

1 **Estimates of exploitation rates of the spiny lobster fishery for *Panulirus argus* from tagging**  
2 **within the Bahía Espíritu Santo “Sian Ka’an” Biosphere Reserve, Mexican Caribbean**

3 Ley-Cooper Kim.<sup>1\*</sup>, de Lestang, Simon.<sup>2</sup>, Phillips, Bruce. F.<sup>1</sup> & Lozano-Álvarez, Enrique<sup>3</sup>

4 <sup>1</sup>Department of Environment and Agriculture, Curtin University, Western Australia.

5 <sup>2</sup>Department of Fisheries, Western Australia.

6 <sup>3</sup>Reef Systems Unit at Puerto Morelos, Institute of Marine Sciences and Limnology, National  
7 Autonomous University of Mexico.

8 \*Corresponding author:

9 MSc. Kim Ley-Cooper

10 Aquatic Science Research Unit, Curtin University of Technology, Brodie Hall Building, 1  
11 Turner Avenue, Technology Park, Bentley, Western Australia 6102,

12 Phone: +52 15554381590. Fax +525555542015

13 e-mail: [kim.ley@postgrad.curtin.edu.au](mailto:kim.ley@postgrad.curtin.edu.au)

14 Running head: Exploitation rates of *Panulirus argus* in Mexico

15 **ABSTRACT** The Caribbean spiny lobster (*Panulirus argus*) fishery is currently being assessed  
16 for a certification process. It is the main economic activity within the Biosphere Reserve of Sian  
17 Ka’an-Mexico (SK), which is a marine protected area where restricted access fishing is allowed.  
18 In this study, commercial catch rates were examined, and lobsters over a wide size range were  
19 tagged throughout the 2010/2011 fishing season, to assess fishing mortality rates and movement  
20 patterns in Bahía Espíritu Santo-SK. Lobster tag recovery data were aggregated into two-week

21 periods and analysed using a modified Brownie model that was parameterized to account for  
22 lobster tag-reporting, and the lobster tag-retention rates. This allowed the estimation of  
23 instantaneous rates of lobster natural and fishing mortality, considering catchability and fishing  
24 effort. Independent aquaria trials were conducted to better estimate lobster tag retention, tagging  
25 induced mortality and interviews with fishers were conducted to better estimate lobster tag  
26 reporting. Based mainly on legal-sized juveniles with fast growth rates found in casitas, the  
27 stock subject to fishing is limited to a maximum depth of 20 m, because of the prohibition of  
28 SCUBA diving and the use of other fishing gear. The Brownie model indicated that exploitation  
29 rates within this bay area were high,  $>0.94$ . Changes in catch per unit of effort and catchability  
30 throughout the season explain how the “casita/campo” system allows for a seasonal  
31 replenishment of juveniles and adults, which has kept the landings relatively stable for the past  
32 decade.

33 **Keywords:** lobster fishery; *Panulirus argus*; tagging, marine protected areas, exploitation rates

## 34 **Introduction**

35 The Caribbean spiny lobster *Panulirus argus* (Latreille, 1804) is widely distributed from the  
36 southern USA to Brazil and throughout the Caribbean (Butler et al. 2011). This species is the  
37 most valuable resource fished within the Mexican Caribbean, but there is still a lack of  
38 knowledge regarding the basic mechanisms and processes that determine the dynamics of the  
39 local populations which are part of the Caribbean’s meta-population. Historically, spiny lobster  
40 fisheries have supported important commercial fisheries along the Caribbean, but increased  
41 fishing pressure has reduced lobster abundance, and currently most fisheries are being depleted  
42 (Ehrhardt et al. 2010).

43           This study was based in Bahía Espiritu Santo, which is located on the central coast of the  
44 State of Quintana Roo, in the Mexican Caribbean (Fig. 1). It is a shallow bay with an area of  
45 approximately 300 km<sup>2</sup> (Sosa-Cordero et al. 1999) with very similar habitat characteristics and  
46 oceanographic conditions to Bahía de la Ascensión (Lozano-Álvarez and Negrete-Soto 1991,  
47 Sosa-Cordero et al. 1998) which is to the north. Bahía de la Ascensión has been more  
48 thoroughly studied (see review by Briones-Fourzán et al. 2000, Sosa-Cordero et al. 2008). Both  
49 bays are within the Sian Ka'an Biosphere Reserve (SK), which is a protected area where the  
50 fishery is currently co-managed by federal government authorities such as the National  
51 Commission for Protected Areas (CONANP) and the National Commission for Fisheries and  
52 Aquaculture (CONAPESCA).

53           In Bahía Espiritu Santo, as in its neighbouring Bahía de la Ascensión (north), the  
54 community-based fishers are organized into cooperatives, and the lobster fishery is based on a  
55 'casita/campo' system. Casitas are large artificial shelters that can hold the full size range of  
56 lobsters. Casitas are made of cement, with a structural frame of approximately 1.8 x 1.2 m, raised  
57 about 10–15 cm above the bottom by two slabs along the longer sides. One or two shorter sides  
58 are fully open, allowing lobsters to freely enter and exit the shelter. Ownership of casitas within  
59 these bays is subject to an organizational scheme, which consists of partitioning the fishing areas  
60 into parcels ("campos") allotted to individual fishers, which vary in size from areas of 3 km<sup>2</sup> to  
61 20 km<sup>2</sup>. A fisher does not own the campo (as ownership of any sea areas is forbidden by law),  
62 but he is free to deploy casitas within it, and hence he owns the casitas and manages his campo in  
63 a semi-ownership arrangement (Lozano-Álvarez *et al.* 1991, Briones-Fourzán et al 2000, Defeo  
64 and Castilla 2005, Sosa-Cordero *et al.* 2008). The fishers fish their campos using small boats (7  
65 m long, 60 HP motor) and some GPS devices, and extract lobsters from casitas by skin diving,

66 exclusively using hand nets or snares (Lozano-Álvarez et al. 1989, INE-SEMARNAT 1996,  
67 Lozano-Álvarez et al. 1993, Sosa-Cordero et al. 1998). Skin diving limits fishing to depths  
68 shallower than 15-20 m, (Lozano-Alvarez et al. 1991a, b; Sosa-Cordero 2003, Ley-Cooper  
69 2006). Both the cooperatives licence and fishers individual permits are renewed annually, and  
70 cooperatives must comply with the federal fishing regulations: a closed season from March to  
71 June, a minimum capture size of 135 mm tail length (~74.5 mm carapace length, CL) (Lozano-  
72 Alvarez et al. 1991b) and a prohibition on the capture of ovigerous females, as well as specific  
73 park management rules, such as no-take zones. Internal regulations of the cooperatives forbid the  
74 use of SCUBA or hookah-diving, the use of lobster traps, and fishing in someone else's campo.  
75 Historically the semi-ownership of the campos has allowed self-surveillance and provided an  
76 efficient management arrangement for these fisheries.

77 Exclusive concession rights have been historically granted to three main fishing  
78 Cooperatives that have managed to maintain the total landings relatively stable in the last two  
79 decades (Sosa-Cordero et al. 2008). The age structure of caught lobsters showed a reduced adult  
80 mortality, since the catch consists mainly of sub-adult lobsters that inhabit the shallow fishing  
81 areas within the bays (Briones-Fourzán et al. 2000), whereas most of the adult lobsters inhabit  
82 deeper areas beyond the bays which are not fished (Lozano-Álvarez et al. 1993).

83 Despite the economic importance of the lobster fishery for the local communities, limited  
84 information exists on the lobster population dynamics and exploitation rates of lobsters in Bahía  
85 Espíritu Santo (Sosa-Cordero et al. 1999). In the present study, we implemented a mark-  
86 recapture programme to assess fishing and natural mortality of adults and sub-adult *P. argus*  
87 lobsters inhabiting this bay.

88

## 89 **Materials and Methods**

### 90 *Pilot study: Examining Lobster Tag loss*

91 *Panulirus argus* lobsters (> 74 mm CL) were collected at the study site of Bahía del Espiritu  
92 Santo, with hand nets. Lobsters were kept in sea cages for a few days and then transported to the  
93 Reef Systems Unit of the National Autonomous University of Mexico at Puerto Morelos. There,  
94 lobsters were distributed among three circular fibreglass tanks (3 m in diameter and 0.9 m in  
95 height), covered with a dark mesh netting for shading, and supplied with a continuous flow-  
96 through seawater system that maintained water temperature and salinity. Each tank contained  
97 one or two relatively small casitas to provide shelter for the lobsters.

98         Other than being fed frozen molluscs and shrimp every three days, the lobsters remained  
99 undisturbed for a 10-d acclimatization period. After this period, lobsters were measured (CL,  
100 from between the rostral horns to the posterior edge of the cephalotorax) using vernier callipers  
101 ( $\pm 0.1$  mm) and sexed based on dimorphic characters. Lobsters were tagged with plastic T-bar  
102 anchor tags of 50 mm in length (Hallprint, Australia) inserted into the tail extensor muscle  
103 between the posterior edge of the carapace and the first abdominal segment using a tag  
104 applicator. The applicator needle was sterilized in 100% ethanol before each subsequent  
105 tagging. Twenty-six and 32 lobsters were tagged in two separate tagging trials. The first trial  
106 lasted for 135 d, and the second trial lasted for 195 d. Throughout these periods, the tanks were  
107 examined daily for dead lobsters and loose tags. Every 2–3 weeks the lobsters were re-measured  
108 and examined daily for the presence and condition of their tags. Although these operations may

109 have disturbed lobsters and/or caused tag shedding, tags that were lost were usually recovered on  
110 dates not related with those in which these operations were performed.

111 The lobster tag-recapture model (see below) runs on a monthly timescale and requires a  
112 single average lobster tag-loss proportion in each time-step. Therefore, to obtain a monthly  
113 average tag loss, the proportion of tagged lobsters in the aquaria each day was averaged by 30-d  
114 periods (Fig. 2). The average monthly proportion of tagged lobsters was then described by a  
115 exponential decay equation for input into the tag-recapture model (Fig.2).

116 We did not differentiate between tag-shedding, natural mortality resulting in the tags to  
117 be lost, and tagged induced mortality because for the purpose of the model, the important factor  
118 was the proportion of lobsters that remained tagged after a specified time.

### 119 ***Brownie Model***

#### 120 *Lobster Tagging Protocol*

121 Lobsters were captured from “casitas” across the bay and lagoon areas of Bahía Espíritu Santo  
122 (1–7 m in depth) by using hand nets and carried back to the boat for tagging. Lobsters were  
123 tagged in the same manner as the tag-loss trials. The tag number, tagging date, location, sex and  
124 carapace length were recorded for each tagged lobster and then the lobster was immediately  
125 returned to the same casita. Tagging was conducted in June (during the closed season) and in  
126 July, September and October 2010 (during the fishing season). Only lobsters  $\geq 44.0$  mm CL were  
127 tagged to reduce incidental mortality (Lozano-Alvarez *et al.* 1991b), with the majority of tagged  
128 lobsters being  $>60$  mm CL.

#### 129 *Estimating the reporting rate*

130 The tagging program was advertised widely and included a reward system to encourage lobster  
131 tag returns. Throughout the July 2010 to February 2011 fishing season, two technicians were in  
132 frequent contact with fishers and administrators of the two cooperatives that fish at Bahía  
133 Espiritu Santo. Fishers were interviewed weekly to discuss the progress of the study and to  
134 prompt them to continue to report tagged lobsters. During these interviews, the fishers were  
135 asked whether they had caught but not reported tagged lobsters and, if so, to provide at least  
136 information on the tag numbers. Because of the high level of interaction between researchers  
137 and fishers the tag-reporting rate was considered to be 100%.

### 138 *Mortality estimates*

139 Instantaneous rates of fishing and other mortality for *P. argus* in Bahía Espiritu Santo were  
140 estimated using a Brownie model (Brownie et al. 1985) that was modified to allow for the  
141 incomplete mixing of tagged lobsters during the first period of recapture and to incorporate  
142 fishing effort data (Hoenig et al. 1998) (Table 1). These modified attributes of the Brownie  
143 model were important as they accounted for the behaviour of *P. argus* after tagging. Since  
144 previous tagging work in Bahia de Ascención (Lozano-Álvarez et al. 1991b) was unable to  
145 determine whether natural mortality or emigration were the cause of tag-loss, our model  
146 assumes that mortality other than that due to fishing is a combination of natural mortality and  
147 emigration.

148 Because tagged lobsters were returned to the same casita from which they were captured,  
149 initially all tagged lobsters were heterogeneously distributed throughout the bay. However  
150 previous studies showed that lobsters actively move between casitas (Briones-Fourzán et al.  
151 2000, Briones-Fourzán and Lozano-Álvarez 2001), and that one month after tagging the

152 distribution of tagged lobsters throughout the bay could be considered more homogeneous  
153 (Lozano-Álvarez et al. 2003). The re-parameterisation of the model accounts for this initial non-  
154 mixing of tagged and untagged lobsters by applying a specific non-mixed estimate of  
155 catchability ( $\hat{q}_j$ ) for all lobsters recaptured with a liberty time of less than two months.

156 In the modified model, other forms of mortality (eg. natural mortality, emigration) were  
157 assumed to remain constant over the entire fishing season, whereas catchability ( $q$ ) was allowed  
158 to vary between months. Estimates of the instantaneous rate of fishing mortality ( $F$ ) were  
159 determined as  $F = qE$ . Catch and effort data were collected from the fishing cooperatives  
160 landing logbooks and verified against data from tax declaration forms. All catches were  
161 converted to whole weights (kg) using a conversion factor based on the relationship between tail  
162 weight and whole weight previously reported by Lozano et al. (1991b). The unit of effort used  
163 was fishing trip adjusted to a monthly proportion of the entire seasons fishing effort.

164 The lobster tag-recovery data were analysed using a Brownie model (Brownie et al. 1985)  
165 as modified by (Hoenig et al. 1998) and described in this paper. The model was constructed in R  
166 (R Development Core Team 2010) with its Log-Likelihood maximised using a Nelder-Mead  
167 method in the optimum routine (Nash 1990, Nelder and Mead 1965)

168



169 **Results**

170 *Examining lobster tag loss and tag reporting rates*

171 In both trials, the rate of average monthly lobster tag loss for each trial was initially rapid before  
172 progressively lessening through time until the end of the experiments, 4.5 and 6.5 months later  
173 respectively (Fig. 2). The relationship between the number of months after release (time at  
174 liberty) and the proportion of lobsters retaining tags was better described by an exponential  
175 decay (Akaike's An Information Criterion (AIC) = -23.2,  $r^2 = 0.86$ ) rather than by a linear  
176 relationship (AIC = -18.9,  $r^2 = 0.8$ ) (Sakamoto et al. 1986). The exponential decay relationship  
177 describing the monthly proportion of tags remaining on lobsters was  $TL = 0.953 * e^{(-0.15 * L)}$ , where  
178  $TL$  is the proportion of lobsters still tagged and  $L$  is the number of months after release.

179 *Brownie model*

180 *Tag recaptures*

181 During the first two months of the fishing season, the fishers actively contacted research  
182 staff and submitted all requested information. However many became less interested from about  
183 the third month of the season, and had to be prompted to provide the data. The cooperatives  
184 retained all information and passed it onto researchers every month. Fishers had three  
185 opportunities to search for tags before handing lobsters to third parties: a) when catching them by  
186 free diving, b) when taking them from personal cages to where they are weighed, and c) when  
187 weighed at the cooperatives headquarters for sale.

188 In total, 786 lobsters were tagged throughout the Bay of Bahía Espíritu Santo, of which  
189 268 (34%) were recaptured, with most recaptures occurring at the beginning of the fishing

190 season from July through September (Table 2). Many lobsters that were recaptured in July and  
191 August had been released near the point of recapture (see Fig. 1), whereas many lobsters that  
192 were recaptured from October to December were found near the coral reef areas located east of  
193 the bay. These lobsters travelled distances ranging from 1.3 to 18.2 km within the bay. However,  
194 five lobsters that were originally tagged in Bahía de la Ascensión on April 2010 as part of a  
195 separate study (Lozano-Álvarez et al. unpublished data) were recaptured in Bahía Espíritu Santo.  
196 These lobsters travelled an average of 43.5 km in a southward direction.

### 197 *Fishery Catch and Effort statistics*

198 In Bahía Espíritu Santo, the monthly catch (whole weight) and effort (fishing trips) showed  
199 slightly different trends over the course of the fishing season (Fig.3). Effort decreased  
200 progressively from its maximum value in the first month of the fishing season (July 2010) to its  
201 minimum level by December 2010 and then remained relatively constant through to the end of  
202 the fishing season (February 2011) (Fig.3). Catch declined after the first month, and then  
203 increased over the next three months before decreasing again to stabilize at a low level (Fig.3).  
204 The catch per unit of effort (CPUE) began in July near maximum levels at 50 kg/trip before  
205 decreasing to the lowest level of 28 kg/trip the following month (August) (Fig. 3). Catch rates  
206 subsequently improved over the next three months to a maximum catch rate of 52 kg/trip in  
207 November before declining slightly to between 35 and 46 kg/trip from December to February  
208 2011 (Fig. 3).

### 209 *Lobster Tag recaptures and estimates of Exploitation rate*

210 A 34% recapture rate allowed for a good estimation of the fishing parameters and criteria for  
211 outputs of the model. Tagged lobsters that were released in the four discrete pulses (June, July,

212 September and October) were recaptured in large proportions, i.e. 33, 39, 100 and 29 %,  
213 respectively. The Brownie model estimated the monthly levels of exploitation required to  
214 reproduce similar proportions of tag-recaptures and the numbers of tags never seen again (NSA)  
215 (i.e. never recaptured) (Table 2, Fig. 4). The numbers of NSA tags were relatively high for the  
216 first, second and fourth tag release pulses since large numbers of lobsters were released in these  
217 pulses and only a fraction of lobsters were recaptured (Table 2, Fig. 4).

218 A residual plot of the model fit to the observed data indicated that the model was able to  
219 closely reproduce the pattern of tagged lobster recaptures with the error in the model estimates  
220 being evenly distributed (unbiased) between the four tagging/release periods and nine recapture  
221 periods (Fig. 5).

222 Model estimates of relative catchability differed dramatically between those representing  
223 unmixed ( $\hat{q}$ ) and mixed catchability for the same time periods (eg. 0.99 and 6.47 (unmixed)  
224 versus 1.89 and 2.47 (mixed) in July through September, respectively). Estimates of relative  
225 catchability ( $q$ ) remained fairly constant between July 2010 and November 2010, with a tight fit  
226 of values and their standard errors (see inset in Fig. 6). Mean catchability ( $q$ ) increased in the last  
227 three months of the fishing season although the confidence intervals for December and January  
228 continued to overlap those of earlier months. Estimates of monthly exploitation rates showed a  
229 slightly different pattern than those of catchability. Exploitation rates ranged between 0.22 and  
230 0.35 with highly overlapping confidence limits in all months (Fig.6). The total exploitation rate  
231 (the cumulative sum of monthly exploitation) and average rate of other mortalities estimated by  
232 the model were 0.94 and 0.24.

233

234 **Discussion**

235 The difficulty in holding lobsters in captivity for a long period limited the number of lobsters that  
236 we used for estimating tag loss. In addition, the potential effects of the behaviour, moulting, and  
237 the size and sex of lobsters on tag shedding have not been explored. Although further studies  
238 taking into account these factors may provide more information on the lobsters shedding of tags  
239 and lobster tag-induced mortality (Ehrhardt 2008), the data on “tag loss” that we obtained during  
240 the tank trials lies within the range obtained in previous studies on other lobster species  
241 (Montgomery and Brett 1996, Dubula et al. 2005). Considering that the trend in tag-loss in both  
242 trials followed the same pattern after time-step 4 (as shown in Fig. 2), the equation for tag loss  
243 would be valid for the time-steps remaining after the experiment was suspended; therefore, the  
244 longer period for field recaptures would be unaffected in the model.

245 The monthly tag loss obtained from the tank trials indicated that over time there is an  
246 exponential-like loss of individuals from a tagged *P. argus* population. Other lobster species,  
247 patterns showing an abrupt initial “tag loss” in the first few weeks post-tagging have been  
248 attributed to tag-induced mortality (Dubula et al. 2005) and tag shedding (Montgomery and Brett  
249 1996). In our tank experiment, we observed that tag loss in lobsters was also associated with  
250 molting activity and occasionally with cannibalism of lobsters about to molt or recently molted.  
251 The latter behavior can reflect stress associated with conditions of captivity (Moriyasu et al.  
252 1995, Lozano Álvarez 1996). However, tag loss because of these factors could be lower in  
253 natural habitats, where lobsters about to molt may isolate themselves.

254 Non-linear mortality patterns among tagged lobsters that were initially greater and then  
255 comparable to the control (untagged lobsters) were reported by Montgomery and Brett (1996).

256 They concluded that tagging causes only a short-term increase in mortality but is not significant  
257 in the long term when compared to the effects of natural mortality. This precedent supports our  
258 case in which an exponential decrease in tag loss (as is shown in Fig. 2) is representative.

259 Having a continued presence of observers on the field, along with a reward-based, tagged  
260 lobster reporting scheme, should be comparable to high-reward tagging studies that have shown  
261 it is possible to determine an exploitation rate from the tagged lobster recovery rate (Hoenig et  
262 al. 1998). The increase in likelihood of fishers not reporting tags as the months advanced was not  
263 a concern in this study for the above reasons. The model under this scheme is hereby considered  
264 an effective mortality estimation tool.

265 We assumed that it would take approximately two months for newly tagged lobsters to  
266 disperse and fully mix with the rest of the population. Consequently, our non-mixing model was  
267 developed to account for these differences, because tagged lobsters were immediately placed  
268 under the same casita in which they were found, and considering that they exhibit a certain  
269 degree of shelter fidelity because of their foraging and gregarious behaviour (Briones-Fourzán et  
270 al. 2007).

271 The model estimated that all other sources of mortality had a rate of  $0.24 \pm 0.02$  (mean  $\pm$   
272 SE). This estimate represents all removals of lobsters by factors other than fishing, tag loss or  
273 tag induced mortality, such as natural mortality and emigration. Since none of the tagged  
274 lobsters were recaptured outside of the bay, and very few were recaptured in the outer reef area,  
275 it is likely our estimate is a measure of predominately natural mortality. This estimate is well  
276 within the range of other estimates form in *P. argus* in the Mexican Caribbean (Lozano-Álvarez  
277 et al. 1991b, Lozano-Álvarez 1994, Arceo et al. 1997, Arce et al. 2001, Sosa-Cordero 2003,

278 Sosa-Cordero et al. 2008) This natural mortality rate, in conjunction with a total annual  
279 exploitation rate of 0.94, imply a large amount of recruitment (growth and/or immigration) must  
280 occur within this bay to maintain catch rates throughout the fishing season. Thus suggesting that  
281 throughout the 2010-2011 fishing season, all legal-sized lobsters found within the bay and at  
282 depths <20 m (depth limit for free-diving) were fished. It remains to be determined whether  
283 lobsters found deeper than 20 m have an effect on the dynamics related to the yearly casita  
284 replenishment processes.

285         In Bahía de la Ascensión, the highest annual CPUE (kg tail weight per boat per day) in  
286 the fishery also occurs in July, at the onset of the fishing season, and declines sharply during the  
287 rest of the fishing season. This trend has been mainly attributed to the combined effects of  
288 natural and fishing mortality (Lozano-Alvarez et al. 1991b). According to Sosa-Cordero et al.  
289 (2008) fishing effort tracks the catch trajectory, an indication that fishers reduce their activities  
290 (and hence their costs) when the local lobster resource is scarce. That is, the annual CPUE  
291 reflects fishing efficiency rather than resource abundance. This notion may also explain why, in  
292 Bahía Espíritu Santo, despite a relatively constant monthly catch from August to November,  
293 CPUE values showed an increasing trend during this period (see Fig. 3). In particular, the high  
294 levels of CPUE during October and November may be associated with the onset of the “Nortes”  
295 (cold fronts arriving from the north), which may increase movements of lobsters into the bay  
296 (Lozano-Álvarez et al. 1993) and the potential for casita replenishment. This is also suggested by  
297 the recapturing in Bahía Espíritu Santo of five lobsters that were tagged outside of the bay  
298 (Lozano-Álvarez et al. unpublished data).

299         At Bahía Espíritu Santo, the regulations on fishing gear result in a limited access to the  
300 deep lobster stock (> 20 m), where reef areas and a high proportion of reproductive adults are

301 found (Lozano-Álvarez et al. 1993, Ley-Cooper 2006). Results from the non-mixing tagging  
302 Brownie models analysis presented here, imply that to sustain a 0.94 annual exploitation rate of  
303 the population subject to fishing, there must be a constant source for lobster replenishment into  
304 the casitas found within this bay.

305 A fishery with a management scheme based on casita/campo system may well be  
306 functioning as an artificial refuge area that increases abundance and population density from  
307 several sources, such as larval recruitment, enhanced juvenile survival and growth (Briones-  
308 Fourzán et al. 2000, Briones-Fourzán et al. 2007) and adult attraction of lobsters migrating  
309 inshore and south during the early winter and closed season (Herrnkind 1980).

310 A trend observed is that as the fishing season advances from July to February, the CPUE  
311 does not decrease substantially, yet total landings do so, while fishing mortality (F) remains  
312 relatively stable. These results suggest a continuous input of lobster biomass into the casitas that  
313 is clearly higher during the closed season, potentially explaining the peak in catch at the  
314 beginning of the fishing season (Lozano-Álvarez et al. 1991b, Lozano-Álvarez et al. 1993).  
315 However, as the season advances, lobster availability decreases as indicated by the increase of  
316 catchability of tagged lobsters (see Fig. 6). Periodic and yearly casita replenishment shows the  
317 need for assessing the proportion of the lobster stocks dwelling outside the bay (Lozano-Álvarez  
318 et al. 1989, Lozano-Álvarez et al. 1993, Sosa-Cordero et al. 1998, Ley-Cooper 2006).

319 An option for dealing with non-mixing models is to increase the number of recaptures  
320 and tagged lobster reporting rates or to use a model that assumes the period of non- mixing lasts  
321 for less than a year. We have provided a type of model intermediate between the Brownie-like  
322 models and exact time of recapture models (see Hearn 1986, Hearn et al. 1987) considered to

323 apportion the total recaptures from a cohort to sub-annual two month periods. We have  
324 determined an expression for the expected value for each cell of the recovery matrix and raised  
325 the expected value to the observed number in the cell, as suggested by Hoenig et al. (1998). In  
326 addition to increasing the precision of the estimates, this type of model enables one to estimate  
327 the fishing mortality in the first time step of the study and consider a solid statistical theory  
328 captured in the Hoenig et al. (1998) models. More definitive information on emigration and the  
329 tagged lobsters reporting rate would be very valuable.

### 330 **Acknowledgements**

331 This project was partially funded by Colectividad Razonatura A.C., MAR Fund, FANP-FMCN  
332 and a scholarship (No. 202251), provided by CONACYT to KLC. We thank Oscar Guzmán and  
333 Mara Ley for technical support in field work, and Fernando Negrete-Soto and Cecilia Barradas-  
334 Ortiz for their help in the tank experiments. Permits to study and collect lobsters were provided  
335 by CONANP and CONAPESCA. We greatly appreciate the field help from fishers from the  
336 cooperatives “Cozumel” and “José Ma. Azcorra”, represented by Eduardo Pérez and Ruben Hoil  
337 respectively.

### 338 **References**

339 Arce M, Clemetson M, González-Cano J, Marshaleck S, OBrian C, Puga R, Restrepo R,  
340 Richards G, Ríos-Lara V, Sosa-Cordero E. 2001. FAO Fish. Rep. 619. Part I: Lobster  
341 Assessment Reports. Región 2: Belize, Southwest Cuba, and México. in: P.A.H. M.,  
342 ed. Report on the FAO/DANIDA/CFRAMP WECAFC regional workshops on the  
343 assessment of the Caribbean spiny lobster *Panulirus argus* Belize City, Belize April  
344 21-May 2, 1997 and Mérida Yucatán, México June 1-12 1998. p 52-74



345 Arceo P, Arce M, Briones-Fourzán P, Lozano-Álvarez E, Salas S, Seijo JC, Sosa-Cordero, E.  
346 1997. La pesquería de Langosta *Panulirus argus* en la plataforma de Yucatán y  
347 Caribe mexicano. Análisis y diagnóstico de los Recursos Pesqueros Críticos del Golfo  
348 de México Universidad Autónoma de Campeche EPOMEX Serie Científica 7:101-  
349 126.

350 Briones-Fourzán P, Lozano-Álvarez E. 2001. Effects of artificial shelters (Casitas) on the  
351 abundance and biomass of juvenile spiny lobsters *Panulirus argus* in a habitat-limited  
352 tropical reef lagoon. Marine Ecology Progress Series 221:221–232.

353 Briones-Fourzán P, Lozano Álvarez E, Eggleston DB. 2000. The use of artificial shelters (casitas)  
354 in research and harvesting of Caribbean spiny lobsters in Mexico in: Phillips B.,  
355 Kittaka J., eds. Spiny Lobster: Fisheries and Culture 2nd Ed Fishing News book-  
356 Blackwell. Oxford, UK; p 420-446.

357 Briones-Fourzán P, Lozano-Álvarez E, Negrete-Soto F, Barradas-Ortíz C. 2007. Enhancement of  
358 juvenile Caribbean spiny lobsters: an evaluation of changes in multiple response  
359 variables with the addition of large artificial shelters. Oecologia 151:401–416

360 Brownie, C., Anderson, D.R., Burnham K.P. & Robson D.S. 1985. Statistical Inference From  
361 Band Recovery Data: a Handbook, 2<sup>nd</sup> Edition, Washington, DC: US Dept. of the  
362 Interior, Fish and Wildlife Service. (U.S. Fish and Wildlife Service Resource  
363 Publication no.156). 305 pages.

364 Butler MJ, Paris C, Goldstein J., Matsuda K Cowen R., 2011. Behavior constrains the dispersal  
365 of long-lived spiny lobster larvae. Marine Ecology Progress Series, 422: 223-237

366 Defeo O, Castilla JC. 2005. More than one bag for the world fisheries crisis and keys for co-  
367 management successes in selected artisanal Latin America shellfisheries. Reviews in  
368 Fish Biology and Fisheries. 15: 265-283.

369 Dubula O, Groeneveld JC, Santos J, van Zyl DL, Brouwer SL, van den Heever N, McCue SA.  
370 2005. Effects of tag-related injuries and timing of tagging on growth of rock lobster,  
371 *Jasus lalandii*. Fisheries Research. 74:1-10.

372 Eggleston DB, Lipcius RN, Miller DL, Cobá-Cetina L. 1990. Shelter scaling regulates survival  
373 of Caribbean spiny lobster, *Panulirus argus*. Marine Ecology Progress Series 62: 70–  
374 88.

375 Eggleston DB, Lipcius R N. 1992. Shelter selection by spiny lobster under variable predation  
376 risk, social condition, and shelter size. Ecology 73:992–1011.

377 Ehrhardt NM. 2008. Estimating growth of the Florida spiny lobster, *Panulirus argus*, from molt  
378 frequency and size increment data derived from tag and recapture experiments.  
379 Fisheries Research 93:332-337.

380 Ehrhardt NM, Puga R, Butler MJ. 2010. Large ecosystem dynamics and fishery management  
381 concepts: The Caribbean spiny lobster. *Panulirus argus*, fisheries in: Fanning L,  
382 Mahon R, McConney P. ed. Towards Marine Ecosystem-Based Management in the  
383 Wider Caribbean. Amsterdam University Press. p157-175.

384 Hearn WS. 1986. Mathematical methods for evaluating marine fisheries. PhD. dissertation,  
385 School of Zoology, University of New South Wales, Kensington, Australia. 195  
386 pages.

387 Hearn WS, Sandland RL, Hampton J. 1987. Robust Estimation of the natural mortality rate in a  
388 completed tagging experiment with variable fishing intensity. International Council  
389 for the Exploration of the Sea 43:107-117.

390 Herrnkind W. 1980. Spiny Lobsters: patterns of movements In: Cobb JS, BF Phillips (eds.). The  
391 Biology and Management of Lobsters. Physiology and Behavior, Academic Press  
392 NY. p 349-407.

393 Hoenig JM, Barrowman JN, Pollock KH, Brooks NE, Hearn WS, Polancheck T. 1998. Models  
394 for tagging data that allow for incomplete mixing of newly tagged animals. Canadian  
395 Journal of Fisheries and Aquatic Sciences. 55:1477-1483.

396 INE-SEMARNAT. 1996. Programa de Manejo de la Reserva de la Biosfera de Sian Ka'an,  
397 Normas de Uso de Los Recursos Naturales. 38 pages.

398 Ley-Cooper K. 2006. Evaluación de estrategias para la explotación óptima de la población de  
399 langosta *Panulirus argus* en la Reserva de la Biósfera de Banco Chinchorro Quintana  
400 Roo. Tesis de Maestría Instituto de Ciencias del Mar y Limnología UNAM. 121  
401 pages.

402 Lozano-Álvarez E. 1994. Análisis del Estado de la Pesquería de la Langosta *Panulirus argus* en  
403 el Caribe Mexicano. In: Yáñez-Arancibia, ed. Recursos Faunísticos del Litoral de la  
404 Península de Yucatán: EPOMEX Universidad Autónoma de Campeche. Serie  
405 Científica 2: 43-55.

406 Lozano-Álvarez E. 1996. Ongrowing of juvenile spiny lobsters, *Panulirus argus* (Latreille,  
407 1804) (Decapoda, Palinuridae), in portable sea enclosures. Crustaceana 69: 958–973.

- 408 Lozano-Álvarez E, Briones-Fourzán P, Phillips BF. 1989. Spiny lobster in Bahía de la  
409 Ascension, Q.R. Mexico. Proceedings of the Mexico-Australia Workshop on Marine  
410 Sciences, CINVESTAV-Mérida, México: 379–391.
- 411 Lozano-Álvarez E, Briones-Fourzán P, Gonzalez-Cano J. 1991a. Pesca exploratoria de langostas  
412 con nasas en la plataforma continental del área de Puerto Morelos, Q.R., México.  
413 Anales del Instituto Ciencias del Mar y Limnología, Universidad Nacional  
414 Autónoma de México 18: 49-58.
- 415 Lozano-Álvarez E, Briones-Fourzán P, Phillips BF. 1991b. Fishery Characteristics, Growth and  
416 Movements of the Spiny Lobster *Panulirus argus* in Bahía de la Ascensión, México.  
417 Fishery Bulletin US 89:79:89.
- 418 Lozano-Álvarez E, Briones-Fourzán P, Negrete-Soto F. 1993. Ocurrence and Seasonal  
419 Variations of Spiny Lobsters, *Panulirus argus*, (Latreille) on the shelf outside Bahía  
420 de la Ascensión, México. Fishery Bulletin US 91:808-815.
- 421 Lozano-Álvarez E, Briones-Fourzán P, Ramos-Aguilar ME. 2003. Distribution, shelter fidelity,  
422 and movements of subadult spiny lobsters *Panulirus argus* in areas with artificial  
423 shelters (casitas) Journal of Shellfish Research 22:533-540.
- 424 Lozano-Álvarez E, Negrete-Soto F. 1991. Pesca exploratoria de la Langosta *Panulirus argus* con  
425 nasas frente a la Bahía de la Ascensión en el Caribe Mexicano. Revista de  
426 Investigaciones Marinas (Cuba) 12:261-268.
- 427 Montgomery SS, Brett PA. 1996. Tagging eastern rock lobsters *Jasus verreauxi*: effectiveness of  
428 several types of tags. Fisheries Research 27:141-152.

429 Moriyasu M, Landsburg W, Conan G. 1995. Sphyrion Tag Shedding and Tag Induced Mortality  
430 of the American Lobster, *Homarus americanus* H. Milne Edwards, 1837 (Decapoda,  
431 Nephropidae). *Crustaceana* 68:184-192.

432 Nash JC. 1990. Compact Numerical Methods for Computers. Linear Algebra and Function  
433 Minimisation. Adam Hilger. 278 pages.

434 Nelder JA, Mead R. 1965. A simplex algorithm for function minimization. *Computer Journal*  
435 7:308–313.  
436

437 Phillips BF. 2006. Lobsters Biology, Management, Aquaculture and Fisheries. Oxford, UK:  
438 Blackwell Publishing. 506 pages.

439 R Development Core Team. (2010). R: A language and environment for statistical computing. R  
440 Foundation for Statistical Computing, Vienna, Austria. ISBN 3-900051-07-0 , URL  
441 <http://www.R-project.org/>. Viewed on April 2011. Computer Program.

442 Sakamoto, Y., Ishiguro, M., and Kitagawa G. (1986). *Akaike Information Criterion Statistics*. D.  
443 Reidel Publishing Company. Hurvich CM & Tsai CL (1989). Regression and Time Series  
444 Model Selection in Small Samples. *Biometrika* 76, 297-307  
445

446 Sosa-Cordero E. 2003. Trends and Dynamics of the Spiny Lobster, *Panulirus argus*, resource in  
447 Banco Chinchorro, México. *Bulletin of Marine Science* 73: 203-217.

448 Sosa-Cordero E, Arce AM, Aguilar-Dávila W, Ramírez-González A. 1998. Artificial shelters for  
449 spiny lobsters *Panulirus argus* (Latreille): an evaluation of occupancy in different  
450 benthic habitats. *Journal of Experimental Marine Biology and Ecology* 229:1-18.

451 Sosa-Cordero E, Ramírez-González A, Dominguez-Viveros M. 1999. La Explotación de  
452 Langosta *Panulirus argus* en Bahía Espíritu Santo, Quintana Roo, México: Un Estudio

453 Descriptivo. Proceedings of the 45th Gulf and Caribbean Fisheries Institute  
454 Charleston 45:820-839.

455 Sosa-Cordero E, Liceaga-Correa MLA, Seijo JC. 2008. The Punta Allen lobster fishery: current  
456 status and recent trends. in: Townsend R, Shotton R, Uchida H (eds) Case studies on  
457 fisheries self-governance, FAO Fisheries Technical Paper. No. 514, Rome FAO  
458 451pp. pp. 149-162.

459

460 **Figure legends**

461 Fig.1. Map of Bahía Espíritu Santo South of “Sian Ka’an” Biosphere Reserve Mexican  
462 Caribbean. The locations and numbers of lobsters tagged are represented by the  
463 numbers in the white circles. The small map insert shows the location of the bay  
464 within the state of Quintana Roo

465 Fig.2. Proportion of lobster retaining tags over time derived from experimental aquaria tanks  
466 using T-bar Hall print tags in two independent trials.

467 Fig.3. Catch per unit of effort (CPUE -Kg/trip) (top figure), Catch (t) [black circles] (bottom figure) and  
468 Effort (trips) [grey squares] (bottom figure), in each month of the fishing season from July  
469 2010 to February 2011.

470 Fig.4. Observed (grey) and estimated (red) tag recoveries in each month of the fishing season  
471 from July to February with all non recovered tags represented on the right hand side  
472 of the plot (Never Seen Again: NSA).

473 Fig.5. Residuals from the Brownie model (observed - estimated).

474 Fig.6. Exploitation rate  $\pm$  1 standard error and relative catchability  $\pm$ 1 standard error (small  
475 insert) in each month of the fishing season from July 2010 to February 2011.

476

477 **Tables**

478 Table 1 Expected number of tag recoveries when additional lobsters are tagged and added to the  
 479 population at the start of each fishing month. Symbols are as follows:  $N_j$  number of  
 480 tagged lobsters released in month  $j$ ;  $\Phi$  probability of retaining a tag;  $\lambda$  tag reporting  
 481 rate;  $q_j$  catchability of the lobsters in month  $j$ ;  $\hat{q}_j$  incomplete mixed catchability of the  
 482 lobsters in month  $j$ ;  $E_j$  proportion of seasons total effort in month  $j$ ;  $M$  instantaneous  
 483 rate of natural mortality.

---

Expected recoveries in each month

Month	1	2	3
1	$\frac{N_1 \phi \lambda \hat{q}_1 E_1}{\hat{q}_1 E_1 + M} (1 - e^{-\hat{q}_1 E_1 - M})$	$\frac{N_1 \phi \lambda q_2 E_2}{q_2 E_2 + M} (1 - e^{-q_2 E_2 - M}) e^{-\hat{q}_1 E_1 - M}$	$\frac{N_1 \phi \lambda q_3 E_3}{q_3 E_3 + M} (1 - e^{-q_3 E_3 - M}) e^{-\hat{q}_1 E_1 - q_2 E_2 - 2M}$
2		$\frac{N_2 \phi \lambda \hat{q}_2 E_2}{\hat{q}_2 E_2 + M} (1 - e^{-\hat{q}_2 E_2 - M})$	$\frac{N_2 \phi \lambda q_3 E_3}{q_3 E_3 + M} (1 - e^{-q_3 E_3 - M}) e^{-\hat{q}_2 E_2 - M}$
3			$\frac{N_3 \phi \lambda \hat{q}_3 E_3}{\hat{q}_3 E_3 + M} (1 - e^{-\hat{q}_3 E_3 - M})$

---

484

485



486 Table 2 Data represented in the input form for the Brownie model with rows and columns  
487 representing tag release and recapture respectively. NSA represents the tags Never Seen Again.

Tag Recaptures by Month											
Month	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Jan	Feb	Released	NSA
Jun	0	105	35	29	9	2	5	0	0	568	383
Jul		27	15	14	5	3	1	1	2	174	106
Aug			0	0	0	0	0	0	0	0	0
Sep				2	1	0	0	0	0	3	0
Oct					0	8	2	2	0	41	29

488