

1 **Nitrogen-doped bamboo-like carbon nanotubes with Ni encapsulation**
2 **for persulfate activation to remove emerging contaminants with**
3 **excellent catalytic stability**

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18
19 **ABSTRACT**

20 Nitrogen-doped bamboo-like carbon nanotubes encapsulated with nickel nanoparticles (Ni@NCNTs)
21 were feasibly fabricated by a one-pot pyrolysis route. The characterization techniques revealed the
22 surface morphology, structure and chemical composition of the as-prepared catalysts synthesized in
23 different conditions. Ni@NCNTs were employed as both adsorbents and catalysts for activating
24 persulfate to remove an emerging pollutant, antibiotic sulfachloropyridazine. The nitrogen
25 modification enhanced both adsorption capacity and catalytic ability of Ni@NCNTs and proved to
26 be a promising alternative to metal oxides and other carbon catalysts. Moreover, Ni@NCNTs showed
27 an excellent stability because of nitrogen heteroatom doping and Ni nanoparticle encapsulation.
28 Electron paramagnetic resonance (EPR) and classical radical quenching tests discovered a
29 simultaneous radical/non-radical mechanism in Ni@NCNTs/persulfate system. The results indicated
30 that the novel nano-architected catalyst materials can show emerging functionalities for
31 environmental remediation processes.

32 **Keywords:** Nitrogen doping, carbon nanotubes, antibiotics, persulfate, sulfate radicals

33 **1. INTRODUCTION**

34 Worldwide drinking water scarcity has become more intense because of the wide water
35 contaminations, which appears as one of the most critical challenges in the world. Many water
36 treatment techniques have been employed to cope with the rapid deterioration of water pollution in
37 terms of the discharge quantity and ever-increasing complexity [1]. Recently, emerging contaminants,
38 for instance pharmaceuticals and personal care products (PPCPs), have attracted tremendous attention.
39 Their intrinsic properties, the continuous and fast development of new species, ubiquity in all surface
40 waters, difficulty of complete remediation and tardy action in early days, as well as lack of
41 understanding of their possible jeopardization to the environment, have brought out great issues to
42 the development of remediation technologies [2, 3].

43 As a representative and main contributor of PPCPs, sulfonamide antibiotics have been intensively
44 used in animal husbandry and human medicine owing to their low cost and good therapeutic effect
45 [4-8]. As a pharmaceutical product, sulfonamides are extremely recalcitrant to complete natural
46 biodegradation because they are biological active [9]. Many studies have detected various
47 sulfonamide compounds in aqueous environment. Even at a low level, a continuous and long-term
48 exposure to such chemicals could lead to appreciable adverse effects on human health [10]. Therefore,
49 the complexity, fine water solubility and trace amount have made them difficult to reach complete
50 degradation via traditional wastewater treatment measures in wastewater treatment plants (WWTP)
51 [11, 12]. Recently, advanced oxidation processes (AOPs) have attracted massive attention as they are
52 very effective for the degradation and mineralization of refractory organic pollutants in the aquatic
53 environment [13]. Typically, AOPs are involving oxidants and superoxides such as ozone, oxygen,
54 hydroperoxide, and various radicals that are capable of destroying target organics to harmless species.
55 Hydroxyl radicals ($\cdot\text{OH}$) have been most consequently observed in AOPs such as Fenton reaction,
56 UV oxidation and ozonation owing to their superior oxidation ability towards organic pollutants with

57 almost no selectivity [14, 15]. Yet, the drawbacks of many $\cdot\text{OH}$ -related Fenton-like reactions, such
58 as excess sludge production, narrow pH (~ 3) requirement and metal leaching, are the bottleneck limits.
59 Comparably, strong oxidizing sulfate radical ($\text{SO}_4^{\cdot-}$) offers an excellent alternative to hydroxyl
60 radical ($\cdot\text{OH}$) because $\text{SO}_4^{\cdot-}$ has a higher oxidation potential (2.5–3.1 V) than hydroxyl radicals (2.7
61 V) and a better flexibility of pH tolerance [15-17]. Analogous to Fenton-like oxidation reactions,
62 sulfate radicals can be obtained by activating persulfate (PS) through various activation methods,
63 such as UV-light irradiation, heating, carbon catalysis, metal ion and metallic oxide catalysis, for
64 organic contaminant degradation [18-21]. Among the catalysts used in AOPs, carbon-based
65 nanomaterials stand out owing to the exceptional adsorption capacity, catalysis ability and
66 environmentally friendly nature [22]. Graphene oxide, reduced graphene oxide, carbon nanotubes,
67 and nanodiamonds have been employed in water treatment as an excellent alternative to metal-based
68 catalysts and shown great competitiveness in terms of avoiding metal toxics and associated
69 contamination [23]. However, pristine carbocatalysts suffer from poor stability in heterogeneous
70 reactions, and the heteroatom doping with N, S, P or B into carbon honeycomb network offers an
71 effective strategy to tackle such an issue [24-32].

72 Herein, we demonstrate a one-pot facile synthesis of high quality N-doped carbon nanotubes with
73 nickel nanoparticle encapsulation (Ni@NCNTs) via direct pyrolysis of melamine as carbon/nitrogen
74 co-precursor and nickel chloride as a Ni source. The prepared product presented a high N-doping
75 level (6.6%) and successful Ni encapsulation. In addition, due to the synergistic effect of N-doping
76 and Ni decoration, the catalytic oxidation and reusability were dramatically enhanced comparing with
77 CNT and N-CNT in activation of PS to effectively remove sulfachloropyridazine (SCP) in water.
78 Electron paramagnetic resonance (EPR) and quenching tests were used to investigate the mechanism
79 of PS activation and SCP oxidation.

80 **2. EXPERIMENTAL**

81 **2.1 Chemicals and materials**

82 Nickel chloride hexahydrate ($\text{NiCl}_2 \cdot 6\text{H}_2\text{O}$, 98%), phenol (99%), sulfachloropyridazine (99%), 5, 5-
83 dimethyl-1-pyrroline (DMPO), Potassium persulfate, and hydrochloride acid (32%) were obtained
84 from Sigma-Aldrich. Multi-walled carbon nanotubes (MWCNTs) were purchased from Timesnano,
85 China. Graphene oxide (GO) was prepared via a modified Hummers' approach and reduced graphene
86 oxide (rGO) was fabricated by annealing the obtained GO in a tubular furnace under nitrogen
87 atmosphere at 700 °C for 1 h. Nitrogen doped reduced graphene oxide (N-rGO) was prepared through
88 a hydrothermal approach, as detailed in previous studies [32-34]. All the chemicals were of analytical
89 grade or higher and used as received without further purification. Deionized (DI) water was used in
90 all the experiments.

91 **2.2 Synthesis of nitrogen-doped carbon nanotubes with nickel encapsulation (Ni@NCNTs)**

92 Nitrogen-doped carbon nanotubes with nickel encapsulation were fabricated by a controlled pyrolysis
93 process using melamine and nickel chloride hexahydrate as carbon/nitrogen and nickel precursors,
94 respectively. In a typical synthesis, melamine ($\text{C}_3\text{N}_6\text{H}_6$, 10 g) were dissolved in 20 mL of DI water at
95 65 °C under continuous stirring for 1 h until melamine was dissolved completely. $\text{NiCl}_2 \cdot 6\text{H}_2\text{O}$ as a
96 nickel precursor with designed loading (1, 2, 3.3, or 7 g) was introduced to the melamine solution to
97 form a homogenous greenish solution. Later, the mixed solution was heated and kept at 80 °C
98 overnight to completely remove water. The resulting precipitates were placed inside a tube furnace
99 and allowed nitrogen gas to flow through for three hours to purge out the air residue. And then the
100 samples were annealed at 700, 800 or 900 °C for 6 h under N_2 atmosphere with a heating rate of 5 °C
101 min^{-1} . After the pyrolysis, the sample was naturally cooled down to room temperature. Finally, the
102 resultant black product was ground to fine powders, followed by thoroughly washing with excessive
103 concentrated hydrochloride acid (32%), by stirring in the acid solution for 6 h. The obtained
104 precipitates were then washed with ethanol and DI water for a few times until the clear solution pH
105 achieved at the value of 7. The as-prepared catalysts were denoted as Ni@NCNT-T-M (where T
106 represents the calcination temperature and M represents the initial weight ratio of melamine to nickel

107 chloride), for instance Ni@NCNT-700-3 means Ni@NCNTs with 3:1 melamine to NiCl₂•6H₂O and
108 heated at 700 °C.

109 **2.3 Catalyst characterization**

110 The crystal structure of the samples determined by X-ray powder diffraction patterns were recorded
111 on a Bruker D8 Advance X-ray diffractometer using Cu K α radiation ($\lambda = 1.5406 \text{ \AA}$) at 40 mA and
112 40 kV with a scanning speed of $0.2^\circ (2\theta) \text{ min}^{-1}$. The Raman spectra were obtained using a laser with
113 an excitation wavelength of 785 nm at room temperature on a Renishaw Raman microscope. The
114 surface contents and types of carbon, nitrogen and metals were measured by X-ray photoelectron
115 spectroscopy (XPS) on a VG Multi lab 2000 spectrometer (Kratos AXIS Ultra DLD) with Al K α
116 radiation as the X-ray source (300 W). The C 1s line at 284.6 eV was applied as a calibration to correct
117 the binding energies. By deducting the Shirley-type background, the core level spectra were
118 deconvoluted into their components with Gaussian-Lorentzian (20:80) shape lines using the CasaXPS
119 software. The Brunauer–Emmett–Teller (BET) specific surface area and pore-size distribution of the
120 samples were determined by N₂ adsorption–desorption at -196°C using a Micrometrics Tristar 3000.
121 The samples were degassed in a vacuum at 110°C overnight before the tests. The morphologies of
122 the carbon samples were characterized using scanning electron microscopy (SEM; FEI Verios XHR
123 460) and transmission electron microscopy (TEM; JEOL 2100). The thermal stability was collected
124 by the thermogravimetric analysis (TGA) using a TGA/DSC1 STAR^e system, Mettler-Toledo thermal
125 analyzer under flowing air. The reactive radicals were detected by electron paramagnetic resonance
126 (EPR) on a Bruker EMS-plus instrument with DMPO as a spin-trapping agent, and the Xeon software
127 (Bruker) was used to analyze the quantitative information. The Ni concentration in reaction solutions
128 was evaluated by an Inductively Coupled Plasma Optical Emission Spectrometer (ICP-OES).

129 **2.4 Adsorption and catalytic oxidation of SCP solutions**

130 SCP adsorption was performed in a 500 mL conical flask at 25°C , unless other temperature
131 specifically mentioned. Typical adsorption experiments were carried out by dispersing the

132 Ni@NCNTs samples (0.2 g L^{-1}) in SCP (20 mg L^{-1} , $\text{pH} = 7$) solutions. At determined intervals, 1
133 mL of solution was withdrawn and immediately filtered by a $0.22 \text{ }\mu\text{m}$ Millipore syringe filter to
134 remove the solid particles.

135 **2.5 Catalytic oxidation of SCP solutions**

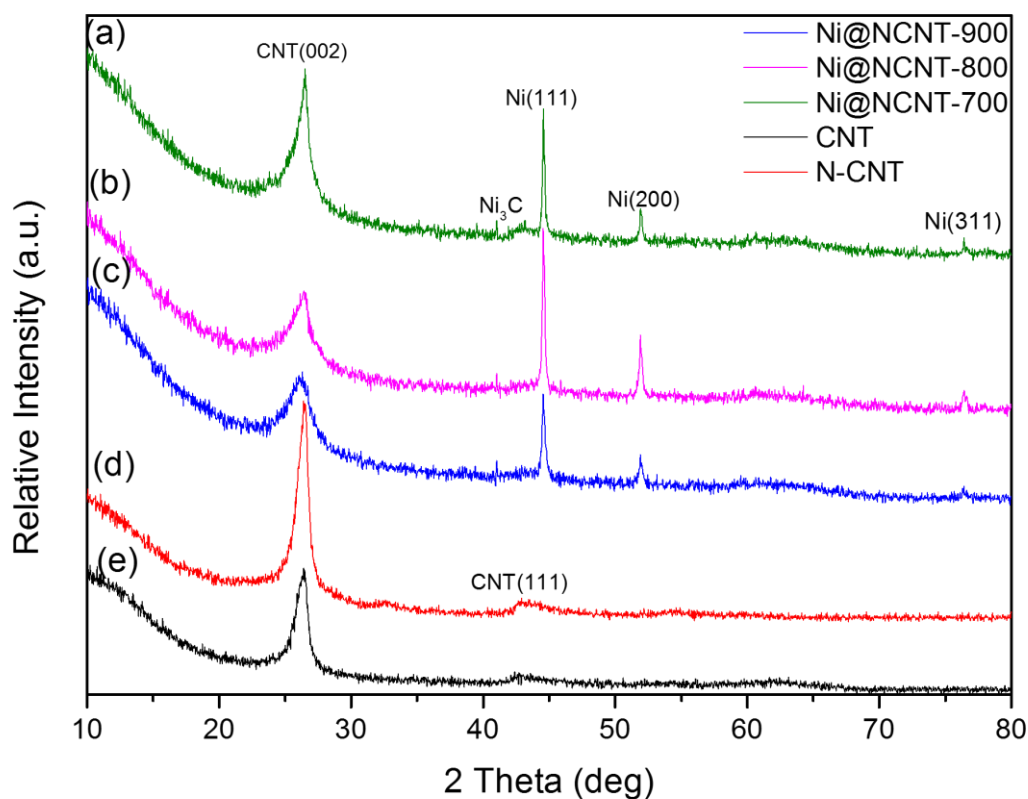
136 The degradation experiments were conducted by introducing designed amount of PS into the reaction
137 solutions when the adsorption/desorption equilibrium was reached. After the reaction was initiated,
138 1 mL of solution was extracted at planned time intervals, and then injected into an UHPLC vial mixed
139 with 0.5 mL methanol as the quenching agent. Each experiment and analysis was repeated twice and
140 the results were reproducible with relative errors less than $\pm 5\%$. The concentrations of SCP were
141 analyzed with an UHPLC system (Thermo-Fisher Scientific 3000) using an UV detector at a detection
142 wavelength of 270 nm. An Acclaim RSLC C-18 column ($2.2 \text{ }\mu\text{m}$) was used as a separation column.
143 The mobile phase was a mixture of diluted acetic acid and methanol (70:30, v/v) with a flow rate of
144 0.3 mL min^{-1} and $20 \text{ }\mu\text{L}$ of injection volume [32]. Total organic carbon (TOC) was measured using a
145 Shimadzu TOC-5000 CE analyzer for a few selected samples. For TOC measurement, 10 mL sample
146 was withdrawn at a designed interval, filtered and quenched with 10 mL of 0.3 M sodium nitrite
147 solution, and then the mixed solution was analyzed [35]. Data reported in the figures are mean values
148 of at least three experiments with error bars representing the standard deviation.

149 **3. Results and discussion**

150 **3.1 Characterization of carbon nanotubes materials**

151 X-ray diffraction (XRD) was employed to investigate the crystalline structure of the prepared CNTs.
152 **Fig. 1** displays the XRD spectra of CNT, N-CNT and Ni@NCNTs composites synthesized at various
153 pyrolysis temperatures. All the carbon nanotubes showed a broad reflection peak at $2\theta = 26.5^\circ$,
154 corresponding to the (002) plane of graphitic carbon with an interlayer spacing of 0.34 nm. The peak
155 intensity increased when temperature elevated from 700 to 800 °C and then decreased at 900 °C.

156 Additionally, the XRD spectra of Ni@NCNTs of different nitrogen loading amounts are shown in
157 **Fig. S1**. It is noticed that the broad background band centred at $2\theta = 26^\circ$ might be owing to the
158 successful incorporated nitrogen defects in the graphitic structure [36]. Aside from the graphitic
159 carbon peak, three pronounced peaks can be observed at $2\theta = 44.5, 51.9$ and 76.4° , corresponding to
160 the (111), in comparison with CNT and N-CNT, (200) and (311) reflections of Ni nanoparticles,
161 respectively and confirming the incorporation of Ni nanoparticles in the CNT catalysts [37]. A less
162 conspicuous peak appearing at 41° indicated the existence of nickel carbide (Ni_3C) in the CNTs. It is
163 worth to notice that there is no obvious nickel oxide peak detected, typically at around $2\theta = 37^\circ$, which
164 might be attributed to the reduction of nickel ions to nickel nanoparticles by carbon atoms. It was
165 found that nickel ions could be reduced to nickel metal nanoparticles by using amorphous carbon as
166 a reductant in argon atmosphere at 500°C and that nickel oxide particles could be reduced by natural
167 graphite at a high pyrolysis temperature [38] [39]. Moreover, the CNT compartment layers may also
168 provide the embedded nickel particles a protection shell from being oxidized.



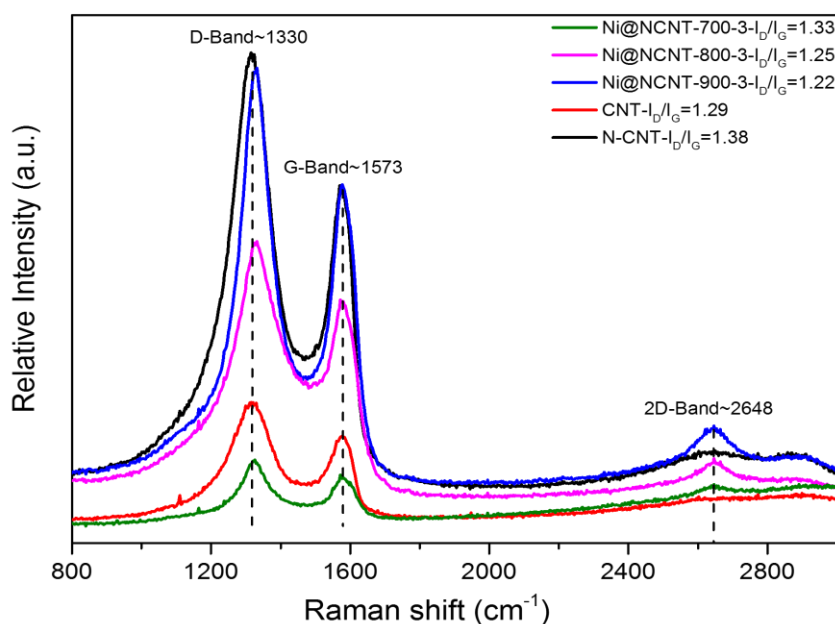
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170 **Fig. 1** XRD patterns of (a) Ni@NCNT-900-3, (b) Ni@NCNT-800-3, (c) Ni@NCNT-700-3, (d) N-
 171 CNT, (e) modified CNT samples.

172 Raman spectroscopy as one of the most sensitive and informative techniques is widely used to
 173 evaluate graphitic materials for graphitization degree, graphene layer number or carbon nanotube wall
 174 layers, nanotube diameter, disorder level in sp^2 network and doping grade. **Fig. 2** shows the Raman
 175 spectra of Ni@NCNTs at various pyrolysis temperatures, e.g. 700, 800 and 900 °C, CNT and N-CNT.
 176 Three prominent peaks are observed at 1330, 1573 and 2648 cm^{-1} , which are attributed to the D, G
 177 and 2D bands, respectively.

178 The intensity ratio of D band to G band (I_D/I_G) is often used to evaluate the structural disorder of
 179 carbon materials. The I_D/I_G of N-CNT (1.38) showed higher value than modified CNT (1.29),
 180 indicating that nitrogen doping on the sp^2 carbon network could largely boost the defect density. The
 181 I_D/I_G ratio of Ni@NCNTs decreases from 1.33 to 1.22 with increasing pyrolysis temperature from

182 700 to 900 °C, suggesting that the carbon nanotubes possess less lattice defects and disorders derived
 183 from high pyrolysis temperature. The disorder property of the catalysts will be further confirmed by
 184 electron microscopy. The G band peak width is another indicator of heteroatom doping [40] [41].
 185 **Fig.2** shows that G band width of CNT synthesized at various temperatures came along with an order
 186 of Ni@NCNT-700 > Ni@NCNT-800 > Ni@NCNT-900, confirming the successful N doping, and
 187 that the doping degree reduced with elevated pyrolysis temperature. **Fig. S2** also shows the variation
 188 of I_D/I_G ratio along with the nitrogen precursor weight percentage, indicating that a higher nitrogen
 189 precursor content would give rise to a higher N doping level. However, when the nitrogen and nickel
 190 precursors loading ratio was raised to 7, the I_D/I_G ratio dropped, suggesting that higher nitrogen
 191 precursor amount would not always give a higher disorder level.



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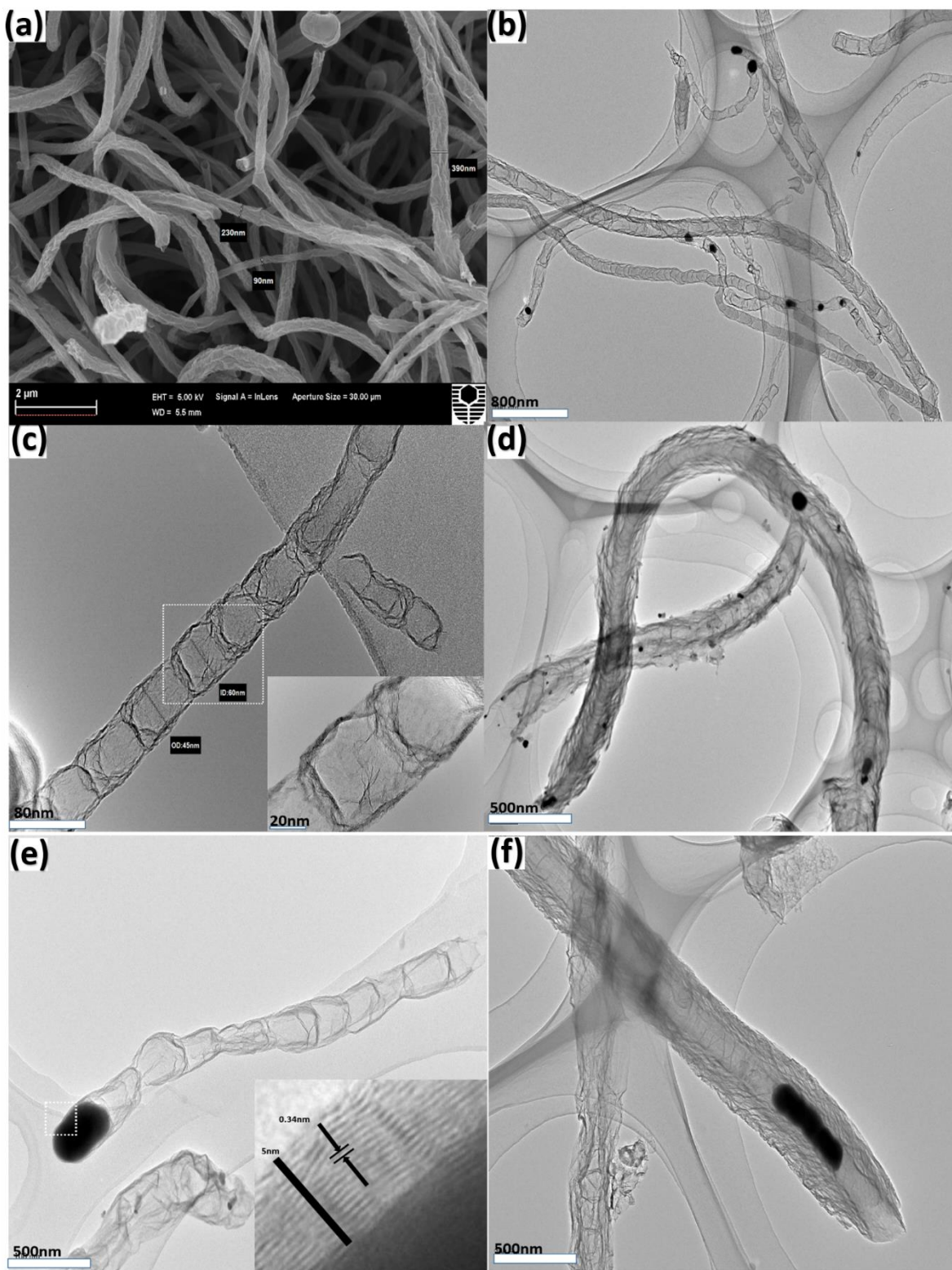
193 **Fig. 2** Raman spectra of Ni@NCNTs at different annealing temperatures, CNT and N-CNT.

194 The surface morphology of the as-synthesized Ni@NCNTs nanostructures was characterized by SEM
 195 and TEM imaging. The SEM images in **Fig. 3(a)** reveal that the product fabricated at 800 °C contains
 196 well grown nanotubes with a herringbone structure and is free from impurities such as amorphous
 197 carbon and polyhedral carbon particles. The diameter of the tubes ranges from 70 to 400 nm, and

198 length can be up to dozens of micrometres. TEM images in **Figs. 3(b)** and (c) further confirm that
199 they are hollow CNTs with a typical bamboo-like and compartmentalized structure. It was also shown
200 that the pea-pod like nickel particles are encapsulated inside the carbon tubes. The inner tube diameter
201 and the wall thickness of a typical carbon nanotube were measured to be 45 and 20 nm, respectively.
202 In addition, it was observed that most of the nickel nanoparticles were kept inside the CNTs after
203 being washed in concentrated hydrochloride acid solution for hours, indicating that the nickel
204 nanoparticles are quite inaccessible from being affected by outside environment, acid solution and
205 oxidation. To further confirm such a point, **Fig. 3(d)** shows the morphology of CNTs before acid
206 treatment. The nickel particles (black dots) are randomly dispersed over the CNTs both inside and
207 outside. On the other hand, the acid treated tubes in **Fig. 3(b)** show a nice and clean tube surface.
208 Many metal-based catalysts have been extensively applied for AOPs, yet the metal leaching is surely
209 a barrier for their further development [29]. In this study, the Ni content of Ni@NCNTs/PS reaction
210 solutions was tested by ICP-OES and it was shown that the Ni concentration was as low as 0.47 mg
211 L⁻¹, which is within the limit of nickel range 0.02 – 1 mg L⁻¹ of the world drinking water quality [42].
212 In addition, the magnetic property of Ni@NCNTs before and after acid wash is shown in **Fig. S3**.
213 The raw Ni@NCNT showed strong magnetism, and the magnetism was largely reserved even after
214 acid wash, which is due to the carbon layers on Ni particles. Such magnetic property can be very
215 convenient for post catalyst recycling process.

216 In **Fig. 3(e)**, nickel particles are detected at the tip of the CNTs, which implies that a tip-growth
217 mechanism might be responsible for the formation of carbon nanotube. Besides, the motion trail of
218 nickel particles is illuminated in **Fig. 3(f)**, suggesting that the nickel particles are able to diffuse along
219 the tube axis and aggregate at the tube end to form a worm-like Ni catalyst lump. A magnified TEM
220 image of encapsulated nickel particles is shown in **Fig. 3(e) (inset)**. The nickel nanoparticles are
221 embedded in a graphitic shell composing of well-ordered carbon layers, and the inter plane spacing
222 of the fringe lattices is in a range from 0.34 to 0.38 nm, which increased with growing tube diameter
223 and agreed to the size effect's rule [43].

224 In order to systematically study the influences of the pyrolysis temperature and amount of nitrogen
225 precursor on CNTs morphology and microstructure, the SEM images of CNTs synthesized at
226 different conditions are displayed in **Fig.S4**. At 700 °C, the obtained sample represents a
227 chrysanthemum-like carbon nanotube cluster with irregular tube sizes and short length, also the
228 graphene like carbon material can be observed (**Figs. S4** and **S5**). As the SEM image was taken from
229 the overlook-angle, one can clearly see that many of the growing tubes are observed to be open ended
230 and interestingly showing a multi-walled carbon tube structure, which is consistent with the Raman
231 results.



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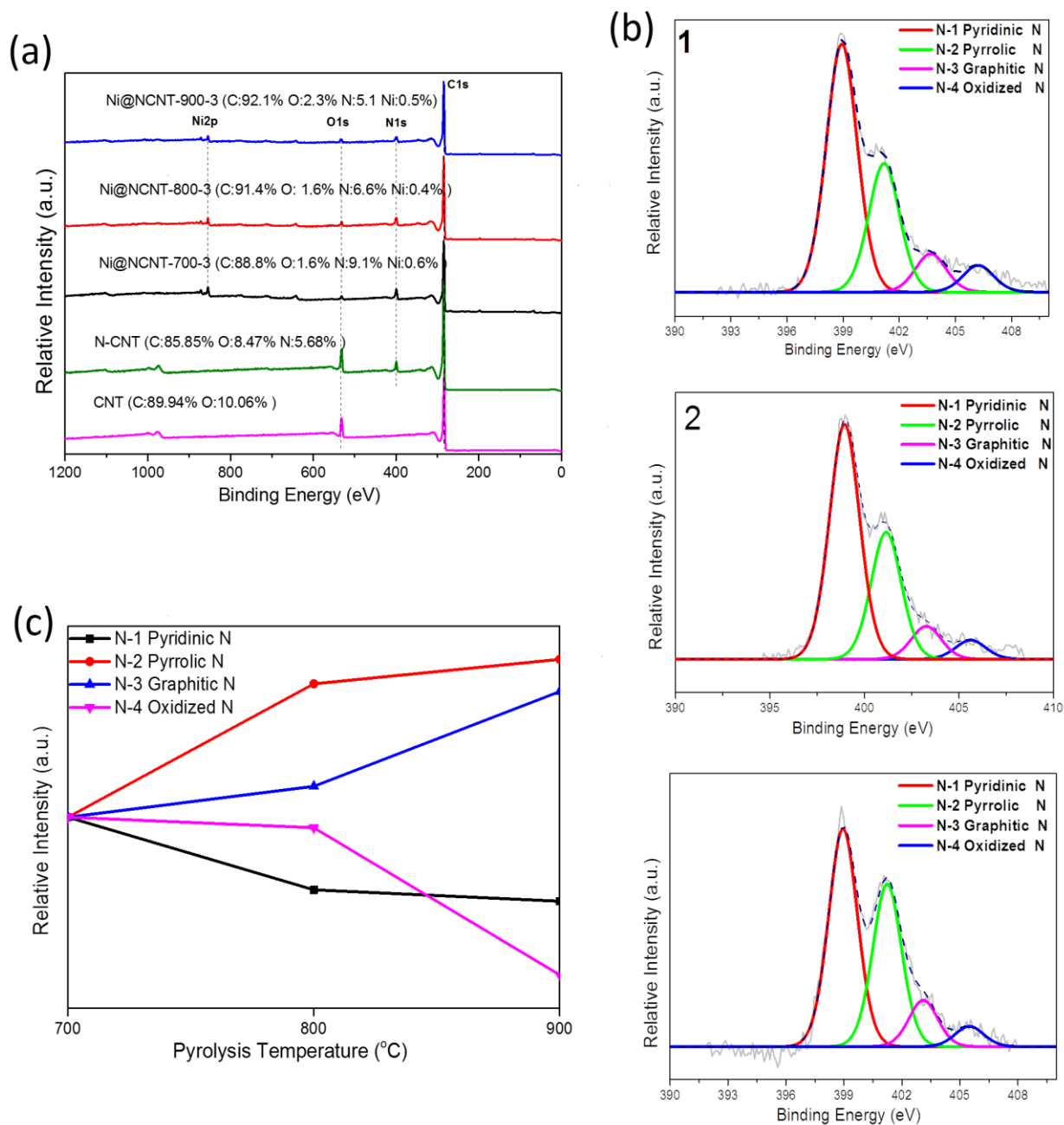
233 **Fig. 3** SEM and TEM characterization of the Ni encapsulated N-doped carbon nanotube. (a) SEM
 234 image of the Ni@NCNT-800-3. (b) TEM image of Ni@NCNT-800-3 with pea-pod like Ni particle
 235 encapsulation. (c) TEM images of bamboo like CNT structure with compartments; inset shows the
 236 magnified part of the compartments.

237 TEM image of Ni@NCNT-800-3 with tip Ni encapsulation; inset displays the graphitic layers of the
238 carbon shell. (f) TEM image of Ni@NCNT-800-3 with worm like Ni particle accumulation at the
239 tube end.

240 The chemical composition and state of the nitrogen doped carbon nanotubes were analyzed by XPS.
241 **Fig.4** displays the XPS survey showing C 1s, N 1s, O 1s and Ni 2p spectra. The emergence of N 1s
242 peak in all samples confirmed that nitrogen atoms were successfully doped into carbon catalysts.
243 Moreover, the XPS element analysis shows the N doping levels are 9.12%, 6.59% and 5.15% in
244 Ni@NCNTs synthesized at 700, 800 and 900 °C, respectively, which are superior to many other
245 reported approaches such as physical vapor deposition (PVD) of 3.5% and chemical vapor deposition
246 (CVD) of 4%. [31, 44] The high nitrogen doping level also indicated that melamine as an excellent
247 nitrogen precursor can efficiently incorporate nitrogen atoms into carbon nanotube network during
248 the CNTs formation process. Table S1 shows the N doping levels in Ni@NCNTs at different synthesis
249 conditions.

250 The chemical states of different types of N dopants were studied by high resolution of N1s spectrum
251 in a specific range (392-408 eV). **Fig. 4(b)** illustrates the N1s spectrum was deconvoluted into four
252 types of N peaks, 398.9, 401.1, 402.7 and 405.2 eV, corresponding to pyridinic N (N at the edges of
253 the graphene layers bonded to two carbon atoms), pyrrolic N (N in five-membered rings), graphitic
254 N or quaternary N (N in six-membered rings within the graphene layers) and oxidized N, respectively.
255 **Fig. 4(c)** shows that the graphitic N content gradually increased from 8.67 to 10.38% with elevated
256 temperature, 700 to 900 °C. This may be due to the better thermal stability of graphitic N and the loss
257 of unstable N at high temperature [45-49]. The oxygen and nickel contents in these samples are also
258 listed in Table S1. All the samples represented a very low oxygen content due to pyrolysing in the
259 inert gas atmosphere at a high temperature, which enables most of the oxygen containing groups to
260 react with carbon to form CO_x. The XPS C 1s spectra are shown in **Fig. S6**. The thermal stability of

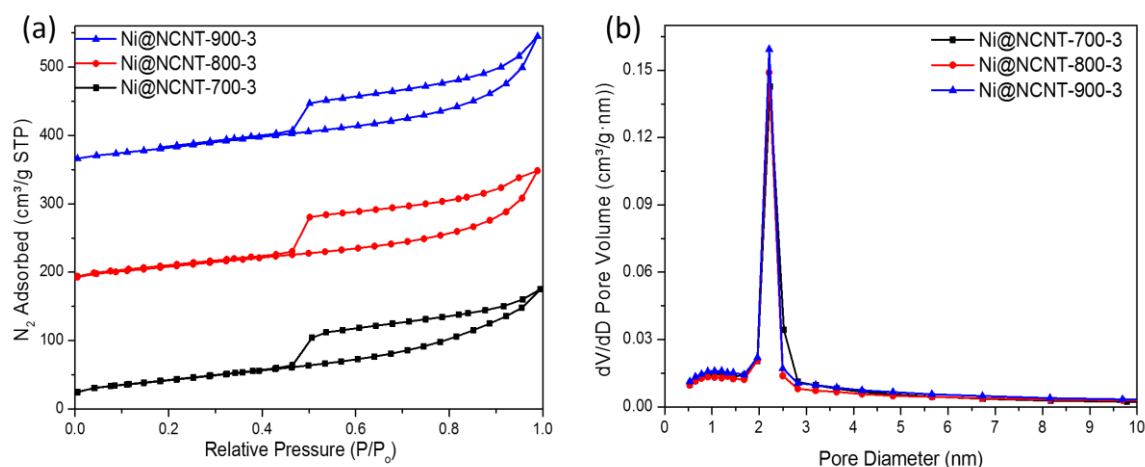
261 the typical Ni@NCNTs was investigated by TGA-DSC (**Fig. S7**). A total of 32% of the initial weight
 262 was remaining which can be nickel compounds such as NiO and Ni₂O₃, at Ni content of roughly 22%.



263

264 **Fig. 4** (a) XPS survey of Ni@NCNT-700, Ni@NCNT-800, Ni@NCNT-900, CNT and N-CNT. (b)
 265 N 1s scan of 1. Ni@NCNT-700, 2. Ni@NCNT-800 and 3. Ni@NCNT-900. (c) Density change of N
 266 components at different temperatures.

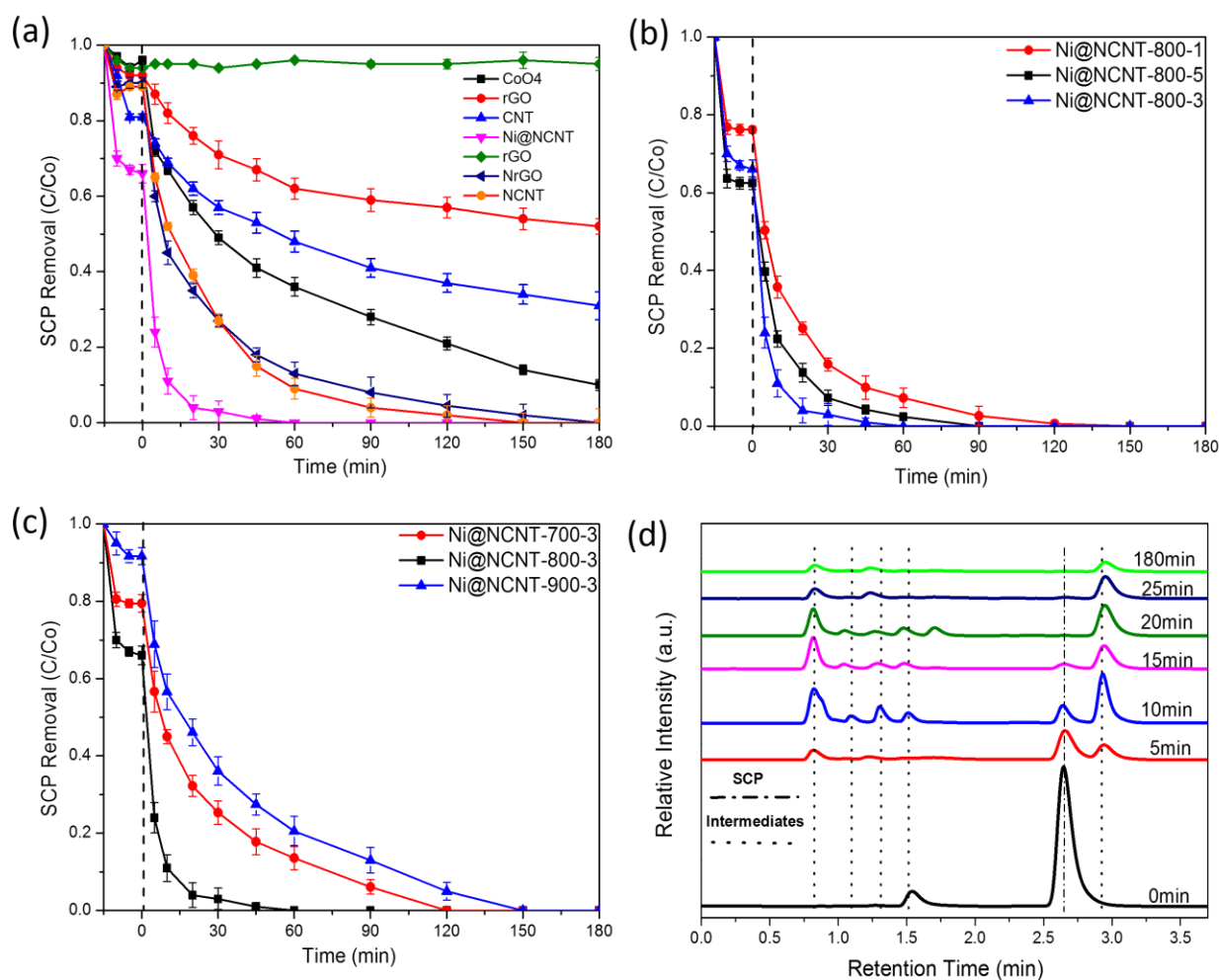
267 **Fig.5** shows the BET specific surface areas and pore structure of the as-synthesized samples at various
 268 fabrication temperatures. All the samples demonstrated a type IV isotherm with a H3 hysteresis loop
 269 that gradually increases at low pressure followed by a sharp jump at half way, illuminating the
 270 mesoporous structure [50]. The hysteresis loop of each sample in **Fig.5 (a)** shows quite a similarity,
 271 Ni@NCNT-800-3 was slightly broader than the other two, suggesting a higher N doping content
 272 would create more defect sites and lead to a more porous structure with a greater surface area and
 273 pore volume (Table S2). The catalysts fabricated at 700 °C had a smaller surface area (135 m²/g) than
 274 800 °C (155 m²/g). But at 900 °C the BET (145 m²/g) declined due to the collapse and aggregation of
 275 carbon nanotube. The pore size distribution (PSD) of samples is displayed in **Fig.5 (b)**, in which a
 276 sharp peak appears in all the three catalysts at 2.2, indicating that the Ni@NCNTs are mostly
 277 composed of micropores (< 2 nm) and mesopores (2-50 nm).



278

279 **Fig.5** (a) Nitrogen sorption isotherms and (b) pore size distributions of Ni@NCNTs at various
 280 synthesising temperatures.

281 3.2 Catalytic SCP removal performance of carbon nanotubes



282

283 **Fig. 6** (a) PS activation over various catalysts. (b) Effect of nitrogen/nickel precursor loading of
 284 Ni@NCNT-800 on SCP removal. (c) Effect of catalyst fabrication temperature on SCP removal. (d)
 285 HPLC spectra of SCP degradation in a 180 min experiment. (Initial SCP concentration = 20 mg/L,
 286 catalyst loading = 0.2 g/L, PS concentration = 2 g/L, and temperature at 25 °C.)

287 Comparative studies of various carbocatalysts and metal oxides were first carried out in catalytic
 288 activation of PS for SCP oxidation in aqueous solutions. Before the PS was introduced into the
 289 solution, a 15-min pre-adsorption was performed in order to distinguish the SCP removal achieved
 290 by adsorption. When PS was added individually, a limited amount of SCP removal (< 5%) was
 291 reached, suggesting that PS could barely oxidize SCP compounds without catalytic activation. Both
 292 rGO and N-rGO showed low adsorption of SCP, however, the catalysis of N-rGO was greater than
 293 rGO. As our previous studies demonstrated, N doping could remarkably improve the carbocatalytic

294 activity by enhancing the π - π bond of the carbon matrix through the lonely pair of electrons of doped
295 nitrogen atoms [27, 30, 32]. In this study, a similar case occurred. Comparing the catalysis on
296 modified commercial CNT, NCNT and as-prepared Ni@NCNT-800-3, MWCNT only reached less
297 than 65% of SCP removal and NCNT presented a higher activity than CNT, giving SCP removal at
298 100% in 3 h while Ni@NCNT-800-3 achieved 100% SCP oxidation efficiency within 1 h, indicating
299 that the N doping and Ni encapsulation induced catalytic enhancement. Nickel nanoparticles were
300 also tested for SCP degradation, but showing much less activity (10%). Moreover, Co_3O_4 as a typical
301 metal based catalyst was employed to compare with the as-prepared Ni@NCNTs on PS activation.
302 **Fig.6(a)** shows that Co_3O_4 presented a poor adsorption capacity and a moderate catalytic activity at
303 5% SCP adsorption and around 90% oxidation in 180 min, respectively, indicating that Ni@NCNTs
304 is more effective than the conventional metal oxide and other carbon catalysts for PS activation.

305 It is believed that a higher N doping level can lead to a more advanced catalysis [51-53]. **Fig. 6(b)**
306 shows the effect of the nitrogen doping level on SCP removal. As shown in Table S1, the nitrogen
307 contents for Ni@NCNT-800-1, Ni@NCNT-800-3 and Ni@NCNT-800-5 are 5.8, 6.6 and 7.7%,
308 respectively. The 15-min pre-adsorption tests show that SCP adsorption capacities are 0.08, 0.11 and
309 0.13 mol g^{-1} for Ni@NCNT-800-1, Ni@NCNT-800-3 and Ni@NCNT-800-5, respectively.
310 Interestingly, the catalytic ability, however, does not follow the adsorption capacity trend where the
311 reaction rate increases in an order of Ni@NCNT-800-3 > Ni@NCNT-800-5 > Ni@NCNT-800-1.

312 The catalysis of Ni@NCNTs was first improved when the nitrogen content increased from 5.8 to 6.6%
313 and more nitrogen functional groups on the CNT surface would work as active sites. However, when
314 the N doping reached its optimum level, a further increased N doping (7.7%) would cause destruction
315 of graphitic network structure and deterioration of the catalyst activity instead.

316 The influence of pyrolysis temperature at 700, 800 and 900 °C on catalytic oxidation was investigated
317 (**Fig. 6(c)**). The adsorption capacity and catalytic oxidation of SCP of the three counterparts are in an
318 order of Ni@NCNT-900-3 < Ni@NCNT-700-3 < Ni@NCNT-800-3, where Ni@NCNT-800-3

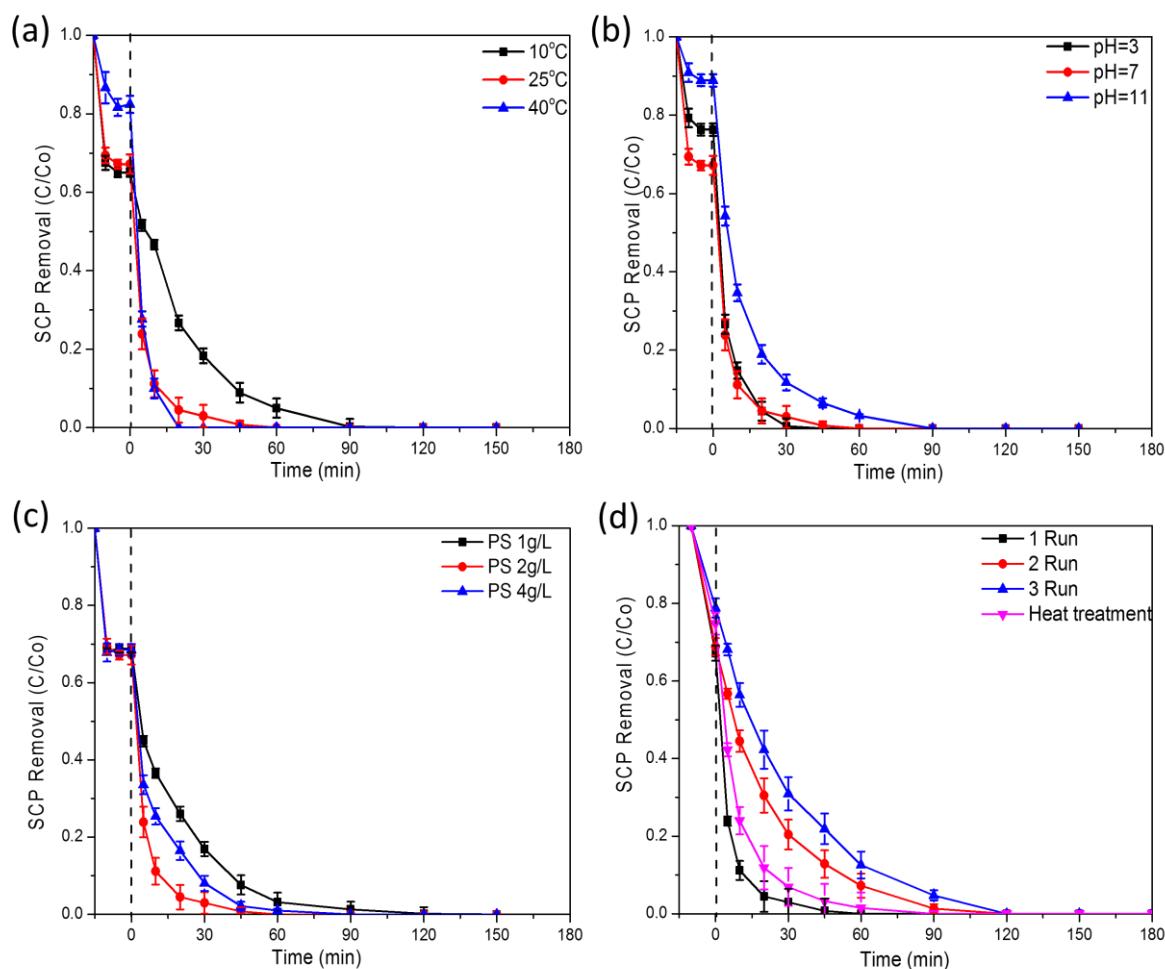
319 exhibited the best performance of complete decomposition of SCP in 60 min. This indicates that a
320 higher pyrolysis temperature did not necessarily endow greater catalysis and adsorption. Moreover,
321 SEM images showed that high annealing temperature can trigger the destruction and aggregation of
322 carbon nanotubes, which might be a main reason for the deactivation of Ni@NCNTs. The decay of
323 the SCP during the oxidation process was monitored by UHPLC as illustrated in **Fig. 6(d)**. It can be
324 seen that a well-defined SCP peak appeared at retention time (t_R) = 2.65 min and its peak intensity
325 gradually decreased while a number of by-products were emerged during the oxidation process. The
326 major intermediates include p-aminophenol (PAB), 3-amino-6-chloropyridazine (3A6C), 3-amino-4-
327 hydroxy-6-chloropyridazine (3A4H6C) and 4-amino-3-hydroxybenzenesulfonic acid (4A3H). As the
328 degradation process continues, benzene rings of the SCP by-products could be further degraded to
329 reach ring-opening to form small molecules such as maleic acid, malic acid and oxalic acid, which
330 would be finally decomposed to water and carbon dioxide. A much detailed investigation on SCP
331 degradation will be carried out in the future.

332 In addition, PS activation rate was also checked on some catalysts. As illustrated in Fig. S8, PS shows
333 continuous decomposition with reaction time. Ni@NCNTs presented high PS decomposition rate
334 than NCNT and Ni@NCNT-800-3 demonstrated a higher PS decomposition rate than other
335 Ni@NCNTs catalysts, which confirms the higher activity of Ni@NCNT-800-3.

336 To further investigate the mineralization of SCP oxidation, total organic carbon (TOC) was
337 determined (**Fig. S9**). The TOC trend chart is well aligning with UHPLC results that SCP was
338 degraded rapidly in the first 30 min and then became relatively steady and achieved 78% of TOC
339 removal in 180 min, indicating that most of the SCP and its intermediates were oxidized into harmless
340 chemicals such as water and carbon dioxide.

341

342 **3.3 Effect of several reaction parameters on SCP removal**



343

344 **Fig.7** (a) Effect of reaction temperature on SCP removal. (b) Effect of reaction solution pH values on
 345 SCP removal. (c) Effect of PS loading on SCP removal. (d) Stability and reusability of Ni@NCNT-
 346 800-3. (Unless specific mentioned, initial SCP concentration = 20 mg/L, catalyst loading = 0.2 g/L,
 347 PS concentration = 2 g/L, and temperature at 25 °C.)

348 **Fig. 7(a)** displays the SCP removal by a synergetic effect of Ni@NCNT-800-3 adsorption and
 349 heterogeneous oxidation of PS in a range of 10-40 °C. The change of the solution temperature had a
 350 noticeable effect on the adsorption of Ni@NCNT, in an order of 0.06 mol g⁻¹ (40 °C) < 0.11 mol g⁻¹
 351 (25 °C) < 0.12 mol g⁻¹ (10 °C), suggesting a better adsorption ability to SCP at a lower temperature.
 352 The complete SCP removal was obtained in 90, 60 and 20 min at 10, 25, and 40 °C, respectively. The
 353 reaction rate constants (Table 1) at different temperatures were well fitted by the first-order kinetics

354 (Eq. 1) with great regression coefficient values ($R^2 = 0.99$) and the activation energy was calculated
355 to be 37.9 kJ mol^{-1} according to the Arrhenius equation. The first order kinetic model is listed below:

$$356 \quad \ln\left(\frac{C}{C_0}\right) = kt \quad (1)$$

357 Where k is the reaction constant and C and C_0 are the SCP concentrations at reaction time (t)= t and
358 $t=0$, respectively.

359 **Fig.7(b)** shows that Ni@NCNT-800-3 presented the best adsorption (0.11 mol g^{-1}) performance at
360 $\text{pH} = 7$. When pH reached as high as 11, the adsorptive SCP removal dropped. Moreover, it is noted
361 that a complete degradation of the antibiotic SCP was achieved in all three cases, depending on the
362 various solution pH values. The reaction rate constants are $k = 0.15, 0.09$ and 0.07 min^{-1} for $\text{pH} = 3,$
363 7 and 11 , respectively.

364 Table 1. Reaction rate constants and activation energy.

<i>Temp</i> ^o C	<i>Reaction rate constant:</i> <i>k(min</i> ⁻¹ <i>)</i>	<i>Regression coefficients:</i> <i>R</i> ²	<i>E</i> _a : <i>Activation energy</i> kJ mol ⁻¹
10	0.04	0.997	
25	0.09	0.997	37.9
40	0.21	0.999	

365

366 The influence of PS loading on SCP degradation rate was studied in the range of $1\text{-}4 \text{ g L}^{-1}$ (**Fig.7(c)**).
367 The change of the PS loading had an effect on the SCP oxidation reaction, the reaction rate constant
368 k varied from 0.05 min^{-1} for 1 g L^{-1} to 0.09 and 0.07 min^{-1} for 2 and 4 g L^{-1} , respectively. Our previous
369 studies have proven that the oxidation reaction rate is consistent with the increase of PS/PMS loading.
370 However, a further adding of PS after the optimum concentration may cause self-quenching reaction,
371 leading to reduced oxidation efficiency [54].

372 The influence of initial SCP concentration on its degradation is illustrated in Fig. S10. SCP
373 concentration also affects the degradation rate. Higher SCP concentration results in lower degradation

374 efficiency. At 100 ppm, only 93% SCP removal could be achieved in 3 h. This may be due to two
375 main factors. More adsorbed SCP molecules and intermediates would cover the active sites of
376 Ni@NCNT, hindering the activation of PS. Meanwhile, insufficient amount of PS would be a limiting
377 factor for degradation of high concentration of SCP solution.

378

379 **3.4 Stability tests**

380 The stability and reusability of prepared Ni@NCNTs and modified commercial CNT were tested by
381 similar SCP oxidation experiments after each use and heat treatment at 350 °C in nitrogen atmosphere
382 for 1 h. In **Fig. 7(d)**, Ni@NCNTs showed outstanding stability being able to achieve completely SCP
383 decomposition in 120 min in all four-run tests, where the fresh catalyst completely degraded SCP in
384 60 min, and recycled catalyst will degrade SCP in 90 and 120 min for the second and third runs,
385 respectively. What more stunning is, after ultrasonic cleaning and heat treatment of the third-run used
386 catalyst, the sample catalysis was recovered showing almost equivalent catalytic ability to the fresh
387 sample. Such good catalysis stability was comparable to or even better than some metal based
388 catalysts [55-57]. The decrease in the second and third runs was mainly due to the changes in surface
389 chemistry and structure of the catalyst. The SSA and pore structures of Ni@NCNTs were reduced
390 after the first run (Table S3). The adsorption of SCP and its intermediates for the coverage of surface
391 active sites will reduce the contact area of Ni@NCNTs and PS, thus leading to the reduction in the
392 catalytic activity. These adsorbed organic molecules can be removed via thermal annealing and the
393 catalytic activity was mostly recovered after the heat treatment. It is not excluded that some of the
394 catalysts were lost during the collection and cleaning processes in recycling runs, resulting in a lower
395 activity in the sequent test. In addition, Ni leaching in the tests was determined (Table S3) and it is
396 found that Ni ion concentrations in the four runs are much low.

397

398 **Fig. S11** shows the reusability of conventional MWCNT on PS activation for SCP degradation. It
399 decomposed 70% SCP in 180 min in the first run, and the SCP decomposition efficiencies were 49%

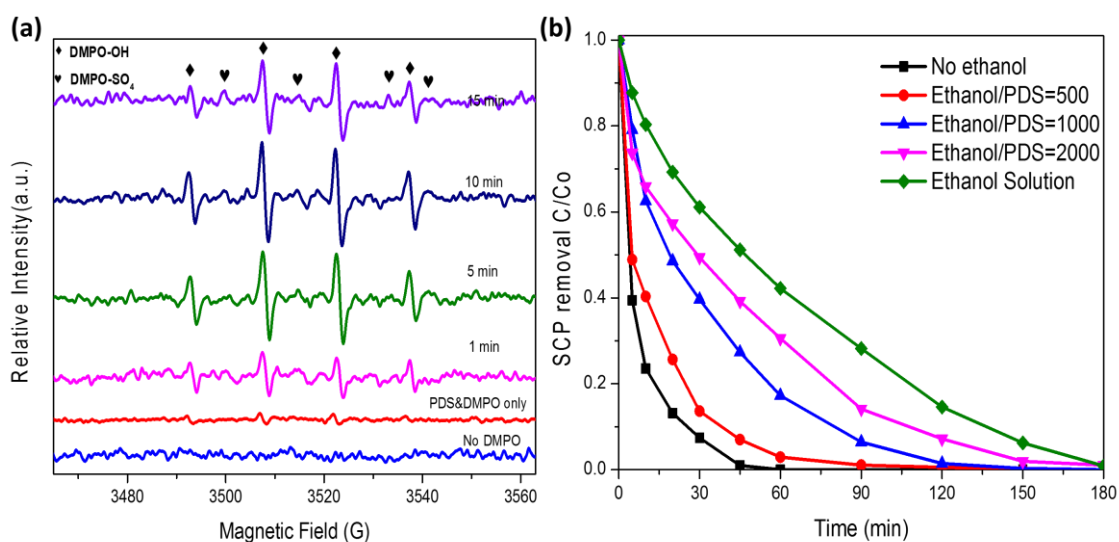
400 and 20% for the second and third runs, respectively. It can be inferred that the nickel encapsulation
401 might be acting as supporting beans inside the nanotubes and prevent the catalyst from collapsing and
402 destructing so that it could maintain good catalysis after numbers of usage. Moreover, the
403 encapsulated Ni particles can accelerate the PS activation efficiency by transferring electrons from
404 the inner Ni metals to the outer carbon tube walls. This effect can be largely conserved due to the
405 carbon wall protection when the catalyst was reused successively.

406 3.5 EPR studies

407 In PS activation involved AOPs, it is well accepted that reactive radicals such as sulfate ($\text{SO}_4^{\cdot-}$) and
408 hydroxyl radicals ($\cdot\text{OH}$) are the predominant active species in attacking organic pollutants [58-60].
409 Here, EPR was carried out to probe the detailed catalytic oxidation mechanism and identify the
410 reactive species of Ni@NCNTs activating PS for SCP degradation. A spin-trapping agent, 5, 5-
411 dimethylpyrroline-oxide (DMPO), was employed to capture the free radicals in the activation reaction.
412 **Fig. 8(a)** shows that no obvious radical peak was observed without adding DMPO and strong
413 characteristic signals of both $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ were observed after DMPO introduction. It can be seen
414 that Ni@NCNTs was able to activate PS to generate a high intensity of hydroxyl radicals and
415 relatively lower amount of sulfate radicals throughout the oxidation reaction. Duan *et al.* recently
416 investigated PS activation on phenol oxidation and suggested persulfate molecules can mainly be
417 activated and decomposed to produce $\text{SO}_4^{\cdot-}$, yet the generation of hydroxyl radicals were presumably
418 attributed to oxidization of water molecules absorbed on carbocatalyst through a one electron transfer
419 process [60]. The variation of free radical concentration versus time was depicted in **Fig. 8(a)**, PS
420 was able to produce a minor amount of radicals by self-activation without introducing a catalyst,
421 however, the intensities of radical peaks were largely enhanced after Ni@NCNTs catalyst was added
422 and reached the maximum in 10 min, which is well agreed with the change of SCP degradation rate.
423 Moreover, both $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ showed a minor receding after first 10 min, ascribed to the
424 consumption of SCP oxidation, and then appeared no obvious change afterwards. This can be
425 attributed to the outstanding catalytic ability and stability of Ni@NCNTs that can continue to activate

426 persulfate throughout the reaction. In lateral comparison among Ni@NCNT-800-1, Ni@NCNT-800-
427 3 and Ni@NCNT-800-5, all three catalysts displayed a similar change in radical concentration that
428 increased first and started to reduce after reaching the maximum (**Fig. S12**).

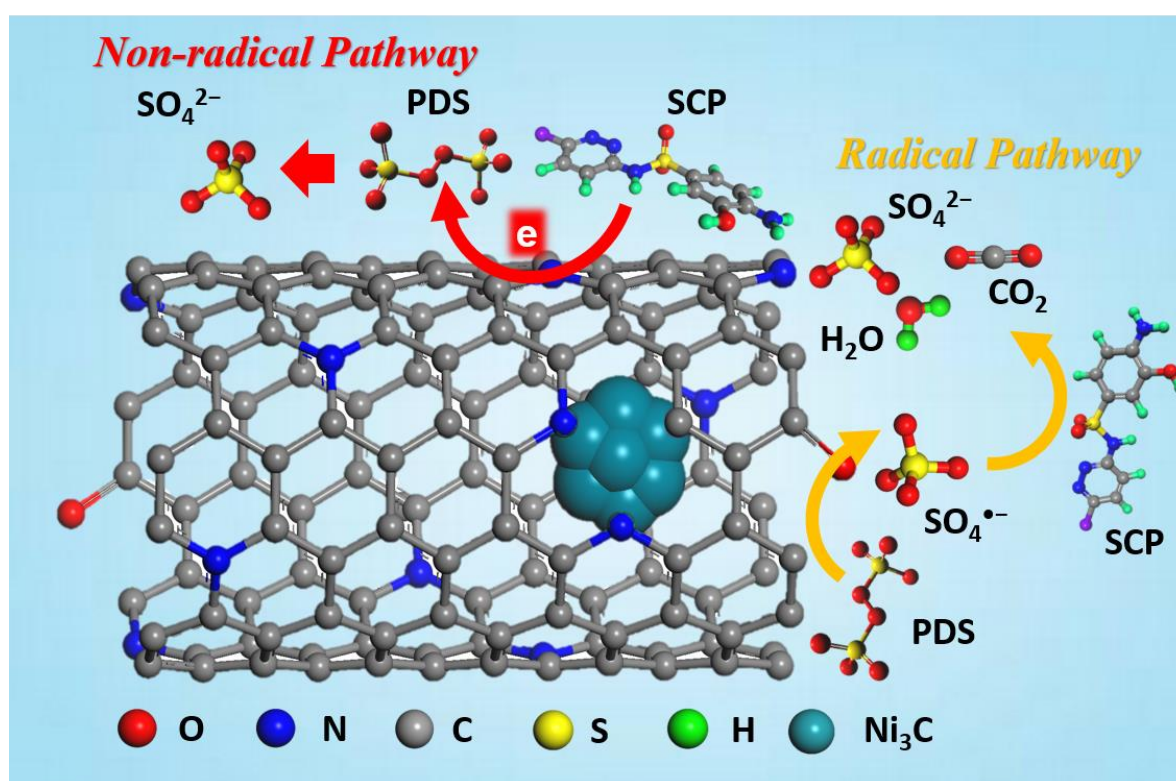
429 In quenching tests, **Fig. 8(b)** shows that SCP removal efficiency was slightly decreased when ethanol
430 was added at an ethanol/PS molar ratio of 500 and the degradation efficiency kept reducing with
431 increasing addition of ethanol. Interestingly, the SCP solution was still able to reach complete
432 removal in 180 min when the aqueous solution was completely replaced by ethanol, suggesting
433 occurrence of non-radical reactions as all radicals were quenched.



434
435 **Fig. 8** (a) EPR spectra of PS activation with Ni@NCNT-800-3 different time intervals. (b) Effect of
436 radical quenching on SCP degradation.

437
438 Other tests for identification of reactive species were also carried out. Degradation of furfuryl alcohol
439 (FFA) as a $^1\text{O}_2$ probe compound was conducted (Fig. S13a). As shown, 11% of 20 ppm of FFA was
440 removed after 3 h, suggesting that singlet oxygen provides a contribution to SCP degradation. Further
441 quenching tests by using tert-butanol (TBA) and NaN_3 for hydroxyl radical and singlet oxygen,
442 respectively, were conducted (Fig. S13b). In the presence of the two quenchers, SCP degradation was
443 reduced, suggesting the contributions of hydroxyl radical and singlet oxygen in SCP degradation.

444 In our previous studies, a series of catalysts were carried out on controlled quenching experiments,
445 cobalt oxide (Co_3O_4) and carbon nanotubes (CNT) were showing harsh catalytic decreasing after
446 ethanol was presented in the solution and the organic degradation was totally prevented when all
447 water was substituted to ethanol [30]. In contrast, nitrogen doped carbon nanotubes (N-CNT)
448 displayed a minor influence from ethanol and maintained good oxidation efficiency even in ethanol
449 solution. It was reported that the sp^2 carbon network can be activated for a better catalytic activity
450 through conjugation with the lone-pair electrons of N atoms, thus, the activated electrons would work
451 together with PS to react with the adsorbed SCP molecules [26][34]. As a result, the co-existing
452 radical and non-radical reactions in Ni@NCNTs/PS system would accelerate the completion of SCP
453 decomposition. From the experiments above and the previous studies, the possible mechanism of
454 Ni@NCNTs/PS activation system is illustrated in **Fig. 9**.



455

456 **Fig. 9** Possible reaction mechanism for organic SCP oxidation by the Ni@NCNTs/PS system.

457

458 **4. Conclusions**

459 Ni@NCNTs was synthesized by a simple pyrolysis method. The fabrication conditions such as
460 pyrolysis temperature and N precursor loading were closely related to the morphologies and catalysis
461 of Ni@NCNTs. It was observed that the Ni@NCNTs was optimally prepared at 800 °C showing the
462 most well-defined regular bamboo-like configuration and outstanding catalysis. The heterogeneous
463 system of Ni@NCNTs and PS was applied for SCP removal, showing efficient adsorption capacity
464 and oxidation efficiency. The effects of reaction temperature, solution pH, PS dosage on the reaction
465 were in detail investigated, indicating the oxidation reaction was effective in a wide pH range and
466 could be improved by increasing reaction temperature and PS amount at a controlled level. The
467 mechanistic study demonstrated that the promoted adsorption and catalytic performances of
468 Ni@NCNTs could be attributed to the unique nanotube morphology and the doping of nitrogen atoms,
469 which acted as active sites for both the SCP adsorption and PS activation. In addition, the synergistic
470 effect of N doping and Ni encapsulation enables Ni@NCNTs to present an excellent stability for
471 reuse. The mechanistic insight of catalytic reaction for PS activation was investigated by EPR and
472 quenching experiments, showing that radical ($\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$) and non-radical reactions were
473 involved in the activation processes.

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480 **Appendix A. Supplementary data**

481 Supplementary data associated with this article can be found, in the online version, at

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