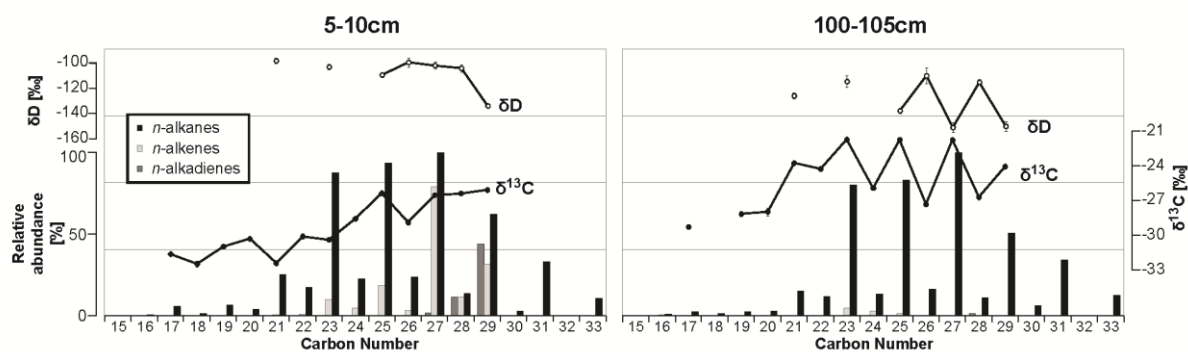
771



772

773 **Fig. 2** Alterations in TOC [%], *n*-alkane distributions (carbon preference index (CPI), average chain length
 774 (ACL)) and abundances of selected biomarkers relative to long-chain *n*-alkanes (a) and δ¹³C- profiles in a core
 775 from the northern Coorong Lagoon (b) put into perspective with events in the region and inferred sediment ages
 776 from ¹³⁷Cs, Pu and ¹⁴C-AMS dating (dating was performed on a neighbouring core). Flood events were marked
 777 according to excursions in the TOC profile. The first occurrence of *Pinus* pollen (*Pinus* line) indicates the
 778 beginning of European influence and roughly coincides with the 1950s (Krull et al., 2009). *n*-C_{*i*} = *n*-alkane with
 779 chain length “*i*”; C₃₁ββH = 17β,21β-22R-homohopane; CPI = $\frac{\sum C_{\text{odd}}}{\sum C_{\text{even}}}$ over the range from C₂₁-C₃₀, where
 780 “C_{odd}” and “C_{even}” are peak areas from TIC chromatograms of *n*-alkanes with odd and even numbered chain
 781 lengths, respectively; ACL = $\frac{\sum (i \times C_i)}{\sum C_i}$ where “C_{*i*}” is the peak area of the *n*-alkane with carbon number “*i*”
 782 over the range from C₂₁-C₃₁; ^a from Krull et al., (2009), ^b from Fluin et al. (2007). OM = organic matter

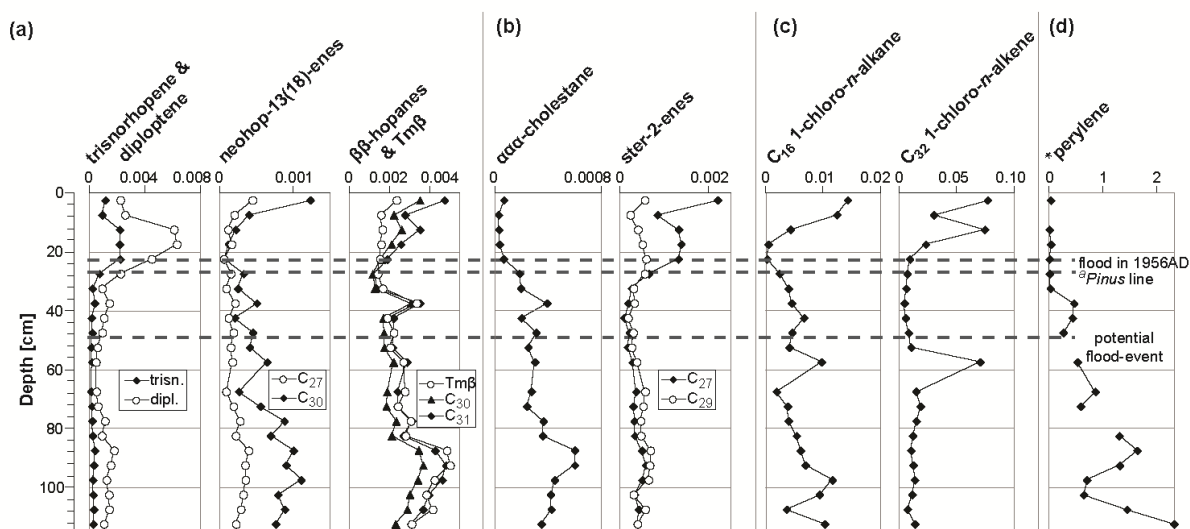
783



784

785 **Fig. 3** Typical distributions and stable isotopic compositions ($\delta^{13}\text{C}$, δD) of *n*-alkanes, *n*-alkenes and *n*-alkadienes
 786 in sediments from the northern Coorong Lagoon before (depth: 100-105cm) and after (depth: 5-10 cm) human
 787 interference with the water-regime. Relative abundances were determined from peak areas in the TIC trace of
 788 GC-MS chromatograms and calculated relative to the most abundant compound. Stable isotopic compositions
 789 include *n*-alkanes and their corresponding monoenes. Error bars indicate standard deviations of 2 replicate
 790 analyses, where error bars are not visible their size is smaller than the symbol

791



792

793 **Fig. 4** Alterations of biomarker abundances relative to $\Sigma\text{C}_{21}\text{-C}_{31}$ *n*-alkanes in a core from the North Lagoon.

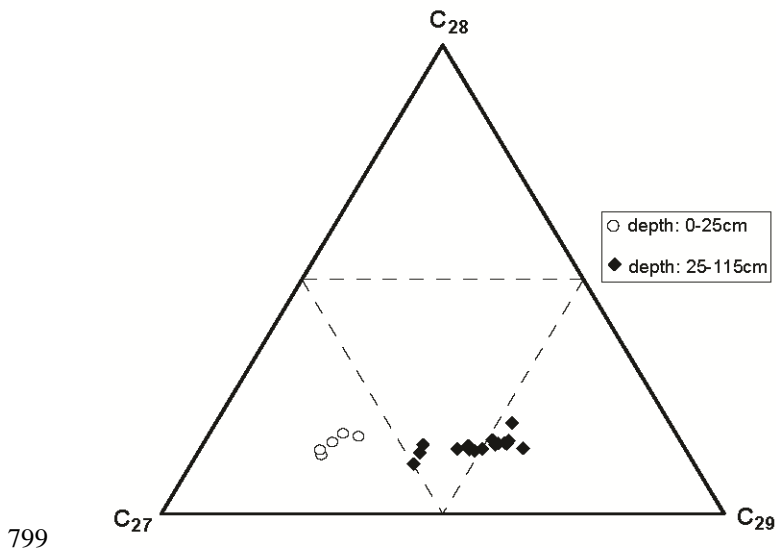
794 *Concentration in dry weight sediment [ng/g]. The first occurrence of *Pinus* pollen (*Pinus* line) indicates the

795 beginning of European influence and roughly coincides with the 1950s (Krull et al., 2009). The corresponding

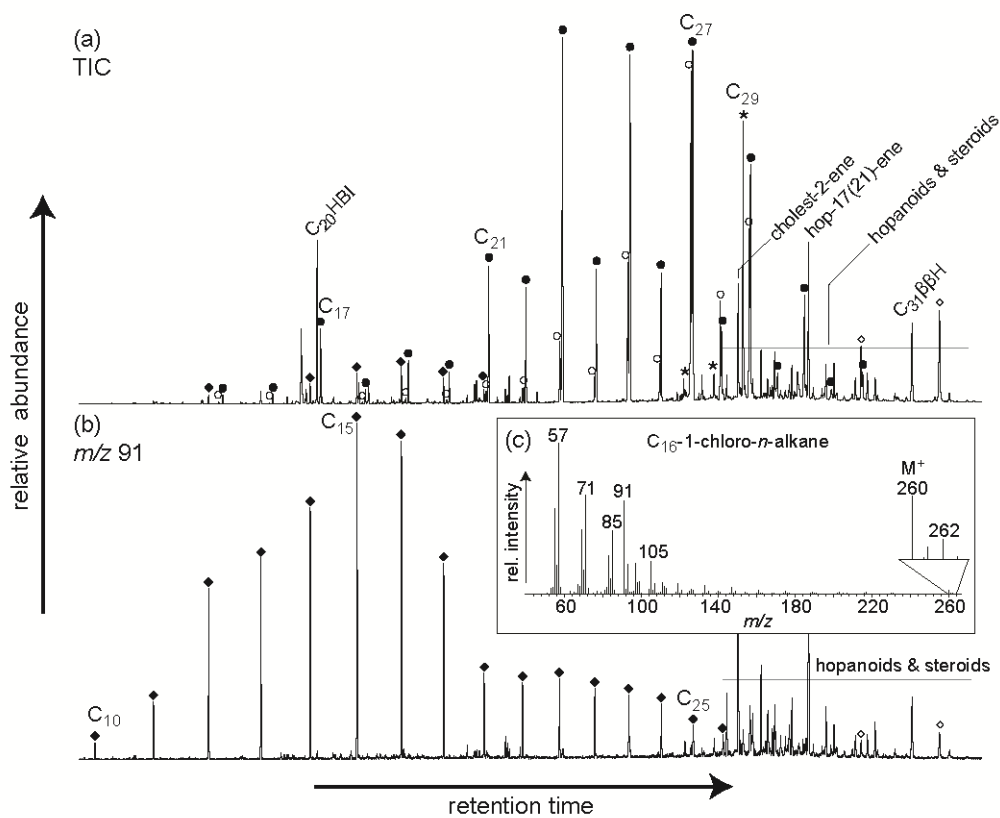
796 depths of flood-events were determined from excursions in the TOC profile. ^a from Krull et al. (2009). Tm-β =

797 17β-22,29,30-trisnorhopane

798

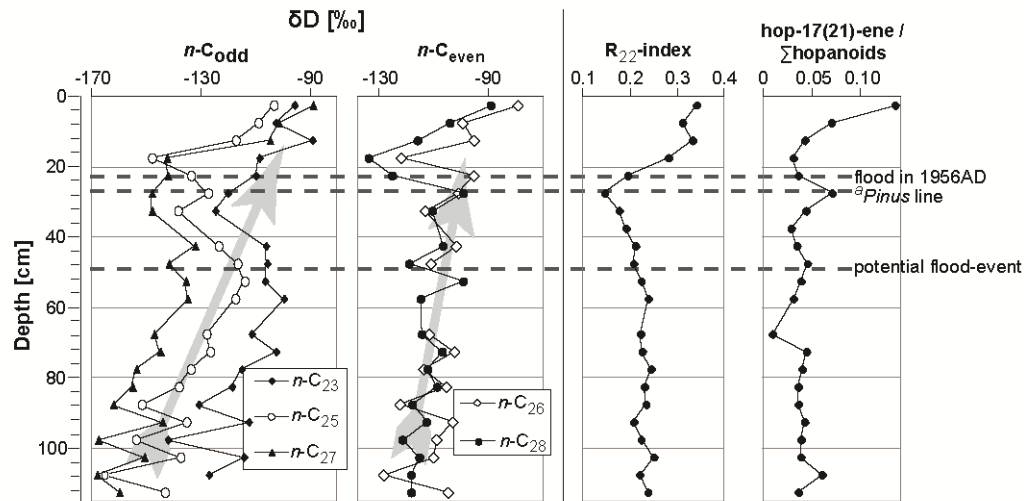


799
 800 **Fig. 5** Ternary diagram of ster-2-enes and cholestane (only steroids in the aliphatic fractions) in a core from the
 801 northern Coorong Lagoon. Sediments between 0-25 cm were deposited after the ~1950s, a period when human
 802 interference with the water regime intensified significantly. Deeper sediments were deposited over several
 803 thousands of years.
 804



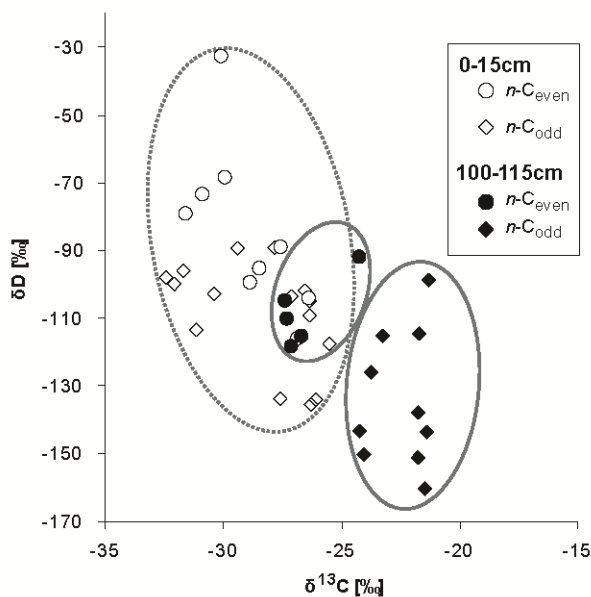
805
 806 **Fig. 6** Total ion chromatogram (TIC, (a)) and m/z 91 extracted ion chromatogram (b) of the aliphatic fraction in
 807 sediments from the North Lagoon between depths of 0-5 cm. (c) is the EI mass spectrum of the C_{16} -1-chloro- n -

808 alkane. C_i : carbon number of displayed compound classes; ●: *n*-alkanes; ○: *n*-alkenes; * : *n*-alkadienes; ◆: 1-
 809 chloro-*n*-alkanes; ◇: tentatively identified 1-chloro-*n*-alkenes; $C_{31}\beta\beta H = 17\beta,21\beta$ -22R-homohopane
 810



811
 812 **Fig. 7** δD profiles of representative *n*-alkanes with odd and even chain lengths (n - C_{even} and n - C_{odd} , respectively)
 813 and other potential salinity indicators throughout a core from the North Lagoon. $R_{22}\text{-index} = 2 \times C_{22} / (C_{21} + C_{23})$
 814 where " C_i " is the peak area in GC-MS chromatograms of the *n*-alkane with chain length " i " (ten Haven et al.,
 815 1988). The first occurrence of *Pinus* pollen (*Pinus* line) indicates the beginning of European influence and
 816 roughly coincides with the 1950s (Krull et al., 2009). The corresponding depths of flood-events were
 817 determined from excursions in the TOC profile. ^a from Krull et al. (2009)

818



819

820 **Fig. 8** Crossplot of $\delta^{13}\text{C}$ vs. δD of long-chain *n*-alkanes ($\text{C}_{21}\text{-C}_{29}$) in representative sediments from the northern
821 Coorong lagoon deposited prior (100-115 cm) and significantly after (0-15 cm) the beginning of human control
822 over the water regime in the Coorong. $n\text{-C}_{\text{even}}$ and $n\text{-C}_{\text{odd}}$ stands for *n*-alkanes with an even- and odd-numbered
823 carbon chain, respectively. The large circles are for enhanced visibility of the different groups, however no
824 cluster analysis has been performed.

825

826 **Table**

827 **Table 1** $\delta^{13}\text{C}$ [‰] \pm standard deviation of 2 replicates of 1-chloro-*n*-alkanes/*n*-alkenes throughout a sediment
828 core from the North Lagoon. C_i = carbon number

829

Depth [cm]	1-chloro- <i>n</i> -alkanes			1-chloro- <i>n</i> -alkene
	C_{15}	C_{16}	C_{17}	$\text{C}_{32:1}$
0-5	-29.0 \pm 0.0	-29.2 \pm 0.4	-31.8 \pm 0.1	-29.8 \pm 0.1
5-10	-28.1 \pm 0.1	-26.5 \pm 0.0	-27.8 \pm 0.2	-26.6 \pm 0.3
10-15	-29.1 \pm 0.2	-30.4 \pm 0.3	n.d.	-25.4 \pm 0.2
15-20	n.d.	n.d.	n.d.	-26.1 \pm 0.4
25-30	n.d.	-26.9 \pm 0.1	n.d.	n.d.
30-35	n.d.	n.d.	-24.8 \pm 0.4	n.d.
40-45	-24.5 \pm 0.3	-27.4 \pm 0.5	n.d.	n.d.
45-50	n.d.	n.d.	n.d.	n.d.
50-55	n.d.	-26.4 \pm 0.0	n.d.	n.d.
55-60	-25.1 \pm 0.2	-25.7 \pm 0.1	-26.0 \pm 0.2	n.d.
70-75	-25.5 \pm 0.2	-24.8 \pm 0.3	-26.4 \pm 0.1	n.d.
85-90	n.d.	-26.0 \pm 0.2	-26.6 \pm 0.1	n.d.
100-105	-25.8 \pm 0.5	-26.1 \pm 0.2	-27.3 \pm 0.5	n.d.
110-115	-26.3 \pm 0.5	-26.3 \pm 0.4	-25.9 \pm 0.0	n.d.

n.d. = not determined

830

831

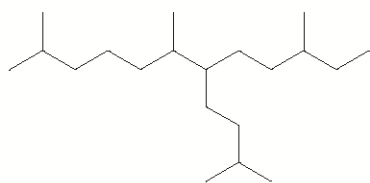
832

833

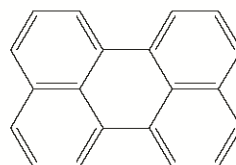
834

835

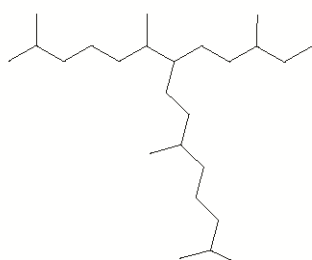
C_{20} HBI (I)



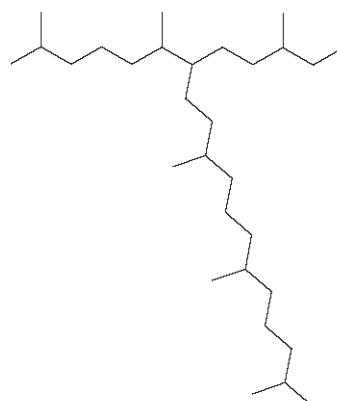
perylene (II)



C_{25} HBI (III)



C_{30} HBI (IV)



837

838 **Fig. A1** Structures referred to in the text