

1 Soy Biodiesel Exhaust is More Toxic than Mineral
2 Diesel Exhaust in Primary Human Airway Epithelial
3 Cells.

4 *Katherine R. Landwehr^{1,2}, Jessica Hillas², Ryan Mead-Hunter¹, Rebecca O'Leary³, Gerhard
5 Knothe⁴, Anthony Kicic^{1,2,5,6,7}, Benjamin J. Mullins¹, Alexander N. Larcombe^{1,2}, on behalf of
6 AusREC^{8,9,10}, WAERP¹

7 ¹Occupation, Environment and Safety, School of Public Health, Curtin University, PO Box U1987,
8 Perth, WA, 6845

9 ²Respiratory Environmental Health, Telethon Kids Institute, Perth Children's Hospital, Nedlands,
10 Perth, WA, 6009

11 ³Department of Primary Industries and Regional Development, Perth, WA, 6008

12 ⁴USDA - Agricultural Research Service, Peoria, IL, USA, 61604

13 ⁵Department of Respiratory Medicine, Princess Margaret Hospital for Children, Perth, WA, 6001

14 ⁶School of Paediatric and Child Health, The University of Western Australia, Perth, WA, 6009

15 ⁷Centre for Cell Therapy and Regenerative Medicine, The University of Western Australia, Perth,
16 WA, 6009

17 8: Telethon Kids Institute, Centre for Health Research, The University of Western Australia,
18 Nedlands, 6009, Western Australia, Australia.

19 9: Priority Research Centre for Asthma and Respiratory Disease, Hunter Medical Research
20 Institute, Newcastle, 2305, New South Wales, Australia.

21 10: Robinson Research Institute, University of Adelaide, North Adelaide, 5006, South Australia,
22 Australia.

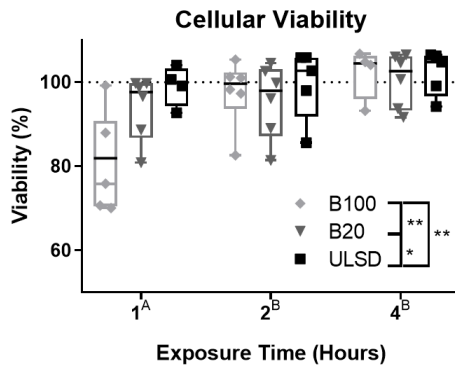
23

24 Key Words: Biodiesel, Soy Biodiesel, Diesel, Exhaust, Exhaust Exposure, Exhaust Gas
25 Analysis, Exhaust Particle Analysis, Health, Airway Epithelial Cell, Cellular Viability,
26 Inflammation, Cytokine Release,

27

28 **ABSTRACT:** As global biodiesel production increases, there are concerns over the potential
29 health impact of exposure to the exhaust, particularly in regards to young children who are at high
30 risk due to their continuing lung development. Using human airway epithelial cells obtained from
31 young children, we compared the effects of exposure to exhaust generated by a diesel engine with
32 Euro V/VI emission controls running on conventional diesel (ULSD), soy biodiesel (B100) or a
33 20% blend of soy biodiesel with diesel (B20). The exhaust output of biodiesel was found to contain
34 significantly more respiratory irritants, including NO_x, CO and CO₂ and a larger overall particle
35 mass. Exposure to biodiesel exhaust resulted in significantly greater cell death and a greater release
36 of immune mediators compared to both air controls and ULSD exhaust. These results have
37 concerning implications for potential global health impacts, particularly for the pediatric
38 population.

39 **ABSTRACT ART:**



40

41 **INTRODUCTION:**

42 Since limited battery storage capability decreases the feasibility of electrical engines in long
43 distance transport and goods shipping¹, and the inefficiency of natural gas storage limits natural
44 gas engine capabilities in long distance haulage², combustion engines are likely to be used for the
45 foreseeable future. However, as the world pushes for cleaner, renewable energy, and fossil fuels
46 become more difficult and expensive to extract, replacements for diesel fuel are currently being
47 explored. Created through the transesterification of lipids into fatty acid methyl esters³, biodiesel
48 is gaining popularity as a renewable, sustainable fuel due to its ability to directly replace diesel
49 fuel in many engines⁴. However, as biodiesel usage is predicted to increase worldwide^{5, 6},
50 concerns have been raised over the health impact of exposure to its exhaust emissions⁷.

51 Most previous studies comparing mineral diesel and biodiesel combustion have found that
52 biodiesel exhaust contains more toxic gases such as nitrogen oxides and a greater proportion of
53 smaller particles which, when inhaled, penetrate deeper into the lungs^{3, 4, 8, 9}. Despite the
54 potentially more toxic effects of biodiesel exhaust, most studies comparing biodiesel to
55 commercial mineral diesel instead focus on fuel economy and engine wear, or the physico-

56 chemical differences between the exhausts^{4, 9}. Few compare health effects to exhaust exposure^{7,}
57 ^{10, 11}. Those that do primarily use the Ames mutagenic assay^{12, 13} or immortalized cell lines^{8, 14}
58 and the majority only focus on the cytotoxic and mutagenic potential of the particulate matter,
59 ignoring the effects of the gaseous components of the exhaust entirely^{7, 15}. Particle concentrations
60 are also rarely relevant to real world exposure levels, often being far too concentrated to simulate
61 a realistic dosage¹⁵. In addition, in *in vitro* based studies, the cell lines used are not always
62 human, or even derived from respiratory tissue^{3, 16}. This brings into question their relevance in
63 human exposure studies where the main exposure route through inhalation of the exhaust means
64 that the respiratory epithelium is among the first tissue exposed and thus likely to be among the
65 most effected. Immortalized cell lines also negate genetic variability, and are limited in how
66 accurately they can model normal human tissue¹⁷.

67 As exhaust is typically inhaled, health complications can occur in the respiratory^{18, 19},
68 circulatory²⁰ and immune systems²¹. Of concern, inhalation of ultrafine exhaust particles has
69 been correlated with exacerbation of childhood asthma²², and associations between air pollution
70 from major roads and decreased lung function in children have been identified^{23, 24}. This suggests
71 children may be at greater risk from adverse health effects caused by exhaust exposure. This is
72 unsurprising as children breathe faster than adults and have higher ventilation to lung surface
73 area/body weight ratios²⁵, meaning that over the same period of time, they are exposed to a larger
74 dosage of exhaust than adults^{25, 26}. In addition, the respiratory and immune systems of children
75 are still developing and insults, such as exposure to large concentrations of exhaust, are known to
76 have lifelong consequences^{23, 27, 28}. Despite this, the effect of exposure to biodiesel exhaust has
77 not yet been studied in children.

78 Due to paucity of information in this setting we tested the hypothesis that soy biodiesel exhaust
79 would contain a greater proportion of ultrafine particles and more oxides of nitrogen and thus
80 exposure would result in more pronounced effects on the airway epithelium. To test this, we
81 exposed primary human airway epithelial cells from young healthy volunteers to whole exhaust
82 from a diesel engine fueled by either pure mineral diesel, a 20% blend of soy biodiesel with
83 mineral diesel or pure soy biodiesel. Physico-chemical exhaust properties were recorded and 24
84 hours' post exposure, cells were analyzed for a variety of health effect endpoints.

85 **MATERIALS AND METHODS:**

86 **Fuel Types and Control:** Three different fuels were used in this study; pure soy biodiesel
87 (B100) created using high quality, food grade, commercial soybean oil (MOI International
88 (AUS.)) converted via a sodium methoxide transesterification process²⁹, a 20% blend (B20) of
89 soy biodiesel in commercial ultra-low-sulfur mineral diesel (SHELL, WA, AUS) and finally,
90 commercial ultra-low-sulfur mineral diesel (ULSD). To negate background effects, HEPA
91 filtered air was used as a control exposure.

92 **Subjects:** This study was approved by the St John of God Hospital Human Ethics Committee
93 (901) and written consent was obtained from each participant's legal guardian after being fully
94 informed about the nature and purpose of the study. Here, airway epithelial cells were derived
95 from 12 healthy, typical volunteers (aged 2.7-11.2yrs, 8 males) undergoing elective surgery for
96 non-respiratory related conditions. Children with existing bacterial or viral chest infections were
97 excluded as was the diagnosis or chronic respiratory diseases including asthma and those with
98 atopy determined by a positive radioallergosorbent test (RAST) to a panel of common childhood
99 allergens.

100 **Sampling & tissue culture:** Airway epithelial cells were derived via trans-laryngeal, non
101 bronchoscopic brushing of the tracheal mucosa of children through an endotracheal tube as
102 previously described^{17, 30}. Primary cell cultures were established as previously described³¹, and
103 grown at 37°C in an atmosphere of 5%CO₂/95% air under aseptic conditions. Cultures were
104 maintained in Basal Epithelial Basal Media (BEBM® ; LONZA, Switzerland) supplemented
105 with growth additives (termed BEGM). Cells were passaged weekly and used before passage 6
106 in all experiments. Twenty four hours prior to experimentation, cells were placed in starvation
107 media, consisting of BEGM minus epithelial growth factor. Volunteers were age and gender
108 matched and split into 2 groups (n=6) whereupon the cells were exposed to either the control or
109 the exhausts. After exposure, supernatants were collected and stored at -80°C for cytokine
110 analysis. Cells were collected and stained with Annexin-V (BD Biosciences, CA, USA) before
111 being analyzed with flow cytometry.

112 **Exposure Methodology:** (See supplementary materials, Figure S1, for Exposure Diagram) All
113 exposures used exhaust generated from a Yanmar L100V engine (Yanmar, Italy). The engine is a
114 single cylinder, 435cc design coupled with a dynamometer and fitted with Euro V/VI after
115 treatment equipment consisting of an oxidation catalyst and diesel particulate filter (Daimler,
116 Germany). The engine was run at a constant load of 40% and speed of 2000rpm. Exhaust was
117 diluted 1 in 10 with HEPA filtered air inside a dilution/mixing chamber attached to the engine
118 exhaust pipe and then extracted at a rate of 10L per minute through an isokinetic sampling point,
119 leading to a sealed incubator (Model 1535, Sheldon Manufacturing, OR, USA) containing the
120 cells. The incubator was kept at 36-37°C and exhaust was injected into each cell containing dish
121 via a manifold arrangement. The exhaust then passed through a baffleplate before extraction at
122 the base of the incubator. A vacuum pump (Part No. D50819, JAVAC, VIC, AUS) and flow

123 controller (10L/min Rotameter (TSI, MN, USA)) were used to ensure a continuous flow of
124 exhaust over the cells. Exhaust removed from the incubator chamber was analyzed for physico-
125 chemical properties.

126 **Gas Measurements:** Exhaust removed from the incubator was analyzed for quantities of
127 combustion gas types using a multi-gas analyzer (TESTO 350, Testo, Lenzkirch, Germany).
128 Measurements of O₂, CO, CO₂, NO_x, NO, NO₂ and SO₂ were taken every 10 minutes.

129 **Particle Analysis:** Exhaust was analyzed for fine particle concentrations between the sizes of 3
130 and 340 nm using a Universal Scanning Mobility Particle Sizer set up (U-SMPS 1700, Palas,
131 Karlsruhe, Germany). Readings were taken every 10 minutes, starting 5 minutes into the
132 exposure to ensure adequate sampling. Mean particle size was calculated using the number of
133 particles mean. Particles were either analyzed as the total number of particles or as particles
134 separated into 2 fractions: particles below 23 nm in size and solid particles above 23 nm³².

135 **Cell Viability:** Cell viability was analyzed 24 hours after exposure using Annexin V staining
136 methodologies (Alexa Fluor® 488 Annexin V/Dead Cell Apoptosis Kit, Thermo Fisher
137 Scientific, MA, USA). Briefly, cells were suspended in 1x Annexin staining buffer and
138 incubated for 15 minutes with a 1/40 dilution of Annexin V, Alexa Fluor™ 488 conjugate
139 solution and 1µg/ml propidium iodide before undergoing flow cytometry analysis. Annexin
140 positive cells were counted as apoptotic, annexin negative/PI positive were counted as necrotic
141 and the double negative population was included as viable cells. The 24 hour incubation time
142 was chosen based on previous research showing that the effect of exhaust exposure was most
143 evident 24 hours after the exposure event⁸. Exhaust exposures were normalized to controls
144 before any statistical analysis on viability occurred.

145 **Mediators:** Cytokine release was analyzed in duplicate 24 hours after exposure using a BioRad
146 27plex human cytokine kit (BioRad, CA, USA). Cytokine release was then calculated using
147 Bio-Plex manager (v6.1.1, BioRad, Tokyo, Japan) and results normalized to cell viability.
148 Readings below detection limits were replaced with a value equal to half the concentration of the
149 lowest standard for ease of statistical analysis.

150 **Statistical Analyses:** The majority of biological results contains data for all patients (n=6),
151 excluding B100 4 hour and ULSD 1 hour exposures (n=4) and ULSD 2 and 4 hours (n=5). Data
152 are presented as mean \pm standard deviation where indicated. Gas measurements were
153 analyzed using general additive model methodologies. All other statistical analysis was performed
154 using multivariate linear regression methodologies. All statistical analyses were completed using
155 R statistical software (v3.4.3) and p-values less than 0.05 were considered significant.

156

157 **RESULTS:**

158 **Gas Analysis:** Combustion gas measurements were taken every 10 minutes over a period of 4
159 hours (figure 1). All fuels show similar trends in combustion gas production over the 4 hour
160 sampling period, with CO₂, NO, NO₂ and SO₂ all increasing rapidly within the first ~60 minutes
161 before plateauing, while O₂ decreased from atmospheric levels to ~19% during this period.
162 Carbon monoxide production peaked rapidly within the first 10 minutes before dropping below
163 detectable limits within 60 minutes. There were, however, significant differences identified
164 between fuels. Compared to ULSD, the B100 exhaust contained significantly higher production
165 of CO, CO₂ and NO₂ and significantly lower levels of O₂ over the exposure period (p<0.05,
166 figures 1a-c and 1e) and B20 contained significantly higher NO (p<0.01; figure 1d). Over the
167 entire exposure period, B100 combustion also produced a greater amount of CO₂ and NO₂

168 compared to B20 combustion ($p < 0.0001$; figure 1c and 1e), whereas B20 combustion produced
169 greater amounts of NO ($p < 0.001$; figure 1d).

170 **Particle Analysis:** An average fine particle size spectrum was obtained for each exhaust at each
171 timepoint (figure 2). The particle spectra for the B100 exhaust was significantly different to the
172 particle spectra for both the B20 and ULSD exhausts ($p < 0.05$ and $p < 0.01$ respectively). Both
173 B100 and B20 showed peaks in particle number concentration around the ultrafine particle size
174 of 100 nm. This peak was largest in B100 and not present in ULSD. In addition, the B100
175 exhaust showed a peak in particle number concentration at approximately 20 nm that was not
176 present in the other 2 exhausts. The 1 hour timepoint was also significantly different to the 2 and
177 4 hours ($p < 0.001$).

178 Particle size and concentrations were obtained for all three exhausts and timepoints (table 1). The
179 difference in particle mass and number concentration between the 1 and 4 hour exhausts was
180 smallest in ULSD and largest in B100. In addition, over 90% of the total number of particles
181 were below the size of 23 nm in all 3 exhaust for the 1 hour timepoint. Comparing within particle
182 sizes, the B100 exhaust contained the largest particle number concentration both above and
183 below 23 nm in size. Thus, the B100 exhaust, particularly in the 1 hour, had the highest particle
184 mass concentration, the largest particle number concentration and the smallest median particle
185 size, as well as the largest differences in particle mass, number and the shape of the particle
186 spectra between the 1 and 4 hour timepoints.

187 **Cellular Viability:** The B100 exhaust demonstrated the highest toxicity with significantly lower
188 viability compared with the B20 and ULSD exposures (figure 3, $p < 0.001$ and $p < 0.0001$

189 respectively). The largest differences in mean viability occurred in the 1 hour exposure, with
190 B100 demonstrating 14.2% and 19.2% more cell death than B20 and ULSD respectively.

191 Comparing timepoints, the 1 hour exposure consistently demonstrated the lowest viability in all
192 fuels, followed by the 2 hour exposure with the second lowest viability. The 1 hour exposure was
193 significantly different to both the 2 and 4 hour exposures ($p < 0.001$ and $p < 0.0001$ respectively).
194 Up to 5% cell death was observed in the 2 hour exposure, however this was not significantly
195 different to the 4 hour exposure timepoint, which consistently demonstrated the highest viability
196 in all fuels. The difference in viability between the 1 and 4 hour timepoints was largest in the
197 B100 fuel, with a 22.3% mean difference, and smallest in ULSD, which demonstrated only a
198 3.0% mean difference. The mean difference in cell death for B20 is 6.5%, which is 18.1% more
199 toxic than ULSD when comparing the overall toxicity of ULSD and B100.

200 Mechanisms of cell death changed significantly after exhaust exposure, with all fuels showing
201 significantly increased necrotic cell death in comparison to the controls (figure 3b, $p < 0.001$), but
202 not each other. Significant differences also occurred between timepoints, with the 4 hour
203 exposure timepoint demonstrating significantly lower levels of necrotic cell death in comparison
204 to the 2 hour timepoint ($p < 0.05$).

205 **Cytokine Release:** Of the 27 cytokines tested, only 11 were released at measurable
206 concentrations; IL-1 β , IL-1RA, IL-6, IL-8, G-CSF, GM-CSF, IP-10, MIP-1 β , IP-10 and
207 RANTES. The results for these 11 cytokines have been displayed in graphical form (figure 5).
208 Both IL-6 and GM-CSF were produced at significantly increased amounts after most exhaust
209 exposures in comparison to the controls ($p < 0.05$ in all cases) with the B20 inducing release of
210 GM-CSF, although this elevated production was not significant ($p = 0.06$). Both RANTES and IP-

211 10 release was significantly decreased after most exhaust exposures ($p < 0.05$ in all cases) with the
212 B100 release of IP-10 decreasing, although not significantly ($p = 0.06$). Exposure to B100 exhaust
213 resulted in significantly increased production of IL-1RA, IL-8, G-CSF and MIP-1 β ($p < 0.05$ in all
214 cases) while ULSD exposure resulted in significantly increased production of IL-1RA, IL-1 β and
215 VEGF ($p < 0.05$ in all cases). Comparing timepoints, the 1 hour exposures had significantly
216 increased release of IL-1 β , IL-6, G-CSF, MIP-1 β and GM-CSF ($p < 0.05$ in all cases) and after 4
217 hours of exposure there was significantly increased release of IL-1RA, VEGF, TNF- α and
218 RANTES ($p < 0.05$ in all cases).

219 **DISCUSSION:**

220 The results of this study show that exposure to mineral diesel, pure soy biodiesel or a 20% blend
221 of soy biodiesel in mineral diesel induced airway epithelial cell death, increased the percentage
222 of necrotic cell death mechanisms and increased the release of immune modulating cytokines
223 compared to control cells. Exhaust characteristics varied significantly between all three fuel
224 types, with B100 containing significantly higher levels of respiratory irritants including NO₂,
225 CO, CO₂ and ultrafine particulate matter at a smaller median particle size, in comparison to both
226 B20 and ULSD. The B20 exhaust contained significantly higher levels of NO in comparison to
227 both B100 and ULSD and more particles than ULSD. Correspondingly, B100 exhaust was
228 significantly more toxic than both B20 and ULSD, resulting in a higher percentage of cell death
229 and the increased release of the largest number of cytokines, particularly in the first hour of
230 exposure. The B20 exhaust was second most toxic with significantly more cell death than ULSD.
231 In contrast, ULSD exposure resulted in a higher release of cytokines than the B20 exposure,
232 suggesting that mineral diesel is more immunogenic. Thus, exposure to the exhaust of all 3 fuels
233 resulted in toxic effects on human airway epithelial cells.

234 Combustion of both diesel and biodiesel results in the generation of toxic gases and respiratory
235 irritants such as carbon dioxide, carbon monoxide, nitrogen oxides and sulfur dioxide⁸. The
236 results of previous studies comparing the gaseous outputs of the two exhausts vary considerably⁴,
237 ^{9, 33, 34}. The majority have found that biodiesel exhaust contains more NO_x as well as a decrease
238 in the average size of the particulate matter ^{4, 7, 8} and this study has found similar results. Long
239 term exposure to NO_x, (made primarily of NO and NO₂) is associated with decreased lung
240 volume, increased allergen response and increased risk of respiratory infections³⁵. Nitrogen
241 monoxide readily oxidizes into nitrogen dioxide at atmospheric conditions³⁶ and nitrogen dioxide
242 interacts with moisture to form nitric acid³⁵. In addition, NO_x is a major contributor to
243 photochemical smog and ozone pollution³⁶. Even at environmental levels, NO_x exposure is
244 implicated with increased health risks, including ischemic stroke and cardiovascular disease^{35, 37}.
245 In children, daily exposure to NO₂ is associated with increased asthmatic symptoms^{35, 38}, which
246 is of concern as both B100 and B20 exhaust contained significantly higher levels of NO₂, either
247 as the pure gas or as NO with the potential to oxidize. As the exhaust used in this study was first
248 diluted with HEPA filtered air in order to simulate a realistic exposure dosage, changes in the
249 sampled exhaust are over 10 fold greater in undiluted exhaust, making increased environmental
250 NO_x levels a potential health threat as biodiesel usage continues to increase.

251 Carbon monoxide oxidizes slowly at atmospheric conditions³⁹ and exposure to low levels of CO
252 for an extended period of time is associated with adverse neurological impacts including
253 emotional instability and difficulty concentrating⁴⁰. The 0.28% increase in CO₂ levels between
254 B100 and ULSD at the end of the 4 hour exposure translates to approximately a 28000 ppm
255 difference in the undiluted exhausts, and represents a considerable increase in total CO₂
256 production. Short-term exposure to only 1500 ppm is associated with cognitive impairment⁴¹ and

257 negative effects on bone formation and blood pressure have been observed at 12000 ppm^{42, 43}.
258 This makes increased biodiesel usage a concern in terms of increased atmospheric CO and CO₂
259 levels, which will in turn have potential impacts on population health and climate change.

260 The particulate matter phase of exhaust is defined as the inert elemental carbon particles created
261 during combustion and the potentially toxic chemicals adhered to their surface⁴⁴; including
262 polyaromatic hydrocarbons, aldehydes, ketones and heavy metals^{33, 34}. Despite the fact that the
263 elemental carbon particles within the particulate matter itself are relatively inert, inhalation
264 causes respiratory irritation¹⁹ and the inhalation of ultrafine particulate matter is associated with
265 worsening health outcomes including pulmonary inflammation⁴⁵, exacerbation of existing
266 respiratory diseases⁴⁶, increased blood pressure and increased risk of heart failure as the small
267 size of the particles allows entrance directly into the cardiovascular system, bypassing the lung
268 barrier function entirely^{47, 48}. Asthma sufferers are particularly susceptible to adverse effects
269 from inhalation, and exacerbation of childhood asthma has been associated with ultrafine
270 particulate pollution²².

271 In addition, inhalation of the particulate matter allows the toxic chemical adhered to the surface
272 to deposit within the lungs and cardiovascular system. Smaller particles have an increased
273 surface to volume ratio, meaning that a greater amount of harmful substances can adhere to the
274 surface for a given mass^{8, 49}. Smaller particles are also able to penetrate deeper into the lungs,
275 causing an irritant effect as well as potentially depositing the harmful chemicals over a greater
276 percentage of the total lung tissue⁴⁵. In this study, biodiesel combustion was found to generate a
277 greater amount of particles, with a smaller mean diameter, making the adverse effects from the
278 inhalation of these chemicals a greater concern, alongside the potential health effects of
279 increased ultrafine particulate pollution. Observing the change in particle spectra over time, the

280 greatest number concentration of ultrafine particles was found in the first hour of the exhaust for
281 all three fuels. The 1 hour B100 exhaust contained both the greatest particle number
282 concentration, above and below 23 nm in size, and the smallest median particle diameter. In
283 addition, it was the only exhaust where the majority of particles below 23 nm in size persisted
284 into the 2nd hour. This change in fine particles over time is likely due to diesel particulate filter
285 loading, as well as possible condensation effects³².

286 Comparing exhaust characteristics between the 1, 2 and 4 hour timepoints, two characteristics
287 stand out in the first hour; the carbon monoxide readings and the particulate matter. The
288 difference in these two characteristics can be attributed to the cold start effect, where the engine
289 is started at below optimum working temperature, and thus the devices such as the catalytic
290 converter and diesel particulate filter are also working at below optimum temperature. This helps
291 to explain why these readings decrease as exposure time, and thus engine temperature, increases.
292 All other characteristics are higher in the 4 hour timepoint, excluding oxygen.

293 The effects of exposure to all exhausts can be observed in the increased cell death for both B100
294 and B20 and the increased release of cytokines for all exhausts. In addition, mechanisms of cell
295 death skew towards necrosis after exhaust exposure in comparison to air, suggesting a high level
296 of cell injury and trauma⁵⁰. Previous studies using adult human airway epithelial cell lines have
297 found similar results, with exposure to diesel exhaust inducing necrotic cell death⁵¹. The
298 cytokine release after exhaust exposure is strongly indicative of an increased inflammatory
299 response with biodiesel having the most pronounced effect, shown by an increase in the release
300 of 6 cytokines in comparison to the controls; IL-1RA, IL-6, IL-8, G-CSF, GM-CSF and MIP-1 β .
301 The next highest inflammatory response is induced by ULSD exhaust exposure, causing
302 increased release of 5 cytokines in comparison to the controls; IL-1RA, IL-1 β , IL-6, VEGF and

303 GM-CSF. Finally, the B20 exhaust exposure significantly increased the release of 1 cytokine, IL-
304 6, indicating a smaller inflammatory response.

305 Very few studies have focused on the effect of biodiesel exposure on human airway epithelial
306 cell mediator production, particularly in less commonly studied cytokines such as IL-1RA, G-
307 CSF and VEGF. The majority of studies that do test the immunogenic effect of biodiesel
308 exposure focus on the release of IL-6 and IL-8, with sporadic attention paid to few other
309 cytokines^{7, 8, 52, 53}. The difference in the release of the inflammatory cytokine panel tested in this
310 study suggests that this may be an oversight, with different exhausts inducing the production of
311 different cytokines in comparison to air exposed controls. For example exposure to ULSD
312 exhaust caused significant release of IL-1 β , which is associated with the acute inflammatory
313 response^{54, 55}, while exposure to soy biodiesel exhaust caused significant release of IL-8 and G-
314 CSF, which are associated with neutrophilic inflammation^{55, 56}. This suggests that exposure to the
315 exhaust of either petroleum diesel or pure soy biodiesel induces different immune reactions,
316 suggesting variations in the nexus points between the adaptive and innate immune responses^{19, 54-}
317 ⁵⁸.

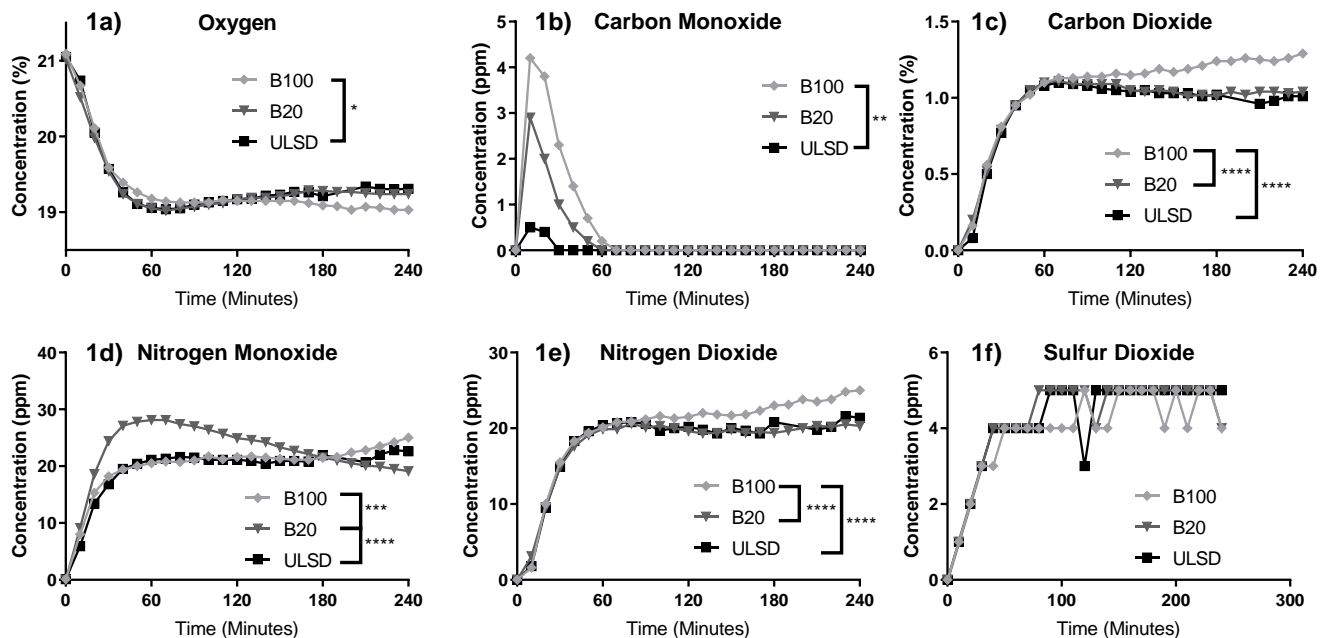
318 Interestingly, the release of cytokines also shows a pattern in response to exposure time. The 4
319 hour exposures show significantly increased production of 4 cytokines in comparison to the other
320 two timepoints; IL-1RA, VEGF, TNF- α and RANTES. The 1 hour exposures show increased
321 production of 5 cytokines in comparison to the other two timepoints; IL-1 β , IL-6, G-CSF, GM-
322 CSF and MIP-1 β . The 2 hour exposures show no significant increase in immune mediators in
323 comparison to the other two timepoints. This suggests that the 1 hour exposure time induces a
324 slightly greater immune response in comparison to the 4 hour exposures and both the 1 and 4
325 hour exposures induce a greater immune response than the 2 hour exposures. Combined with the

326 significantly increased cell death in the 1 hour exposures, with some cell death and significantly
327 increased necrotic cell death also observed in the 2 hour exposure, the earlier timepoints are
328 more toxic than the later, with the 1 hour timepoint showing the most toxic effects followed by
329 the 2 hour timepoint.

330 As the health effects of carbon monoxide are attributed to the binding of hemoglobin and thus
331 low blood oxygen levels instead of a direct toxic effects on cells^{59, 60}, and the presence of carbon
332 monoxide is eliminated in the exhaust in all fuels by the second hour via the catalytic converter
333 and is thus unlikely to explain the presence of increased necrotic cell death at that timepoint, the
334 cause of the more toxic effects in the 1 hour are likely to be attributed to the particulate matter
335 which has a greater mass and number concentration and a smaller median size in the 1 hour
336 when compared to both the 2 and 4 hour timepoints.

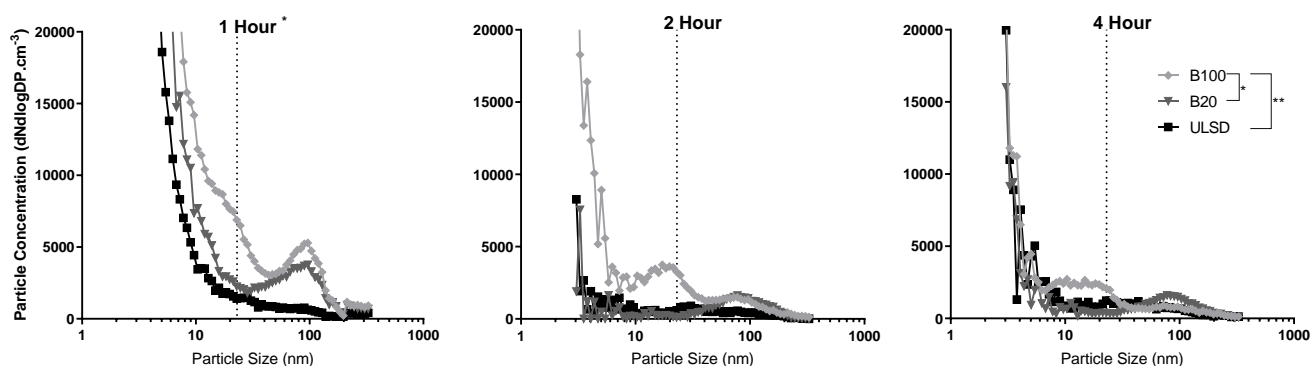
337 This is not the first study to find that exposure to a lower dosage of particles results in the most
338 toxic effects^{61, 62}. Previous studies have attributed this effect to particle agglomeration and we
339 propose a similar effect occurred in this study. Smaller particles can more easily penetrate the
340 cell membrane and cause injury to the cells. As exposure time increases, more and larger
341 particles are added to the media. As diesel particulate matter is known to agglomerate readily in
342 liquid¹⁹, we propose that the addition of the larger particles causes particle concentration to reach
343 a point that agglomeration occurs with the initial smaller, more damaging particles, preventing
344 them from penetrating the cell membrane. This explains why the most cell death is observed in
345 the 1 hour exposure where particle dosage is lowest and median particle size is smallest. A
346 smaller amount of cell death is observed in the 2 hour, which has the second lowest particle
347 dosage, and no death observed by the 4 hour timepoint. Indications of cell injury, including
348 inflammatory cytokine release and necrotic cell death also decrease over time.

349 In conclusion, in our study, soy-based biodiesel exhaust contained more and smaller particulate
 350 matter, more gaseous respiratory irritants and exposure to the exhaust resulted in a higher
 351 percentage of cell death and a wider release of cytokines for a more varied immune reaction in
 352 comparison to mineral diesel exposure. As biodiesel usage becomes more widespread,
 353 environmental NO_x levels and ultrafine particulate pollution are likely to increase, leading to
 354 worrying concerns on health impacts in the wider community, particularly the impacts on
 355 childhood asthma severity.



356
 357 **Figure 1:** Combustion gas analysis from the diluted exhaust of the three different fuels types.

358 Measurements were taken every 10 minutes for 4 hours (*=p value<0.5, **=p value<0.01, ***=p
 359 value<0.01, ****=p value<0.001). Figure 1a) and 1c) concentration measurements as a
 360 percentage, all other figures show concentration in parts per million (ppm).



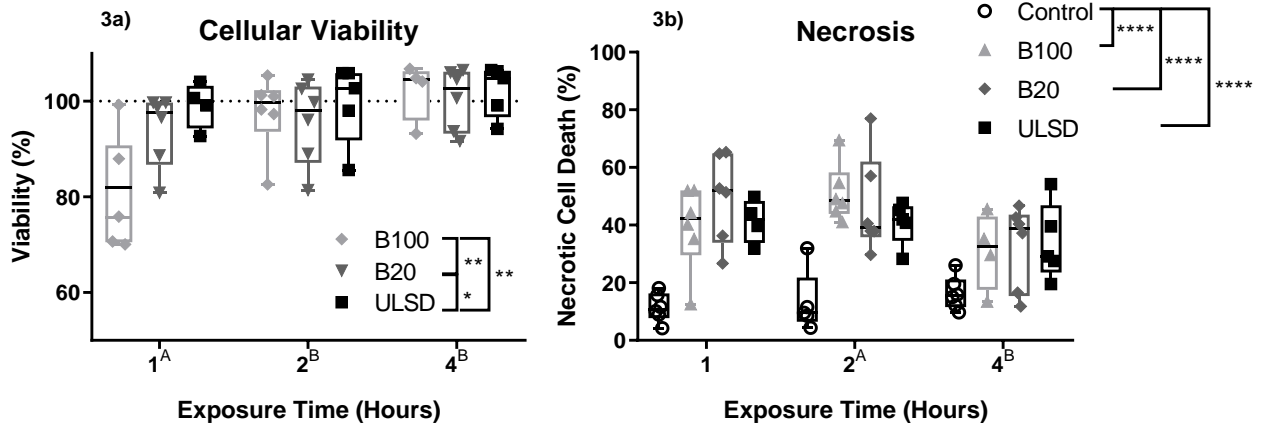
361
 362 **Figure 2:** Particle size spectra for all three fuels (*= p value <0.05 , **= p value <0.01 , ****= p
 363 value <0.001). The dotted line indicates the particle size of 23 nm. Within fuels, particle size spectra
 364 are significantly different between the 1 hour and the 2 and 4 hour timepoints ($p<0.001$). Both
 365 B100 and B20 show peaks around the ultrafine particle size of 100 nm which is absent in the
 366 mineral diesel exhaust.

367 **Table 1:** Particle characteristics between the sizes of 3-340 nm for all fuels and timepoints.

Fuel	ULSD			B20			B100		
Time (Hours)	1	2	4	1	2	4	1	2	4
Particle Concentration ($\mu\text{g}/\text{m}^3$)	34	10	19	84 (2.76) ^a	19 (1.90)	27 (1.42)	105 (3.09)	29 (2.9)	24 (1.26)
Median Particle Size (nm)	20	20	10	10	10	20	<10	20	10
Total Particle Number (particles/ cm^3)	521539	47765	113216	881298 (1.69)	45690 (0.96)	94510 (0.83)	1301691 (2.50)	224339 (4.70)	148467 (1.31)
Particle Number >23 nm (particles/ cm^3)	24117 [4.62%] ^b	15044 [31.50%]	23298 [20.58%]	75980 [8.62%]	26894 [58.86%]	28641 [30.30%]	113029 [8.68%]	40377 [18.00%]	24595 [16.57%]
Particle Number <23 nm (particles/ cm^3)	497422 [95.38%]	32721 [68.50%]	89918 [79.42%]	805319 [91.38%]	18797 [41.14%]	65869 [69.70%]	1188662 [91.32%]	183962 [82.00%]	123872 [83.43%]

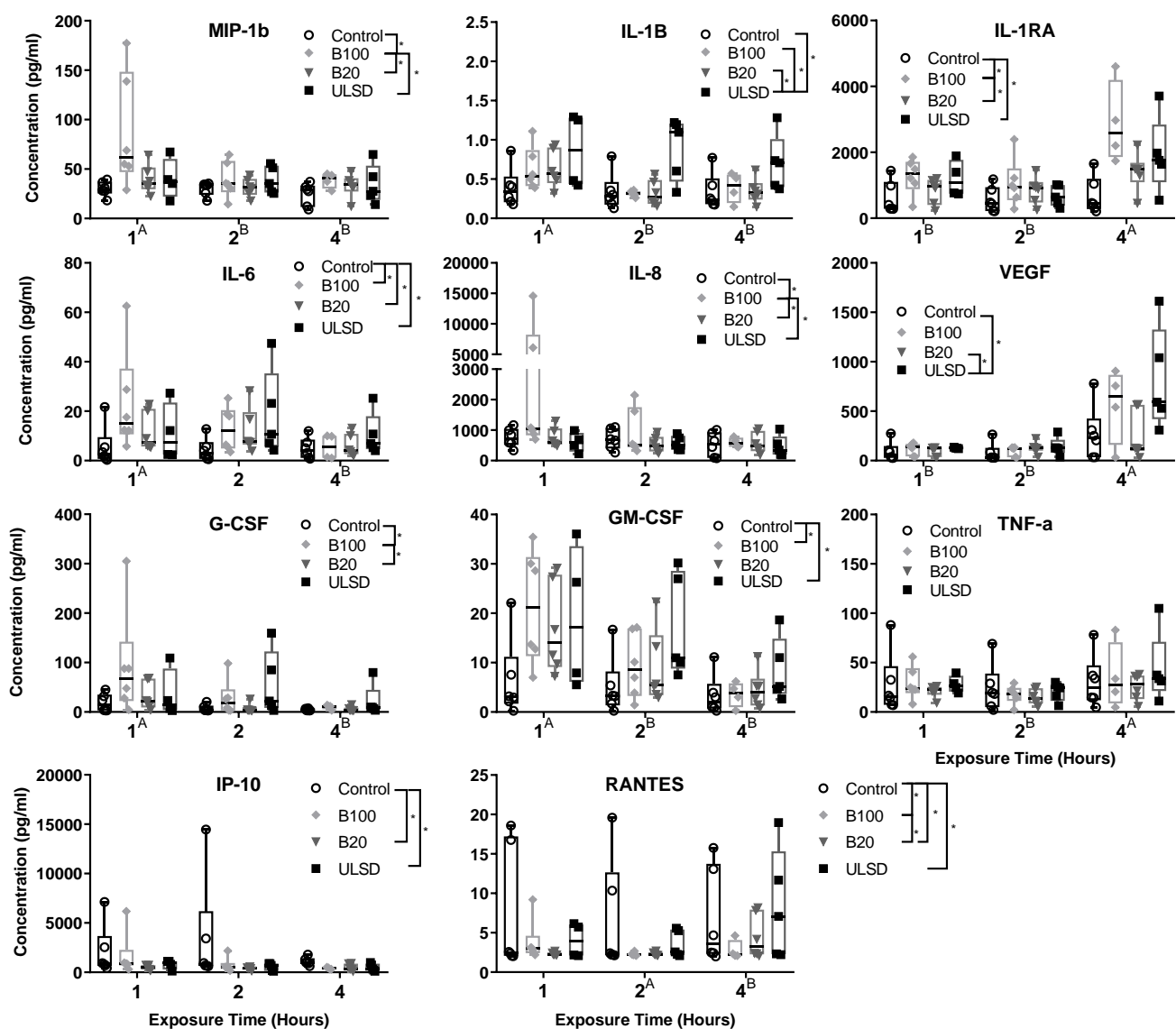
368 ^a Values in round brackets represent proportional increases in comparison to ULSD values.

369 ^b Values in square brackets represent the percentage of the total particle number concentration.



371

372 **Figure 3:** a) Cell viability measurements 24 hours after exposure using Annexin V staining. All
 373 results are normalized to control measurements (dotted line). 3b) Percentage of cell death via
 374 necrotic mechanisms 24 hours after exposure. Asterisk symbols on legend indicate significance
 375 between fuels (*=p value<0.05, **=p value<0.01, ****=p value<0.0001). Superscripts on x-axis
 376 indicate significant differences across time. A superscript of “A” indicates significant increase to
 377 a superscript of “B” (3a) p<0.001 and p<0.0001 for 1 vs 2 and 4 hours respectively, 3b) p<0.05).
 378 Bars indicate spread of data and median.



379

380 **Figure 4:** Measured cytokine release for all fuels and times for 9 of the 11 cytokines released
 381 above limit of detection (*= p value <0.05). A superscript of A indicates significant increase to a
 382 superscript of B between timepoints ($p<0.05$). Boxplots indicate spread of data and median values.

383 **ASSOCIATED CONTENT**

384 Supporting information (PDF) including a diagram of the experimental setup and flow cytometry
 385 gating is available free of charge at:

386 **DISCLOSURE:** The authors of this paper have no conflicts of interest to declare.

387 AUTHOR INFORMATION

388 **Corresponding Author**

389 *Email: Katherine.Landwehr@telethonkids.org.au

390 **Author Contributions**

391 KRL, JH, AK, BJM and ANL performed the exposures. BJM and RM-H created the fuels with
392 input from GK. KRL, JH and AK grew the cells for exposure. KRL performed analysis assays
393 with input from JH and flow cytometry data analysis was performed by JH. RO'L advised on
394 statistical analyses. The manuscript was written through contributions of all authors. All authors
395 have given approval to the final version of the manuscript.

396 **Funding Sources**

397 This research was supported by the ARC Discovery Project Grant- DP170104346 and Curtin
398 University, Western Australia.

399 **ACKNOWLEDGMENT**

400 We would like to acknowledge the Australian Research Council for funding via the Discovery
401 Project Grant DP170104346 as well as Curtin University's School of Public Health and Graduate
402 Research School for their support and funding. We would like to the West Australian Epithelial
403 Research Program for supplying the patient samples. Thank you to all participants and families
404 for contributing to the study.

405 **ABBREVIATIONS**

406 B100, Pure soy biodiesel; B20, A 20% blend of soy biodiesel in mineral diesel; ULSD,
407 commercial mineral diesel, O₂, oxygen; CO, carbon monoxide, CO₂, carbon dioxide; NO_x,

408 nitrogen oxides; NO, nitrogen monoxide; NO₂, nitrogen dioxide; SO₂, sulfur dioxide; IL-1 β ,
409 interleukin 1 beta; IL-1RA, interleukin 1 receptor antagonist; IL-6, interleukin 6; IL-8,
410 interleukin 8; G-CSF, granulocyte colony stimulating factor; GM-CSF, granulocyte-macrophage
411 colony stimulating factor; IP-10, interferon gamma inducible protein 10; MIP- β , macrophage
412 inflammatory protein; RANTES, regulated on activation, normal T-cell expressed and secreted;
413 TNF- α , tumour necrosis factor alpha; VEGF, vascular endothelial growth factor.

414 **REFERENCES**

- 415 1. Amjad, S.; Neelakrishnan, S.; Rudramoorthy, R., Review of design considerations and
416 technological challenges for successful development and deployment of plug-in hybrid electric
417 vehicles. *Renewable and Sustainable Energy Reviews* **2010**, *14*, (3), 1104-1110.
- 418 2. Camuzeaux, J. R.; Alvarez, R. A.; Brooks, S. A.; Browne, J. B.; Sterner, T., Influence of
419 Methane Emissions and Vehicle Efficiency on the Climate Implications of Heavy-Duty Natural
420 Gas Trucks. *Environmental Science & Technology* **2015**, *49*, (11), 6402-6410.
- 421 3. Bünger, J.; Krahl, J.; Baum, K.; Schröder, O.; Müller, M.; Westphal, G.; Ruhnau, P.;
422 Schulz, T. G.; Hallier, E., Cytotoxic and mutagenic effects, particle size and concentration
423 analysis of diesel engine emissions using biodiesel and petrol diesel as fuel. *Archives of*
424 *Toxicology* **2000**, *74*, (8), 490-498.
- 425 4. Fontaras, G.; Karavalakis, G.; Kousoulidou, M.; Tzamkiozis, T.; Ntziachristos, L.;
426 Bakeas, E.; Stournas, S.; Samaras, Z., Effects of biodiesel on passenger car fuel consumption,
427 regulated and non-regulated pollutant emissions over legislated and real-world driving cycles.
428 *Fuel* **2009**, *88*, (9), 1608-1617.
- 429 5. Organisation for Economic Co-operation and Development, OECD-FAO agricultural
430 outlook 2015–2025 data. <http://stats.oecd.org/index.aspx?queryid=48184#>
- 431 6. EIA, International energy statistics.
432 https://www.eia.gov/beta/international/data/browser/#/?pa=000000g&c=41000000020000600000000000g0002000000000000000001&tl_id=79-A&vs=~~~~~INTL.81-1-WORL-TBPD.A&cy=2014&vo=0&v=T&start=2000&end=2014&showdm=y
- 434 [A&vs=~~~~~INTL.81-1-WORL-](https://www.eia.gov/beta/international/data/browser/#/?pa=000000g&c=41000000020000600000000000g0002000000000000000001&tl_id=79-A&vs=~~~~~INTL.81-1-WORL-TBPD.A&cy=2014&vo=0&v=T&start=2000&end=2014&showdm=y)
435 [TBPD.A&cy=2014&vo=0&v=T&start=2000&end=2014&showdm=y](https://www.eia.gov/beta/international/data/browser/#/?pa=000000g&c=41000000020000600000000000g0002000000000000000001&tl_id=79-A&vs=~~~~~INTL.81-1-WORL-TBPD.A&cy=2014&vo=0&v=T&start=2000&end=2014&showdm=y)
- 436 7. Larcombe, A. N.; Kicic, A.; Mullins, B. J.; Knothe, G., Biodiesel exhaust: The need for a
437 systematic approach to health effects research. *Respirology* **2015**, *20*, (7), 1034-1045.
- 438 8. Mullins, B. J.; Kicic, A.; Ling, K.-M.; Mead-Hunter, R.; Larcombe, A. N., Biodiesel
439 Exhaust–Induced Cytotoxicity and Proinflammatory Mediator Production in Human Airway
440 Epithelial Cells. *Environmental Toxicology* **2016**, *31*, (1), 44-57.
- 441 9. Graver, B. M.; Frey, H. C.; Hu, J., Effect of Biodiesel Fuels on Real-World Emissions of
442 Passenger Locomotives. *Environmental Science & Technology* **2016**, *50*, (21), 12030-12039.
- 443 10. Swanson, K. J.; Madden, M. I. C.; Ghio, A. J., Biodiesel Exhaust: The Need for Health
444 Effects Research. *Environmental Health Perspectives* **2007**, *115*, (4), 496-499.

- 445 11. Madden, M. C., A paler shade of green? The toxicology of biodiesel emissions: Recent
446 findings from studies with this alternative fuel. *Biochimica et Biophysica Acta - General Subjects*
447 **2016**, *1860*, (12), 2856-2862.
- 448 12. Bünger, J.; Müller, M. M.; Krahl, J.; Baum, K.; Weigel, A.; Hallier, E.; Schulz, T. G.,
449 Mutagenicity of diesel exhaust particles from two fossil and two plant oil fuels. *Mutagenesis*
450 **2000**, *15*, (5), 391-397.
- 451 13. Mutlu, E.; Warren, S. H.; Matthews, P. P.; Schmid, J. E.; Kooter, I. M.; Linak, W. P.;
452 Gilmour, I. M.; DeMarini, D. M., Health effects of soy-biodiesel emissions: bioassay-directed
453 fractionation for mutagenicity. *Inhalation Toxicology* **2015**, *27*, (11), 597-612.
- 454 14. Cervena, T.; Rossnerova, A.; Sikorova, J.; Beranek, V.; Vojtisek-Lom, M.; Ciganek, M.;
455 Topinka, J.; Rossner, P., DNA Damage Potential of Engine Emissions Measured In Vitro by
456 Micronucleus Test in Human Bronchial Epithelial Cells. *Basic & Clinical Pharmacology &*
457 *Toxicology* **2017**, *121*, 102-108.
- 458 15. André, V.; Barraud, C.; Capron, D.; Preterre, D.; Keravec, V.; Vendeville, C.; Cazier, F.;
459 Pottier, D.; Morin, J. P.; Sichel, F., Comparative mutagenicity and genotoxicity of particles and
460 aerosols emitted by the combustion of standard vs. rapeseed methyl ester supplemented bio-
461 diesel fuels: Impact of after treatment devices: Oxidation catalyst and particulate filter. *Mutation*
462 *Research-Genetic Toxicology and Environmental Mutagenesis* **2015**, *777*, 33-42.
- 463 16. Jalava, P. I.; Aakko-Saksa, P.; Murtonen, T.; Happonen, M. S.; Markkanen, A.; Yli-Pirilä, P.;
464 Hakulinen, P.; Hillamo, R.; Mäki-Paakkanen, J.; Salonen, R. O.; Jokiniemi, J.; Hirvonen, M.-R.,
465 Toxicological properties of emission particles from heavy duty engines powered by conventional
466 and bio-based diesel fuels and compressed natural gas. *Particle and Fibre Toxicology* **2012**, *9*,
467 (1), 37.
- 468 17. Kicic, A.; Sutanto, E. N.; Stevens, P. T.; Knight, D. A.; Stick, S. M., Intrinsic
469 Biochemical and Functional Differences in Bronchial Epithelial Cells of Children with Asthma.
470 *American Journal of Respiratory and Critical Care Medicine* **2006**, *174*, (10), 1110-1118.
- 471 18. Benbrahim-Tallaa, L.; Baan, R. A.; Grosse, Y.; Lauby-Secretan, B.; El Ghissassi, F.;
472 Bouvard, V.; Guha, N.; Loomis, D.; Straif, K., Carcinogenicity of diesel-engine and gasoline-
473 engine exhausts and some nitroarenes. *The Lancet Oncology* **2012**, *13*, (7), 663-664.
- 474 19. Larcombe, A. N.; Phan, J. A.; Kicic, A.; Perks, K. L.; Mead-Hunter, R.; Mullins, B. J.,
475 Route of exposure alters inflammation and lung function responses to diesel exhaust. *Inhalation*
476 *Toxicology* **2014**, *26*, (7), 409-418.
- 477 20. Mills, N. L.; Törnqvist, H.; Gonzalez, M. C.; Vink, E.; Robinson, S. D.; Söderberg,
478 S.; Boon, N. A.; Donaldson, K.; Sandström, T.; Blomberg, A.; Newby, D. E., Ischemic and
479 Thrombotic Effects of Dilute Diesel-Exhaust Inhalation in Men with Coronary Heart Disease.
480 *New England Journal of Medicine* **2007**, *357*, (11), 1075-1082.
- 481 21. Nejad, S. H.; Takechi, R.; Mullins, B. J.; Giles, C.; Larcombe, A. N.; Bertolatti, D.;
482 Rumchev, K.; Dhaliwal, S.; Mamo, J., The effect of diesel exhaust exposure on blood-brain
483 barrier integrity and function in a murine model. *Journal of Applied Toxicology* **2015**, *35*, (1),
484 41-47.
- 485 22. Evans, K. A.; Halterman, J. S.; Hopke, P. K.; Fagnano, M.; Rich, D. Q., Increased
486 ultrafine particles and carbon monoxide concentrations are associated with asthma exacerbation
487 among urban children. *Environmental Research* **2014**, *129*, 11-19.
- 488 23. Gauderman, W. J.; Vora, H.; McConnell, R.; Berhane, K.; Gilliland, F.; Thomas, D.;
489 Lurmann, F.; Avol, E.; Kunzli, N.; Jerrett, M.; Peters, J., Effect of exposure to traffic on lung
490 development from 10 to 18 years of age: a cohort study. *The Lancet* **2007**, *369*, (9561), 571-577.

- 491 24. Gauderman, W. J.; Urman, R.; Avol, E.; Berhane, K.; McConnell, R.; Rappaport, E.;
492 Chang, R.; Lurmann, F.; Gilliland, F., Association of Improved Air Quality with Lung
493 Development in Children. *New England Journal of Medicine* **2015**, *372*, (10), 905-913.
- 494 25. Ginsberg, G. L.; Perkovich Foos, B.; Firestone, M. P., Review and Analysis of Inhalation
495 Dosimetry Methods for Application to Children's Risk Assessment. *Journal of Toxicology and*
496 *Environmental Health, Part A* **2005**, *68*, (8), 573-615.
- 497 26. Saadeh, R.; Klaunig, J., Child's Development and Respiratory System Toxicity. *Journal*
498 *of Environmental & Analytical Toxicology* **2014**, *4*, (5), 1.
- 499 27. Svanes, C.; Omenaas, E.; Jarvis, D.; Chinn, S.; Gulsvik, A.; Burney, P., Parental smoking
500 in childhood and adult obstructive lung disease: results from the European Community
501 Respiratory Health Survey. *Thorax* **2004**, *59*, (4), 295-302.
- 502 28. Chen, Z.; Salam, M. T.; Eckel, S. P.; Breton, C. V.; Gilliland, F. D., Chronic effects of air
503 pollution on respiratory health in Southern California children: findings from the Southern
504 California Children's Health Study. *Journal of Thoracic Disease* **2015**, *7*, (1), 46-58.
- 505 29. Knothe, G.; de Castro, M. E. G.; Razon, L. F., Methyl Esters (Biodiesel) from and Fatty
506 Acid Profile of *Gliricidia sepium* Seed Oil. *Journal of the American Oil Chemists' Society* **2015**,
507 *92*, (5), 769-775.
- 508 30. Lane, C.; Burgess, S.; Kicic, A.; Knight, D.; Stick, S., The use of non-bronchoscopic
509 brushings to study the paediatric airway. *Respiratory Research* **2005**, *6*, (1), 53-53.
- 510 31. Martinovich, K. M.; Iosifidis, T.; Buckley, A. G.; Looi, K.; Ling, K.-M.; Sutanto, E. N.;
511 Kicic-Starcevic, E.; Garratt, L. W.; Shaw, N. C.; Montgomery, S.; Lannigan, F. J.; Knight, D.
512 A.; Kicic, A.; Stick, S. M., Conditionally reprogrammed primary airway epithelial cells maintain
513 morphology, lineage and disease specific functional characteristics. *Scientific Reports* **2017**, *7*,
514 (1), 17971.
- 515 32. Amanatidis, S.; Ntziachristos, L.; Giechaskiel, B.; Bergmann, A.; Samaras, Z., Impact of
516 Selective Catalytic Reduction on Exhaust Particle Formation over Excess Ammonia Events.
517 *Environmental Science & Technology* **2014**, *48*, (19), 11527-11534.
- 518 33. Prokopowicz, A.; Zaciera, M.; Sobczak, A.; Bielaczyc, P.; Woodburn, J., The Effects of
519 Neat Biodiesel and Biodiesel and HVO Blends in Diesel Fuel on Exhaust Emissions from a
520 Light Duty Vehicle with a Diesel Engine. *Environmental Science & Technology* **2015**, *49*, (12),
521 7473-7482.
- 522 34. Gioda, A.; Rodríguez-Cotto, R. I.; Amaral, B. S.; Encarnación-Medina, J.; Ortiz-
523 Martínez, M. G.; Jiménez-Vélez, B. D., Biodiesel from Soybean Promotes Cell Proliferation in
524 Vitro. *Toxicology In Vitro* **2016**, *34*, 283-288.
- 525 35. Chen, T.-M.; Kuschner, W. G.; Gokhale, J.; Shofer, S., Outdoor Air Pollution: Nitrogen
526 Dioxide, Sulfur Dioxide, and Carbon Monoxide Health Effects. *The American Journal of the*
527 *Medical Sciences* **2007**, *333*, (4), 249-256.
- 528 36. Chameides, W. L.; Fehsenfeld, F.; Rodgers, M. O.; Cardelino, C.; Martinez, J.; Parrish,
529 D.; Lonneman, W.; Lawson, D. R.; Rasmussen, R. A.; Zimmerman, P.; Greenberg, J.;
530 Middleton, P.; Wang, T., Ozone Precursor Relationships in the Ambient Atmosphere. *Journal of*
531 *Geophysical Research: Atmospheres* **1992**, *97*, (D5), 6037-6055.
- 532 37. Zhu, N.; Li, H.; Han, M.; Guo, L.; Chen, L.; Yun, Y.; Guo, Z.; Li, G.; Sang, N.,
533 Environmental nitrogen dioxide (NO₂) exposure influences development and progression of
534 ischemic stroke. *Toxicology Letters* **2012**, *214*, (2), 120-130.
- 535 38. Smith, B.; Nitschke, M.; Pilotto, L. S.; Ruffin, R.; Pisaniello, D.; Willson, K. J., *Health*
536 *effects of daily indoor nitrogen dioxide exposure in people with asthma*. 2000; Vol. 16, p 879-85.

- 537 39. Jaffe, L. S., Ambient Carbon Monoxide And Its Fate in the Atmosphere *Journal of the*
538 *Air Pollution Control Association* **1968**, *18*, (8), 534-540.
- 539 40. Weaver, L. K., Carbon Monoxide Poisoning. *New England Journal of Medicine* **2009**,
540 *360*, (12), 1217-1225.
- 541 41. Allen, J. G.; MacNaughton, P.; Cedeno-Laurent, J. G.; Cao, X.; Flanigan, S.; Vallarino,
542 J.; Rueda, F.; Donnelly-McLay, D.; Spengler, J. D., Airplane pilot flight performance on 21
543 maneuvers in a flight simulator under varying carbon dioxide concentrations. *Journal of*
544 *Exposure Science & Environmental Epidemiology* **2018**.
- 545 42. Elliott, A. R.; Prisk, G. K.; Schollmann, C.; Hoffmann, U., Hypercapnic ventilatory
546 response in humans before, during, and after 23 days of low level CO₂ exposure. *Aviation,*
547 *space, and environmental medicine* **1998**, *69*, (4), 391-6.
- 548 43. Drummer, C.; Friedel, V.; Borger, A.; Stormer, I.; Wolter, S.; Zittermann, A.; Wolfram,
549 G.; Heer, M., Effects of elevated carbon dioxide environment on calcium metabolism in humans.
550 *Aviation, space, and environmental medicine* **1998**, *69*, (3), 291-8.
- 551 44. Kittelson, D. B., ENGINES AND NANOPARTICLES: A REVIEW. *Journal of Aerosol*
552 *Science* **1998**, *29*, (5), 575-588.
- 553 45. Oberdörster, G.; Celein, R. M.; Ferin, J.; Weiss, B., Association of Particulate Air
554 Pollution and Acute Mortality: Involvement of Ultrafine Particles? *Inhalation Toxicology* **1995**,
555 *7*, (1), 111-124.
- 556 46. Seaton, A.; Godden, D.; MacNee, W.; Donaldson, K., Particulate air pollution and acute
557 health effects. *The Lancet* **1995**, *345*, (8943), 176-178.
- 558 47. Brook, R. D.; Rajagopalan, S.; Pope, C. A.; Brook, J. R.; Bhatnagar, A.; Diez-Roux, A.
559 V.; Holguin, F.; Hong, Y. L.; Luepker, R. V.; Mittleman, M. A.; Peters, A.; Siscovick, D.; Smith,
560 S. C.; Whitsel, L.; Kaufman, J. D., Particulate Matter Air Pollution and Cardiovascular Disease:
561 An Update to the Scientific Statement From the American Heart Association. *American Heart*
562 *Association* **2010**, *121*, (21), 2331-2378.
- 563 48. Goodson, J. M.; Weldy, C. S.; MacDonald, J. W.; Liu, Y.; Bammler, T. K.; Chien, W.-
564 M.; Chin, M. T., *In utero* exposure to diesel exhaust particulates is associated with an altered
565 cardiac transcriptional response to transverse aortic constriction and altered DNA methylation.
566 *The FASEB Journal* **2017**.
- 567 49. Yoza, B.; Matsumoto, M.; Matsunaga, T., DNA extraction using modified bacterial
568 magnetic particles in the presence of amino silane compound. *Journal of Biotechnology* **2002**,
569 *94*, (3), 217-224.
- 570 50. Øvrevik, J.; Refsnes, M.; Låg, M.; Holme, J. A.; Schwarze, P. E., Activation of
571 Proinflammatory Responses in Cells of the Airway Mucosa by Particulate Matter: Oxidant- and
572 Non-Oxidant-Mediated Triggering Mechanisms. *Biomolecules* **2015**, *5*, (3), 1399-1440.
- 573 51. Totlandsdal, A. I.; Cassee, F. R.; Schwarze, P.; Refsnes, M.; Låg, M., Diesel exhaust
574 particles induce CYP1A1 and pro-inflammatory responses via differential pathways in human
575 bronchial epithelial cells. *Particle and fibre toxicology* **2010**, *7*, 41-41.
- 576 52. Swanson, K. J.; Kado, N. Y.; Funk, W. E.; Pleil, J. D.; Madden, M. C.; Ghio, A. J.,
577 Release of the pro-inflammatory markers by BEAS-2B cells following in vitro exposure to
578 biodiesel extracts. *Open Toxicology Journal* **2009**, *3*, 8-15.
- 579 53. Skuland, T. S.; Refsnes, M.; Magnusson, P.; Oczkowski, M.; Gromadzka-Ostrowska, J.;
580 Kruszewski, M.; Mruk, R.; Myhre, O.; Lankoff, A.; Øvrevik, J., Proinflammatory effects of
581 diesel exhaust particles from moderate blend concentrations of 1st and 2nd generation biodiesel

- 582 in BEAS-2B bronchial epithelial cells—The FuelHealth project. *Environmental Toxicology and*
583 *Pharmacology* **2017**, *52*, (Supplement C), 138-142.
- 584 54. Hiraiwa, K.; van Eeden, S. F., Contribution of lung macrophages to the inflammatory
585 responses induced by exposure to air pollutants. *Mediators of inflammation* **2013**, *2013*, 619523-
586 619523.
- 587 55. Barnes, P. J., The cytokine network in asthma and chronic obstructive pulmonary disease.
588 *The Journal of clinical investigation* **2008**, *118*, (11), 3546-3556.
- 589 56. Xu, S.; Höglund, M.; Håkansson, L.; Venge, P., Granulocyte colony-stimulating factor
590 (G-CSF) induces the production of cytokines in vivo. *British Journal of Haematology* **2000**, *108*,
591 (4), 848-853.
- 592 57. Steiner, S.; Bisig, C.; Petri-Fink, A.; Rothen-Rutishauser, B., Diesel exhaust: current
593 knowledge of adverse effects and underlying cellular mechanisms. *Archives of Toxicology* **2016**,
594 *90*, 1541-1553.
- 595 58. Menten, P.; Wuyts, A.; Van Damme, J., Macrophage inflammatory protein-1. *Cytokine &*
596 *Growth Factor Reviews* **2002**, *13*, (6), 455-481.
- 597 59. Ghio, A. J.; Stonehuerner, J. G.; Dailey, L. A.; Richards, J. H.; Madden, M. D.; Deng, Z.;
598 Nguyen, N.-B.; Callaghan, K. D.; Yang, F.; Piantadosi, C. A., Carbon monoxide reversibly alters
599 iron homeostasis and respiratory epithelial cell function. *American journal of respiratory cell*
600 *and molecular biology* **2008**, *38*, (6), 715-723.
- 601 60. Fisher, A. B.; Hyde, R. W.; Baue, A. E.; Reif, J. S.; Kelly, D. F., Effect of carbon
602 monoxide on function and structure of the lung. *Journal of applied physiology* **1969**, *26*, (1), 4-
603 12.
- 604 61. de Brito, J. M.; Mauad, T.; Cavalheiro, G. F.; Yoshizaki, K.; de André, P. A.; Lichtenfels,
605 A. J. F. C.; Guimarães, E. T.; Rivero, D. H. R. F.; Antonangelo, L.; Oliveira, L. B.; Pedrosa, L.
606 R. M.; Macchione, M.; Saldiva, P. H. N., Acute exposure to diesel and sewage biodiesel exhaust
607 causes pulmonary and systemic inflammation in mice. *Science of The Total Environment* **2018**,
608 *628-629*, 1223-1233.
- 609 62. Seriani, R.; Junqueira, M. d. S.; de Toledo, A. C.; Martins, M. A.; Seckler, M.; Alencar,
610 A. M.; Negri, E. M.; Silva, L. F. F.; Mauad, T.; Saldiva, P. H. N.; Macchione, M., Diesel exhaust
611 particulates affect cell signaling, mucin profiles, and apoptosis in trachea explants of Balb/C
612 mice. *Environmental Toxicology* **2015**, *30*, (11), 1297-1308.

613